

Impending Regeneration Failure of the IUCN Vulnerable Borneo Ironwood (*Eusideroxylon zwageri*)

Authors: Qie, Lan, Elsy, Alexander D. , Stumvoll, Ashley, Kwasnicka, Magdalena, Peel, Anna L. , et al.

Source: Tropical Conservation Science, 12(1)

Published By: SAGE Publishing

URL: <https://doi.org/10.1177/1940082918823353>

The BioOne Digital Library (<https://bioone.org/>) provides worldwide distribution for more than 580 journals and eBooks from BioOne's community of over 150 nonprofit societies, research institutions, and university presses in the biological, ecological, and environmental sciences. The BioOne Digital Library encompasses the flagship aggregation BioOne Complete (<https://bioone.org/subscribe>), the BioOne Complete Archive (<https://bioone.org/archive>), and the BioOne eBooks program offerings ESA eBook Collection (<https://bioone.org/esa-ebooks>) and CSIRO Publishing BioSelect Collection (<https://bioone.org/csiro-ebooks>).

Your use of this PDF, the BioOne Digital Library, and all posted and associated content indicates your acceptance of BioOne's Terms of Use, available at www.bioone.org/terms-of-use.


Usage of BioOne Digital Library content is strictly limited to personal, educational, and non-commercial use. Commercial inquiries or rights and permissions requests should be directed to the individual publisher as copyright holder.

BioOne is an innovative nonprofit that sees sustainable scholarly publishing as an inherently collaborative enterprise connecting authors, nonprofit publishers, academic institutions, research libraries, and research funders in the common goal of maximizing access to critical research.

Impending Regeneration Failure of the IUCN Vulnerable Borneo Ironwood (*Eusideroxylon zwageri*)

Tropical Conservation Science
Volume 12: 1–6
© The Author(s) 2019
DOI: 10.1177/1940082918823353
journals.sagepub.com/home/trc



Lan Qie¹ , Alexander D. Elsy¹, Ashley Stumvoll¹, Magdalena Kwasnicka¹, Anna L. Peel¹, Joseph A. Sullivan¹, Maisie S. Ettinger¹, Alasdair J. Robertson¹, Jeanelle K. Brisbane¹, Amber L. Sawyer¹, Yan N. Lui¹, Siew Ngim Ow¹, Matteo Sebastianelli¹, Bartosz Majcher¹, Muying Duan¹, Hannah Vigus¹, Grace Pounsin², Reuben Nilus³, and Robert Ewers¹

Abstract

The regeneration of many climax species in tropical forest critically depends on adequate seed dispersal and seedling establishment. Here, we report the decreased abundance and increased spatial aggregation of younger trees of the Borneo ironwood (*Eusideroxylon zwageri*) in a protected forest in Sabah Malaysia. We observed a high level of seedling herbivory with strong density dependence, likely exacerbated by local aggregation and contributing to the progressively shrinking size distribution. We also note the largely undocumented selective herbivory by sambar deer on *E. zwageri* seedlings. This study highlights the combined impact of altered megafauna community on a tree population through interlinked ecological processes and the need for targeted conservation intervention for this iconic tropical tree species.

Keywords

seedling survival, neighborhood density, dispersal limitation, spatial clustering, threatened species, wildings

Introduction

The loss of megafauna due to climate change, habitat alteration, and hunting has a strong impact on plants that are dependent on large-bodied animals for seed dispersal, and this effect is particularly strong in tropical forests (Corlett, 2010; Galetti et al., 2017). Many tree species with large fruits have lost the animal dispersers they coevolved with, and consequently suffer reduced seed dispersal and increased spatial aggregation of seedlings, leading to lower survival and reduced gene flow (Galetti et al., 2017; Harrison et al., 2013). Such trees are typically late-succession to canopy species (Harrison et al., 2016), and their regeneration is crucial to the future of intact tropical forests and restoration efforts to accelerate succession in disturbed forests.

Borneo ironwood (*Eusideroxylon zwageri*) is a classic example of rainforest climax tree species facing regeneration challenges. *E. zwageri* is distributed in eastern and

southern Sumatra, Bangka, Belitung, Borneo, the Sulu archipelago, and Palawan. In Borneo, *E. zwageri* was formerly a common component of the main middle story in the mixed dipterocarp forest. It is a prized timber species, renowned for its extraordinary strength, durability, and rot resistance, and it is a cultural keystone species to the indigenous people of the region

¹Department of Life Sciences, Silwood Park Campus, Imperial College London, UK

²Maliau Basin Conservation Area, Yayasan Sabah, Malaysia

³Sabah Forestry Department Forest Research Centre, Sabah, Malaysia

Received 5 November 2018; Revised 10 December 2018; Accepted 12 December 2019

Corresponding Author:

Lan Qie, School of Life Sciences, University of Lincoln, Brayford Pool, Lincoln, UK.

Email: qie.lan@gmail.com



Creative Commons CC BY: This article is distributed under the terms of the Creative Commons Attribution 4.0 License (<http://www.creativecommons.org/licenses/by/4.0/>) which permits any use, reproduction and distribution of the work without further permission provided the original work is attributed as specified on the SAGE and Open Access pages (<https://us.sagepub.com/en-us/nam/open-access-at-sage>).

(Franco, Ghani, & Hidayati, 2014). Despite being long-living (>1,000 years), due to overexploitation (Peluso, 1992) and slow natural regeneration (typically requires >100 years to reach 30 cm diameter; Irawan, 2005), *E. zwageri* has become scarce across its distributional range and is classified as Vulnerable in the IUCN Red List (1998). The species produces large drupaceous fruits that measure 10 – 18 cm × 5 – 10 cm and contain a single large seed, measuring 7 – 15 cm × 4 – 7 cm, with a very hard seed coat (Irawan, 2005). Water dispersal for the heavy seeds of *E. zwageri* is possible but limited as the species often occur away from rivers (Irawan, 2005). Fruits of *E. zwageri* often have vertebrate feeding damage to the fleshy mesocarp and nut, suggesting this species may be dispersed by animals, although evidence of this is anecdotal and mainly refers to porcupines (Kostermans, Sunamo, Martawijaya, & Sudo, 1994; Phillipps & Phillipps, 2016). Only megafauna (>44 kg in body mass) would be capable of long-distance dispersal of these large seeds (Galetti et al., 2017). It has been speculated that *E. zwageri* seeds were dispersed by the Sumatran rhinoceros (*Dicerorhinus sumatrensis*; Phillipps & Phillipps, 2016), formerly occurring throughout Borneo but now virtually extinct in the wild in Sabah, and down to the last few individuals, if any, still surviving in Borneo (Hance, 2015). The other possible long-distance dispersal agent for seeds of this size is the Borneo pygmy elephant (*Elephas maximus borneensis*), the origin of which remains a subject of debate. The diet of these elephants has been extensively studied, but they have never been reported to interact with *E. zwageri* fruits. Despite the value and threatened status of *E. zwageri*, we still have very limited understanding of its ecology and, in particular, its natural regeneration from seeds.

Post germination, the seedling stage represents another bottleneck for the regeneration of a tree species under survival pressures such as pathogens and seedling herbivory, which are typically density dependent processes (Wang & Smith, 2002). There have been anecdotal reports that one of the common herbivores in Bornean forests, the sambar deer (*Rusa unicolor*), preferentially browsed on *E. zwageri* seedlings. This would potentially exacerbate the regeneration challenge already faced by *E. zwageri*. Once stems survive the dynamic stage and reach 10 cm diameter, the degree of spatial aggregation is expected to be stable across size classes (Condit et al., 2000). Therefore, in a long-living *E. zwageri* population (oldest trees may be > 1,000 years), spatial aggregation in younger versus older trees should reflect long-term change in the strength of seed dispersal. We conducted a survey of *E. zwageri* across all life stages from seedlings to adult trees in an old growth Bornean lowland forest and tested three hypotheses: (a) reduced regeneration in this local population will be reflected in its tree size

distribution curve with a flat gradient at small size classes or a unimodal shape (Halpin & Lorimer, 2017), (b) there is increased spatial aggregation in the younger subpopulation compared to the older subpopulation of adult trees (≥ 10 cm diameter), and (c) seedlings herbivory damage has density dependence.

Methods

We conducted this study in January 2018 at Maliau Basin Conservation Area (4.74° N, 116.97° E, 256 m a.s.l.), Sabah, Malaysia. The survey area was in old growth forest on the south bank of the Maliau River near the edge of the conservation area, in the vicinity of a large buffer zone consisting of selectively logged and secondary forests. A 4-ha square plot was set up, within which a grid of 10 by 10 m subplots was established for tree surveys. All *E. zwageri* trees with diameter at breast height (DBH) ≥ 1 cm were mapped and recorded to diameter classes in 10 cm intervals.

There were abundant *E. zwageri* seedlings in the area from the January 2017 fruiting event. Many of these had noticeable herbivory damage with young leaves missing from the top, characteristic of that by mammalian herbivores, and often resulting in complete defoliation. We surveyed seedlings (all stages with DBH < 1 cm) in a 1.35 ha area in the centre of the 4-ha plot, using 5 by 5 m subplots within the 10 by 10 m grid. In each seedling subplot, we counted *E. zwageri* seedlings in two categories: *surviving* if the apical meristem was intact, and *fatally damaged* if the apical meristem was eaten, including dead *E. zwageri* seedlings with evidence of herbivory damage. Leafless and dead *E. zwageri* seedlings were identified by their characteristic straight and reddish bare stem, often with the stony seed coat still attached at root. For each subplot, we also recorded ground vegetation cover as a proxy for interspecific competition, estimated into four classes: 0% to 25%, 25% to 50%, 50% to 75%, and 75% to 100%.

We quantified spatial aggregation of *E. zwageri* adult trees (DBH ≥ 10 cm) using the relative neighborhood density Ω_x , defined as the average density of neighbors in the annulus at distance x for each focal tree, standardized by mean population density (Condit et al., 2000). We calculated Ω_x for the 0 to 100 m distance range in 10 m steps and compared this metric for the older and younger subpopulations, categorized as those above or below the median tree size respectively.

For *E. zwageri* seedlings, we used a generalized linear model with a binomial error distribution and a logit link function to test the effects of conspecific seedling density and vegetation cover on the proportion of fatal herbivory damage. All analyses were conducted in the R statistical computing environment (R Core Team, 2016).

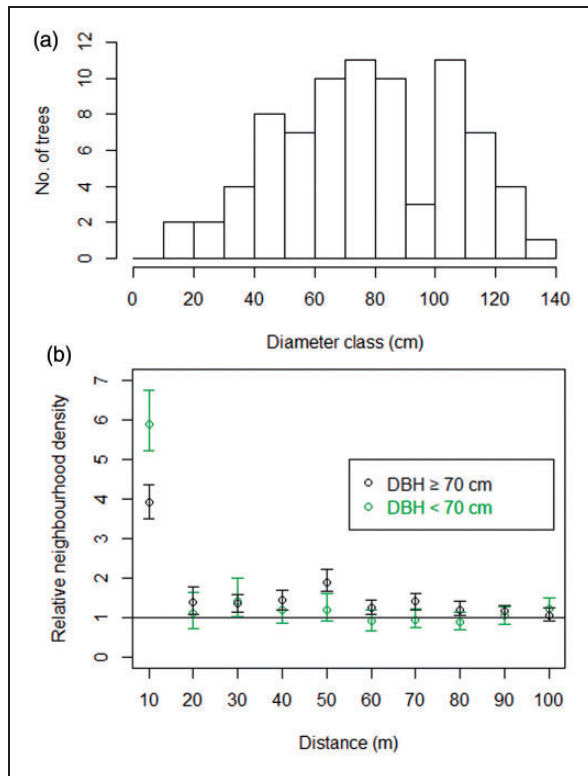


Figure 1. Demographic and spatial distribution of *E. zwageri* adult trees (DBH ≥ 10 cm) in the 4-ha plot. (a) Population size distribution with a unimodal shape and a mode at the 70 cm class. (b) Spatial aggregation measured as the mean relative neighborhood density in 10-m annuli around each focal tree, standardized by mean population density. The population was divided into the older (DBH ≥ 70 cm) and younger (DBH < 70 cm) subpopulations. Bars represent 95% bootstrapped confidence intervals. DBH = diameter at breast height.

Results

We recorded 90 *E. zwageri* trees (DBH ≥ 1 cm) in the 4-ha plot, among which 80 were adult trees (DBH ≥ 10 cm). As predicted from the reduced regeneration hypothesis, the adult tree size distribution had a unimodal shape peaking at the 70 cm class (median) and decreasing toward smaller classes (Figure 1(a)). All adult trees were aggregated at the 0 to 10 m scale ($\Omega_x > 1$), and the degree of aggregation was significantly higher in the younger subpopulation, consistent with the expectation of increased dispersal limitation over time (Figure 1(b)).

A total of 495 *E. zwageri* seedlings were recorded in 197 of the 540 5 by 5 m subplots surveyed, with an overall mean density of 367 seedlings ha⁻¹. Nearly all appeared to be less than 1 year old and resulting from the fruiting event in January 2017. Among these, 53.3% had fatal herbivory damage, already dead or highly unlikely to survive. The proportion of fatal herbivory damage increased rapidly with conspecific seedling density at subplot level ($p < .001$), predicted to reach 0.97 at 1 seedling m⁻²

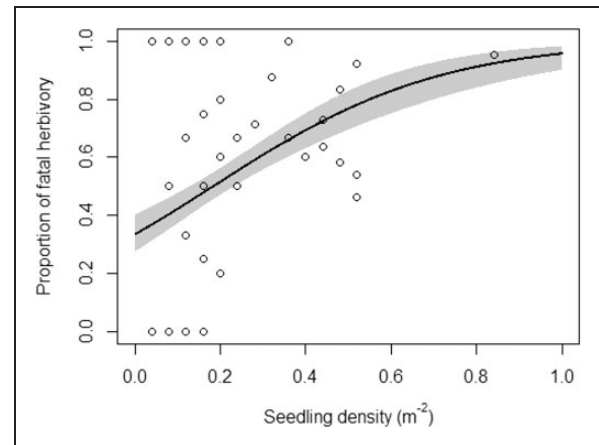


Figure 2. Relationship between the rate of fatal herbivory of *E. zwageri* seedlings and conspecific seedling density at 5 \times 5 m subplot level. Line was fitted using a generalized linear model. Shaded areas represent confidence intervals.

(Figure 2). Fatal herbivory was not significantly associated with vegetation cover ($p = .09$).

Discussion

The remarkably low abundance of young *E. zwageri* trees indicates that the population of this species is not sustainable in this forest stand (Halpin & Lorimer, 2017). *E. zwageri* adult trees were aggregated at the 0 to 10 m spatial scale, characteristic of tropical trees (Condit et al., 2000); a similar aggregation pattern at this scale was observed at genotypical level in a recent study (Shao et al., 2017). However, the increased spatial aggregation of *E. zwageri* in younger trees supports the hypothesis of increasing seed dispersal limitation over time due to the lack of seed dispersers, as shown in another Bornean forest (Harrison et al., 2013). We highlight the effect of dispersal limitation and negative density-dependent seedling survival, operating together to influence species spatial distribution (Shao et al., 2017).

A survey of 13 other Bornean canopy species found a seedling survival rate of approximately 50% over 10 years (Delissio, Primack, Hall, & Lee, 2002). In contrast, we observed an overall survival rate of 46.7% in *E. zwageri* seedlings in less than 1 year since germination. The actual survival rate may be even lower as we only accounted for dead stems with detectable remains. High seedling mortality in the first year may not be extreme, but strikingly, *E. zwageri* seedlings more than 1 year old (height > 1 m) were almost absent, suggesting complete failure of seedling establishment in recent years. Using our model as a conservative approximation of the density dependent annual mortality rate and applying it to the surviving *E. zwageri* seedlings at subplot level, we predict that after 6 years seedling density

would drop to merely 3 individuals ha^{-1} , leading to a progressively shrinking demography (Figure 1(a)). Seedling height may not reach 2m in this time (Irawan, 2005), and thus, they remain susceptible to ground level herbivory.

The heavy herbivory observed here could be associated with a single mammalian herbivore, the sambar deer, based on characteristics of the stem damage. Local rangers reported that sambar deer repeatedly broke into the nursery at the Maliau Basin Study Centre and selectively browsed on the *E. zwageri* seedlings cultivated there, but not other species. Adapted to surviving shade conditions for relatively long period, *E. zwageri* seedlings allocate a high proportion of resources towards defense in the form of concentrated tannin and lignin (Kurokawa, Kitahashi, Koike, Lai, & Nakashizuka, 2004). This is an evolved mechanism against invertebrate herbivores and pathogens, but mammalian herbivores such as sambar deer have coevolved to produce a tannin-binding protein in their saliva and in fact show a preference toward plant species high in concentrated tannin and lignin (Semiadil, Barry, Muir, & Hodgson, 1995). Evidence of preferred browsing of *E. zwageri* seedlings by sambar deer remains, however, anecdotal, and further verification is needed.

It is possible that at local scale, browsing pressure can influence *E. zwageri* regeneration. A generalist grazer and browser, the sambar deer appears to benefit from intermediate forest disturbance, preferring forest edges and selectively logged forest due to the increased availability of phytomass on the ground (Corlett, 2007; Meijaard & Sheil, 2007). A recent study in our research area found a 19% increase in the abundance of mammalian herbivores in logged forest compared to unlogged forest and, in particular, the more than doubling of sambar deer abundance (Wearn et al., 2017). In the absence of tigers and leopards in Borneo, the Sunda clouded leopard (*Neofelis diardi*) may be the most important natural predator of the sambar deer; however, with relatively small body size, the clouded leopard can probably only take the fawns of sambar. Predation pressure on the sambar deer largely comes from hunting (Bennett, Nyaoi, & Sompud, 2000). It is likely that in protected areas where hunting is prohibited, the sambar deer has enjoyed predator release. Our study site is in a protected area adjacent to a large buffer zone consisting of selectively logged forests. The sambar deer population here therefore have probably benefited from both favorable habitat and the absence of hunting. As an interesting contrast, sambar deer population was low in another Sabah forest (Kabili-Sepilok; Ross, Hearn, Johnson, & Macdonald, 2013), and greater densities of young *E. zwageri* trees were recorded, on average 49 stems with DBH 5 to 10 cm per 4-ha (Qie and Nilus, unpublished data).

Our results shine a spotlight on an iconic tree species *E. zwageri*. We show the intensified interaction between this native tree and a native mammalian herbivore resulting from changes in the megafauna community and forest landscape in Borneo. Anthropogenic forces on the forest ecosystem in recent centuries seem to have eradicated its long-distance seed dispersers and in some places favored its seedling herbivores. Where these impacts co-occur, they may lead to an impending regeneration failure of *E. zwageri* at local scale, as we are starting to see in this population.

Implications for Conservation

As a threaten species, an updated IUCN Red List assessment is due for *E. zwageri* population status across its distribution range. Our study highlights the impending challenges faced by its natural regeneration and call for more survey and monitoring work to be carried out at other localities where *E. zwageri* populations persist. *E. zwageri* requires active conservation intervention and efforts should focus on its most vulnerable seedling stage. Once survived through this recruitment bottleneck, the tall saplings and trees of *E. zwageri* are known to be exceptionally resilient, including being drought- and fire-tolerant (Delmy, 2001; Van Nieuwstadt & Sheil, 2005). Ex situ propagation of this species has only been carried out on small scales partly due to inadequate supply of seeds and seedlings (Azani, Majid, & Meguro, 2001; Irawan, 2012), yet we show that a great number of *E. zwageri* seedlings in natural forests are being lost to high mortality. Our results suggest that at local density of $<0.2 \text{ m}^{-2}$ *E. zwageri* seedlings may have optimal survival. Therefore, moderate harvest of *E. zwageri* seeds and seedlings in natural forests (i.e., wildlings) for ex situ cultivation and reintroduction once they reach sapling stage should be feasible. This approach may simultaneously increase *E. zwageri* regeneration and population gene flow, providing a cost-effective way of assisted forest regeneration. As an indigenous cultural icon, its conservation strategy should not be limited to protected forests but also include community managed forests where local people may be incentivized and empowered to take positive actions (Santika et al., 2017). Across its distribution range, *E. zwageri* cultivation can be readily integrated into community based agroforestry initiatives to provide long-term benefits for local people.

Acknowledgments

The authors thank the Sabah Biodiversity Centre for research permit and the excellent support of staff members at Maliau Basin Studies Centre.

Declaration of Conflicting Interests

The author(s) declared no potential conflicts of interest with respect to the research, authorship, and/or publication of this article.

Funding

The author(s) disclosed receipt of the following financial support for the research, authorship, and/or publication of this article: This study was supported by the Imperial College London MRes Tropical Forest Ecology field course program. L.Q. was supported by a Natural Environment Research Council grant (NE/P00363X/1).

ORCID iD

Lan Qie  <http://orcid.org/0000-0002-7309-7523>

References

- The IUCN Red List of Threatened Species (1998). e. T31316A9624725. Asian Regional Workshop (Conservation & Sustainable Management of Trees, Viet Nam, August 1996). *Eusideroxylon zwageri*. Available at: <https://www.iucnredlist.org/species/31316/9624725>
- Azani, A. M., Majid, N. M., & Meguro, S. (2001). Rehabilitation of tropical rainforests based on indigenous species for degraded areas in Sarawak, Malaysia. In S. Kobayashi, J. W. T. Tunbull, T. Toma, Mori, & N. M. N. A. Majid (Eds), *Rehabilitation of tropical degraded forest ecosystems* (pp. 141–148). Bogor, Indonesia: Center for International Forestry Research.
- Bennett, E., Nyaoi, A., & Sompud, J. (2000). Saving Borneo's bacon: The sustainability of hunting in Sarawak and Sabah. In J. Robinson & E. Bennett (Eds), *Hunting for sustainability in tropical forests* (pp. 305–324). New York, NY: Columbia University Press.
- Condit, R., Ashton, P. S., Baker, P., Bunyavejchewin, S., Gunatilleke, S., Gunatilleke, N., . . . Yamakura, T. (2000). Spatial patterns in the distribution of tropical tree species. *Science*, 288(5470), 1414–1418. doi:10.1126/science.288.5470.1414
- Corlett, R. T. (2007). The impact of hunting on the mammalian fauna of tropical Asian forests. *Biotropica*, 39(3), 292–303. doi:10.1111/j.1744-7429.2007.00271.x
- Corlett, R. T. (2010). Megafaunal extinctions and their consequences in the tropical Indo-Pacific. *Terra Australis*, 32, 117–131.
- Delissio, L. J., Primack, R. B., Hall, P., & Lee, H. S. (2002). A decade of canopy-tree seedling survival and growth in two Bornean rain forests: Persistence and recovery from suppression. *Journal of Tropical Ecology*, 18(5), 645–658. doi:10.1017/S0266467402002420
- Delmy, A. (2001). Fire resistance of tree species in Bukit Soeharto Education Forest, East Kalimantan, Indonesia. In S. Kobayashi, J. W. T. Tunbull, T. Toma, Mori, & N. M. N. A. Majid (Eds), *Rehabilitation of tropical degraded forest ecosystems* (pp. 27–34). Bogor, Indonesia: Center for International Forestry Research.
- Franco, F., Ghani, B., & Hidayati, S. (2014). *Terras* (*Eusideroxylon zwageri* Teijsm. & Binn.), a cultural keystone species of the Berawan people of Sarawak, Malaysia. *Pertanika Journal of Social Sciences and Humanities*, 22(3), 891–902.
- Galetti, M., Moleón, M., Jordano, P., Pires, M. M., Guimarães, P. R., Pape, T., . . . Svenning, J. C. (2017). Ecological and evolutionary legacy of megafauna extinctions. *Biological Reviews*. doi:10.1111/brv.12374
- Halpin, C., & Lorimer, C. (2017). A demographic approach to evaluating tree population sustainability. *Forests*, 8(2), 46.
- Hance, J. (2015). *Officials: Sumatran rhino is extinct in the wild in Sabah*. Retrieved from <https://news.mongabay.com/2015/04/officials-sumatran-rhino-is-extinct-in-the-wild-in-sabah/>
- Harrison, R. D., Sreekar, R., Brodie, J. F., Brook, S., Luskin, M., O'Kelly, H., . . . Velho, N. (2016). Impacts of hunting on tropical forests in Southeast Asia. *Conservation Biology*, 30, 972–981. doi:10.1111/cobi.12785
- Harrison, R. D., Tan, S., Plotkin, J. B., Slik, F., Detto, M., Brenes, T., . . . Davies, S. J. (2013). Consequences of defaunation for a tropical tree community. *Ecology Letters*, 16, 687–694. doi:10.1111/ele.12102
- Irawan, B. (2005). *Ironwood (Eusideroxylon zwageri Teijsm. & Binn.) and its varieties in Jambi, Indonesia*. (PhD), University of Göttingen, Göttingen.
- Irawan, B. (2012). Growth performance of one year old seedlings of ironwood (*Eusideroxylon zwageri* Teijsm. & Binn.) varieties. *Jurnal Manajemen Hutan Tropika*, 18(3). doi:10.7226/jmht.18.3.184
- Kostermans, A. J. G. H., Sunamo, B., Martawijaya, A., & Sudo, S. (1994). *Eusideroxylon Teijsm. and Binnend*. In L. Soerianegara & R. H. M.J. Lemmens (Eds), *Plant resources of South-East Asia no. 5(1) Timber trees: Major commercial timbers*. Bogor, Indonesia: PROSEA.
- Kurokawa, H., Kitahashi, Y., Koike, T., Lai, J., & Nakashizuka, T. (2004). Allocation to defense or growth in dipterocarp forest seedlings in Borneo. *Oecologia*, 140(2), 261–270. doi:10.1007/s00442-004-1566-7
- Meijaard, E., & Sheil, D. (2007). The persistence and conservation of Borneo's mammals in lowland rain forests managed for timber: Observations, overviews and opportunities. *Ecological Research*, 23(1), 21. doi:10.1007/s11284-007-0342-7
- Peluso, N. L. (1992). The ironwood problem: (Mis) Management and development of an extractive rainforest product. *Conservation Biology*, 6(2), 210–219. doi:10.1046/j.1523-1739.1992.620210.x
- Phillipps, Q., & Phillipps, K. (2016). *Phillipps' field guide to the mammals of Borneo and their ecology: Sabah, Sarawak, Brunei, and Kalimantan*. Princeton, NJ: Princeton University Press.
- R Core Team. (2016). *R: A language and environment for statistical computing*. Vienna, Austria: R Foundation for Statistical Computing. Retrieved from <https://www.R-project.org/>
- Ross, J., Hearn, A. J., Johnson, P. J., & Macdonald, D. W. (2013). Activity patterns and temporal avoidance by prey in response to Sunda clouded leopard predation risk. *Journal of Zoology*, 290(2), 96–106. doi:10.1111/jzo.12018

- Santika, T., Meijaard, E., Budiharta, S., Law, E. A., Kusworo, A., Hutabarat, J. A., . . . Wilson, K. A. (2017). Community forest management in Indonesia: Avoided deforestation in the context of anthropogenic and climate complexities. *Global Environmental Change*, *46*, 60–71. doi:10.1016/j.gloenvcha.2017.08.002
- Semiadil, G., Barry, T. N., Muir, P. D., & Hodgson, J. (1995). Dietary preferences of sambar (*Cervus unicolor*) and red deer (*Cervus elaphus*) offered browse, forage legume and grass species. *The Journal of Agricultural Science*, *125*(1), 99–107. doi:10.1017/S0021859600074554
- Shao, X., Brown, C., Worthy, S. J., Liu, L., Cao, M., Li, Q., . . . Swenson, N. G. (2017). Intra-specific relatedness, spatial clustering and reduced demographic performance in tropical rainforest trees. *Ecology Letters*. Advance online publication. doi:10.1111/ele.13086
- Van Nieuwstadt, M. G. L., & Sheil, D. (2005). Drought, fire and tree survival in a Borneo rain forest, East Kalimantan, Indonesia. *Journal of Ecology*, *93*(1), 191–201. doi:10.1111/j.1365-2745.2004.00954.x
- Wang, B. C., & Smith, T. B. (2002). Closing the seed dispersal loop. *Trends in Ecology & Evolution*, *17*(8), 379–386. doi:10.1016/S0169-5347(02)02541-7
- Wearn, O. R., Rowcliffe, J. M., Carbone, C., Pfeifer, M., Bernard, H., & Ewers, R. M. (2017). Mammalian species abundance across a gradient of tropical land-use intensity: A hierarchical multi-species modelling approach. *Biological Conservation*, *212*, 162–171. doi:10.1016/j.biocon.2017.05.007