

Population assessment and habitat use by marine otter *Lontra felina* in southern Chile

Authors: Medina-Vogel, Gonzalo, Bartheld, José Luis, Pacheco, Ricardo Alvarez, and Rodríguez, Claudio Delgado

Source: Wildlife Biology, 12(2) : 191-199

Published By: Nordic Board for Wildlife Research

URL: [https://doi.org/10.2981/0909-6396\(2006\)12\[191:PAAHUB\]2.0.CO;2](https://doi.org/10.2981/0909-6396(2006)12[191:PAAHUB]2.0.CO;2)

BioOne Complete (complete.BioOne.org) is a full-text database of 200 subscribed and open-access titles in the biological, ecological, and environmental sciences published by nonprofit societies, associations, museums, institutions, and presses.

Your use of this PDF, the BioOne Complete website, and all posted and associated content indicates your acceptance of BioOne's Terms of Use, available at www.bioone.org/terms-of-use.

Usage of BioOne Complete content is strictly limited to personal, educational, and non - commercial use. Commercial inquiries or rights and permissions requests should be directed to the individual publisher as copyright holder.

BioOne sees sustainable scholarly publishing as an inherently collaborative enterprise connecting authors, nonprofit publishers, academic institutions, research libraries, and research funders in the common goal of maximizing access to critical research.

Population assessment and habitat use by marine otter *Lontra felina* in southern Chile

Gonzalo Medina-Vogel, José Luis Bartheld, Ricardo Alvarez Pacheco & Claudio Delgado Rodríguez

Medina-Vogel, G., Bartheld, J.L., Pacheco, R.A. & Delgado, R. 2006: Population assessment and habitat use by marine otter *Lontra felina* in southern Chile. - Wildl. Biol. 12: 191-199.

The marine otter *Lontra felina* lives along the Pacific coast of South America from 6°S to 56°S. A method for estimating marine otter abundance is described and a comparison of the characteristics of the terrestrial habitat of this species in southern Chile is made. From June 1999 to June 2000, we conducted eight hours of observations on one day each month at four study sites. An average of 3.8 observable otters/km were recorded, with significant differences between sites but no systematic trends with regard to season. Observable pups were recorded year-round. Otters spent 80% of their time out of view, engaged in behaviours other than feeding; observable otters were mostly seen feeding. The methodology described is effective for indexing abundances at small geographic scales. The marine otter needs both an aquatic habitat for foraging and a terrestrial habitat with an abundance of safe shelters, which, in turn, could limit the distribution of the species when human settlers and domestic dogs utilise the terrestrial habitats.

Key words: abundance, observable otters, shelters, terrestrial and aquatic habitat

Gonzalo Medina-Vogel*, José Luis Bartheld**, Ricardo Alvarez Pacheco*** & Claudio Delgado Rodríguez***, Instituto de Ecología y Evolución, Universidad Austral de Chile, Casilla 567, Valdivia, Chile - e-mail addresses: gmedina@unab.cl (Gonzalo Medina-Vogel); jlbarthled@yahoo.com (José Luis Bartheld); ralvarez@conservacionmarina.cl (Ricardo Alvarez Pacheco); cdelgado@conservacionmarina.cl (Claudio Delgado Rodríguez)

Present addresses:

*Escuela de Medicina Veterinaria, Facultad de Ecología y Recursos Naturales, Universidad Nacional Andrés Bello, República 252, Santiago, Chile

**Comité Nacional pro Defensa de la Fauna y Flora, Carlos Anwandter 849, Valdivia, Chile

***Irlanda 1802, Valdivia, Chile

Corresponding author: Gonzalo Medina-Vogel

Received 17 August 2004, accepted 23 December 2004

Associate Editor: Paolo Cavallini

Data on distribution, abundance, habitat selection and population trends are an important tool for animal conservation and management of species of interest. These

data can aid species protection and habitat conservation programmes, which are dependent on information and evidence at regional, national and international levels.

Such data are also essential for research programmes, either to provide a safeguard against projects that threaten important species or to evaluate the impact of measures taken (Reuther et al. 2000, Gese 2001). In order to make the most efficient use of such information, data should be obtained in a standardised and optimal form following a scientific assessment. The benefits of standardised collection include increased efficiency, where information on species distribution, abundance and behaviour can be made available together with other relevant ecological and biological information, thereby providing a more holistic perspective for management. Furthermore, efficiency could be improved if data were prepared in a comparable and transformable form, since personnel and financial resources in nature conservation are limited (Reuther et al. 2000).

Among carnivores, Mustelidae (otters), Phocidae (true seals), Otariidae (fur seals and sea lions) and Odobenidae (walrus) are the only species which are exclusively associated with aquatic habitats. Two species of otters have an obligatory marine existence: the sea otter *Enhydra lutris* and the marine otter *Lontra felina* (Estes 1989). Of these, the marine otter has most recently adapted to marine living conditions; it is also the smallest otter in its genus (Estes 1989). However, it is poorly known along virtually all imaginable ecological dimensions (see Larivière 1998 for a review) and, therefore the development of appropriate management or conservation policies is difficult. Marine otters typically inhabit coasts with exposed rocky shores, covering areas approximately 30 m inland and 150 m offshore along the Pacific coast (Castilla & Bahamondes 1979, Ostfeld et al. 1989). Their distribution ranges from 6°S (near Chimbote, Peru) to 56°S (at Cape Horn and Isla de los Estados, Argentina; Larivière 1998, Parera 2002). In Chile, latitudinal differences in marine otter population abundance have been documented, ranging from 0.04 otter/km in the Beagle Channel (55°S) to 6.9 otter/km on Chiloe Island (42°S), and 1.0–4.4 otter/km north of 29°S (Castilla 1982, Rozzi & Torres-Mura 1990, Ebensperger & Castilla 1991, Sielfeld 1992, Medina 1995b). However, previous studies have encountered problems in assessing animal distribution, abundance, behaviour and the relative importance of aquatic and terrestrial habitats due to difficulties in recognising individual animals, the lack of sexual dimorphism, and the long time that otters are obscured by caves, rocks and other terrestrial structures (Ebensperger & Castilla 1991, Sielfeld 1992, Ostfeld et al. 1989, Medina 1995b, Larivière 1998). Our study therefore had two objectives: 1) to develop a method for estimating the abundance of the marine otter within a region, and 2) to assess its use of terrestrial habitat.

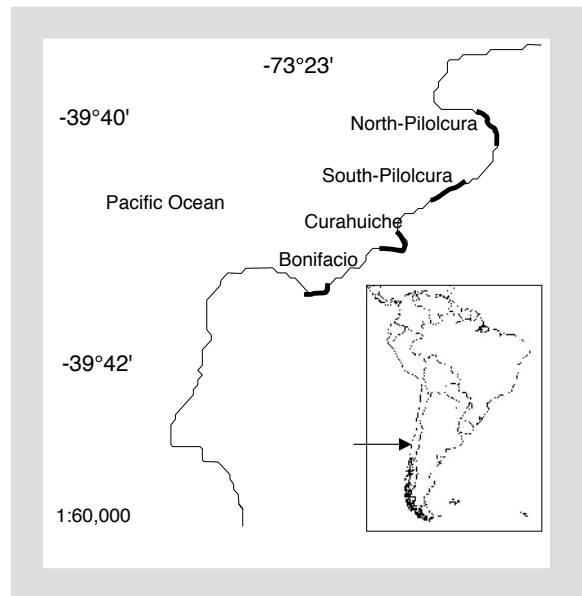


Figure 1. Location of the four coastal study sites in Chile. The bold lines indicate the 1-km lengths of coastline that were surveyed.

Material and methods

Our study was conducted at four sites (Bonifacio, Curahuiche, North-Pilolcura and South-Pilolcura) between Punta Bonifacio (39°41'30"S, 73°W) and Punta Chungungo (39°39'30"S, 73°W) in southern Chile (Fig. 1). The sites were selected based on relatively high otter numbers observed in preliminary surveys, and also for ease of access and observation. Each study site consisted of 1 km of coastline (with shoreline measured on a 1:50,000 scale map). Sites were separated by 1.4–1.6 km, and each site could be entirely viewed from any of three elevated observation points. The shoreline at the four study sites is rocky and exposed to strong winds and large waves, with tidal heights ranging within 0.6–3.0 m. The shoreline is dominated by brown algae (*Durvillaea antarctica*, *Macrocystis pyrifera*, *Mazzaella laminarioides* and *Lessonia nigrescens*), with a notable presence of other algae (*Corallina* sp. and *Lithothamnion* sp.; Moreno et al. 1984, Medina 1995a,b, Medina-Vogel et al. 2004).

The method used to estimate marine otter abundance was based on direct observation over the entire area of each study site. From June 1999 to June 2000 three observers worked at each site for one day per month. In order to separate the area scanned by each observer inside each study site, the shoreline of each study site was subdivided perpendicularly into three zones (x,y,z) with similar shoreline lengths. Each zone encompassed the area from 30 m above the high tide line to 150 m off-

shore and was delineated using natural markers (i.e. reef rocks, rocks or cracks), so that each observer counted and registered the otters and their activities only inside his observation zone. Observations using 10 × 50 binoculars were made simultaneously in each zone, during 2-minute surveys with an interval of eight minutes, over two continuous 4-hour periods (i.e. morning and afternoon) per day (method adapted from Estes et al. 1982, Ostfeld et al. 1989). Because of weather conditions, this procedure was conducted every 28–45 days on 10 occasions, by the same three observers, resulting in a total of four days of observation per field trip. At the study sites, observers hid in the vegetation above the rocky shoreline. The otters did not display any avoidance or fear behaviours related to the presence of observers. Any observation which involved walking along the coast (e.g. measurements of habitat characteristics and collection of spraints and prey remains) were made after the observation days. As marine otters have no discernible body markings or sexual dimorphism, and can not be counted at night or during the time they are inside caves or rock cracks (Castilla 1982, Ebensperger & Castilla 1991, Sielfeld 1992, Medina 1995b Larivière 1998), our method could not assess total otter numbers at any one time. Estimates and comparisons of otters were therefore determined by counting otters seen in the water or on rocks during daylight hours only. The otters that were counted and observed were termed 'observable otters'. Therefore, our study compared the maximum number of observable otters seen by the three observers during one simultaneous scanning (i.e. two minutes of simultaneous scanning during the 8-hour observation period). It did not correspond to the cumulative number of otters counted. As otters below the surface were not recorded, a small bias could be inherent in our scanning technique. To reduce this bias we scanned the ocean surface at a rate slow enough for every point within the observed zones to be observed, for a period longer than the duration of the longest described dive (about 65 seconds), and 2–4 times the duration of an average dive (25–44 seconds; Ostfeld et al. 1989, Medina 1995b). Otters observed during the scanning of the sea surface were classified as adults or pups, and their activity was described as grooming, feeding, travelling, resting or socialising (i.e. two or more otters interacting; Shimek & Monk 1977). As only one case was recorded, otters feeding together were classified as feeding, and an otter was considered to be resting when it was observed to be stationary on a rock.

The habitat characteristics of each study site were evaluated using 10-m wide transects perpendicular to the shore from 30 m above the high tide line to the sea surface, with a distance between transects of 100 m, mak-

ing a total of 10 transects per study site. In each transect, the number and frequency of three categories of rock sizes were recorded: small (i.e. < 40 cm in diameter, corresponding to rocks smaller than an adult otter), medium (i.e. rocks 40–65 cm in diameter, similar to the size of an adult otter), and large rocks (i.e. rocks > 65 cm in diameter, corresponding to rocks larger than an adult otter (Fig. 2). The otter habitat was also classified according to the total number of rock caves, cracks and reef rocks (i.e. large rocks permanently surrounded by water), per transect for each study site (see Fig. 2). These habitat characteristics were selected because they are thought to be important habitat characteristics for the marine otter (Castilla & Bahamondes 1979, Ebensperger & Castilla 1992, Sielfeld 1990). In order to assess terrestrial structures associated with highly used otter shelters, we tagged and characterised all locations inside each study site that were visited by ≥ 1 otter during more than three different observation periods (i.e. during three different months). Shelters were assessed using the same habitat characteristics that were described above (i.e. rock sizes and types). Finally, all study sites were patrolled on 10 occasions every 28–45 days from June 1999 to June 2000 looking for otter spraints and prey remains. Prey remains refer to prey brought ashore but not completely consumed. All spraints and prey remains were removed or collected and counted.

For the analysis of otter abundance, sites and seasons were used as independent variables. For the assessment of relationships between otter abundance and habitat characteristics, abundance was treated as the dependent variable and habitat characteristics as the independent variables. Since the data did not present a normal distribution (Shapiro-Wilks, symmetry and kurtosis test), a Kruskal-Wallis one-way analysis of variance by ranks was used to assess differences in otter density and habitat characteristics between study sites, seasons and daytime periods (Sokal & Rohlf 1981, Zar 1996). To assess differences in spraints and prey remains between sites, the Wilcoxon signed-rank test was used, with each study site treated as an independent sample (Zar 1996). The Spearman correlation coefficient was used to test the relationship between spraints, prey remains and otter densities (Zar 1996). The significance level was set at 5% for all tests.

Results

Otter abundance

We carried out 1,960 separate counts of observable otters, consisting of 2-minute scans (49 otter samples/day for



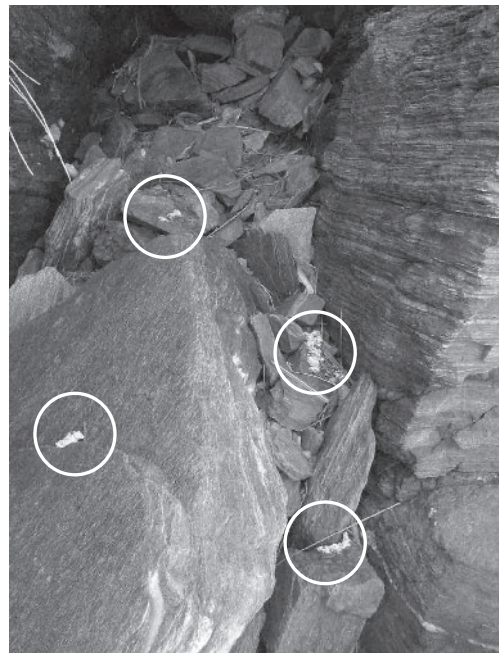
Large rocks with cracks



Reef rocks



Large, medium and small rocks without cracks



Cave with otter spraints

Figure 2. Rock sizes and types (i.e. with or without cracks), and a cave with otter spraints. (See 'Material and methods').

10 months) during eight hours of daytime with a total of 320 hours of observation. One or more otters were recorded 1,764 (90%) times. The otter's activity schedule was 54% feeding, 28% traveling, 8% grooming, 6% socialising and 4% resting with no differences between seasons ($H_3 = 2.6, P = 0.5$). In summer more otters were recorded during the first four hours of observation (morning) than during the ensuing four hours (afternoon; $H_3 = 5.3, P = 0.02$). The mean distribution was 3.8 otters/km, of which 18% were pups born during the year (Table 1).

Table 1. Estimated mean maximum abundance of the marine otter during 10 months of observation at the four study sites in Chile.

Study site	Range (otters/km)		Mean maximum (otters/km)		Total mean maximum
	Adults	Cubs	Adults (SD)	Pups (SD)	
Bonifacio	1-5	0-2	2.4 (0.5)	0.4 (0.7)	2.8
Curahuiche	2-5	0-4	3.4 (0.4)	1.5 (1.4)	4.8
South-Pilolcura	2-4	0-1	2.7 (0.2)	0.4 (0.5)	3.1
North-Pilolcura	2-5	0-2	3.7 (0.2)	0.5 (0.7)	4.2
Total	1-5	0-4	3.1 (0.2)	0.7 (1.0)	3.8

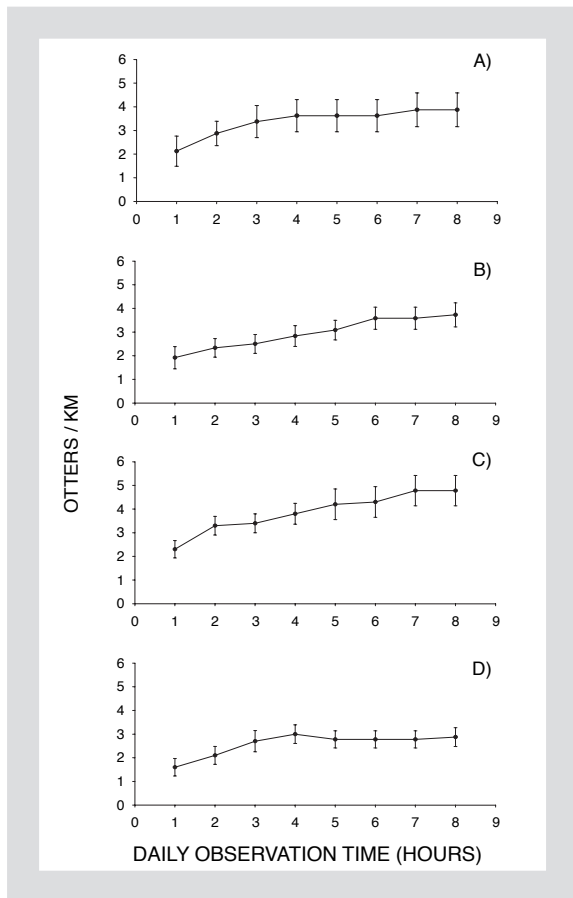


Figure 3. Variation in the maximum number of otters counted simultaneously by three observers. Data are shown as the accumulative hours of observation for the seasons winter (A), spring (B), autumn (C) and summer (D); observations from all study sites are pooled. The vertical bars indicate standard error.

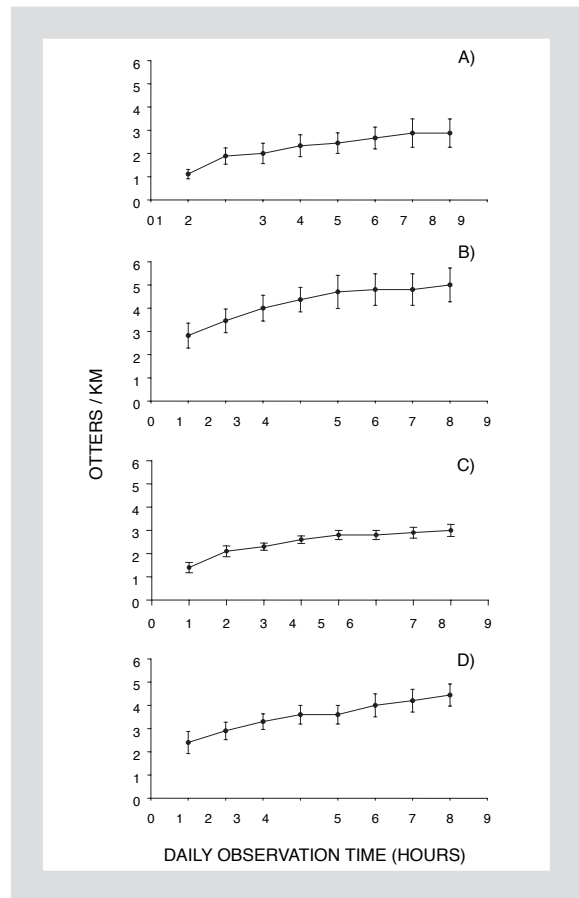


Figure 4. Variation in the maximum number of otters counted simultaneously by three observers. Data are shown as the accumulative hours of observation at each of the four study sites Bonifacio (A), Curahuiche (B), South-Pilolcura (C) and North-Pilolcura (D); observations from all months are pooled. The vertical bars indicate standard deviation.

No counts were conducted during August and May due to bad weather. The greatest number of otters ($H_3 = 9.0$, $P = 0.03$) were recorded at Curahuiche, but no differences were recorded between the other study sites (see Table 1). Pups were recorded year-round, with most pups in spring, i.e. between September and November ($H_3 = 7.8$, $P = 0.05$). There was no seasonal variation in adults ($H_3 = 1.1$, $P = 0.78$). Using our technique for estimating otters, the highest abundance (i.e. the maximum number of otters recorded simultaneously by all three observers) was obtained quickest at the sites with the lowest estimated numbers of otters (Fig. 3). During summer (December-February), only four hours were needed to record the highest number at all study sites, whereas in spring (September-November), the asymptote was not reached after eight hours (Fig. 4). However, if the maximum number of otters counted (see Table 1) during this study had been visible at all time to any of the three

observers during the 1,960 otter counts, a total of 8,702 activity events would have been recorded instead of the 1,764 activity events actually recorded during the year. This suggests that only 20% of the daily activity of the estimated number of otters in our study sites were recorded. Therefore, during 80% of the observation periods, the otters were not observed either because they were out of the study site, obscured by a rock or inside a cave or rock crack. As observations were only carried out during eight hours of daylight, they represent only 7% of the total activity of the otters in a 24-hour schedule.

We recorded a total of 565 spraints, with significant differences observed between sites. Bonifacio and North-Pilolcura had significantly fewer spraints than Curahuiche (Wilcoxon: $P = 0.01$), and North-Pilolcura had fewer spraints than South-Pilolcura (Wilcoxon: $P = 0.04$). There were no significant differences in spraint number between Curahuiche and South-Pilolcura or Bonifacio

Table 2. Characterisation of habitats used by marine otters at the four study sites in Chile. Values in parentheses indicate the relative frequency (in %) of habitat characteristics within each study site and shelters.

Study site	Bonifacio	Curahuiche	South-Pilolcura	North-Pilolcura
Otter abundance	2.8	4.8	3.1	4.2
Number of transects	10	10	10	10
Rock size (study site)				
Small rocks	5 (45.5)	3 (27.3)	1 (9.1)	2 (18.2)
Medium rocks	1 (16.7)	2 (33.3)	2 (33.3)	1 (16.7)
Large rocks	4 (17.4)	5 (21.7)	7 (30.4)	7 (30.4)
Rock type (study site)	26 (25.0)	27 (26.0)	12 (11.0)	39 (38.0)
Reef rocks				
Caves and cracks	26 (16.7)	61 (38.9)	56 (35.7)	14 (8.9)
Number of Shelters	N = 9	N = 30	N = 30	N = 7
Rock type (shelter)				
Caves and cracks	2 (5.1)	27 (69.2)	8 (20.6)	2 (5.1)
Rocks without cracks	7 (18.9)	3 (8.1)	22 (59.5)	5 (13.5)
Rock size (shelter)				
Small rocks	23 (27.7)	13 (15.7)	6 (7.2)	42 (50.6)
Medium rocks	9 (4.4)	45 (22.1)	139 (68.1)	11 (5.4)
Large rocks	22 (13.7)	117 (72.7)	22 (13.7)	0
Total	54	175	177	53

and North-Pilolcura. Data pooled from all sites did not show a significant difference between seasons. However, there were significant differences in numbers of prey remains found between study sites: South-Pilolcura had the highest amount of prey remains (425 ± 44.9 ; SD) and North-Pilolcura the lowest (30 ± 4 ; SD). Both of these sites were significantly different to all the other sites (Wilcoxon: $P < 0.01$). Additionally, Curahuiche and Bonifacio had significantly fewer prey remains than South-Pilolcura (Wilcoxon: $P = 0.02$). There was no correlation between either the estimated number of otters or pups versus the amount of spraints and prey remains.

Terrestrial habitat characteristics

At all sites, > 40% of rocks were in the large category, and Bonifacio had the greatest proportion of small rocks (Table 2). A total of 76 otter shelters were registered from all study sites, which corresponded to areas with large rocks with caves and cracks (51%) and rocks without cracks (49%; see Table 2). There were no significant differences in the size or type of the rocks (i.e. with or without cracks or caves) used by marine otters between shelters (see Table 2). Different otters were observed using the same shelters on several occasions, and were also seen copulating in the water close to the shelters. During January 2000 four different male otters were live-trapped in the same shelters within 12 hours in Curahuiche (N = 2 otters) and North-Pilolcura (N = 2 otters).

Discussion

Most previous studies of marine otter abundance and activities were based on a census technique consisting of transect walks along a determined distance of shoreline (see Ebensperger & Castilla (1991) for a review). In such studies observation times varied between 0.5 and 2 hours/km, with an average of 1 hour/km of seashore (Ebensperger & Castilla 1991). If this methodology had been adapted for this study, approximately 221 activity events would have been recorded. This is equivalent to 12.5% of the actual number of recorded events from our study, without considering possible within-a-day or between-season variability. Consequently, as our observation methodology most probably represents only 7% of the total activity of the otters in a 24-hour schedule, previous studies using the transect methodology may be underestimating or overestimating the abundance of otters and their activity budget, as such observations are based on a sample size of only about 1% of total daily otter activity. Otters that hide due to the disturbance made by observers walking along the seashore may be missed, or otters may be counted more than once. Therefore, the technique we describe here is effective at providing samples of up to 10% of daily otter activity.

Future research on aspects of the behaviour, time activity budget, population variation and social interactions of this species should consider the use of radio-tracking as a practical method for obtaining more accurate data (e.g. Melquist & Hornocker 1979, Kruuk et al.

1989, Kruuk 1995, Sauer et al. 1999). Additionally, due to the spatial limits of the method presented in this paper, further applications should consider the stratification of shorelines based on the characteristics described in this paper, and shoreline sampling of the stratifications, in a way that allows extrapolation to a larger area of interest. The methods to estimate and compare otter activity described in our study are best considered as an index of minimum abundance, that may be better suited for monitoring populations rather than for surveying populations.

In our study the variation in the number of observable otters was 1-5 otters/km of seashore, with an average of 3.8. The asymptote curves from Figures 3 and 4 do not show cumulative data, rather the time and effort needed for the three observers to survey the maximum number of otters in a given place. The failure of most curves to reach an asymptote implies that the maximum number of otters was underestimated to an unknown extent. During the year in which our study was conducted, the number of otters did not vary between seasons. Therefore, it is likely that when otters were out of view, they were inside caves, cracks or obscured behind other terrestrial structures within the study sites. This is supported by the high daily variation in the maximum number of otters counted. Furthermore, we found that when otters are visible, they are most often feeding or moving between feeding sites. The values of feeding activity agree with values reported by Ostfeld et al. (1989) for Pan de Azúcar (56%) and Chiloé (54.7%), but they are smaller than those reported by Medina (1995b) in Curiñanco (63-70%), which is near to our study sites, and in Los Molles (83%) by Ostfeld et al. (1989). These observations indicate that activity data on marine otters that are analysed without considering the time during which otters are out of view, should be treated with caution, and are likely to be biased in favour of foraging.

Other studies have estimated the population abundance of otter in marine habitats by counting the number of otter holts (i.e. a tunnel system with signs of regular use by otters; van der Zee 1981, Kruuk et al. 1989), or in terrestrial habitats by assessing the concentration of spraints (Ebensperger & Castilla 1992, Sielfeld 1992). In our study it was demonstrated that because of the number of rock cracks, caves and galleries between rocks the latter procedure is not effective. Furthermore, there was no relation between the number of caves, and rock cracks and the number of otters counted. Our study also showed that the concentration of spraints and prey remains are not a reliable indicator of the number of otters or of aquatic habitat use and selection by the marine otter. Spraint concentrations differed only between sites with

very different number of otters (i.e. Cuarhuiche vs Bonifacio), and as previously described for Eurasian otter *L. lutra* in the sea, the best index for evaluating habitat selection and use is to directly count the otters and observe their behaviours and activities (Kruuk 1995). If sprainting serves as a proxy for the use of key resources (Kruuk 1992), we might expect that the accumulation of spraints in several rock caves and cracks in our study sites would indicate the presence of important resources for otters (Hutchings & White 2000). However, our study demonstrated that spraints and prey remains were only useful for determining and describing otter shelters and feeding sites. This lack of correlation between spraints or prey remains and resource availability may be explained by the fact that the marine otter lives on the Pacific coast, where it is exposed to wind and waves (Larivière 1998) which, at our study sites, broke harshly on the rocky shore. The sea continuously removes spraints and food remains from the inter-tidal region and reef rocks, which make sampling of spraint and prey remains incomplete for comparisons of terrestrial and aquatic habitat use by marine otters.

Although previous research has described the mating season for the marine otter to occur between December and January, and parturition to occur from January to March (see Larivière 1998 for a review), pups were recorded year-round in our study, with a peak between September and November. This discrepancy with previous observations could be the result of a lack of studies during different seasons and regions inside the distribution range of the species. This aspect of marine otter biology warrants further research and evaluation, given the great importance of seasonal pupping patterns for management and conservation concerns.

Previous studies have mentioned large rocks, rock caves and tree roots as important resources for the marine otter (Castilla & Bahamondes 1979, Ebensperger & Castilla 1992, Sielfeld 1990). However, our results did not demonstrate significant differences in habitat characteristics between study sites with different abundances of otters (see Table 2). Furthermore, there was no significant difference between the type of rocks for shelters; approximately half of them consisting of caves and cracks, and the other half consisting of tunnels between rocks without cracks (see Table 2). However, as recorded in our study, the marine otter apparently spends a great time of the day on the coast obscured inside caves or rock cracks. This could be a strategy for temperature conservation. Therefore, we urgently recommend that future studies should compare sites from different regions, in order to gain a better understanding of the importance of terrestrial habitat as it pertains to the aquatic living

condition of the marine otter, and to determine whether the distribution of the species is limited when human settlers and domestic dogs utilise the terrestrial habitats.

Acknowledgements - we would like to thank the Frankfurt Zoological Society, the Comité Nacional Pro Defensa de la Fauna y la Flora and the International Otter Survival Fund for financial support and provision of equipment. We are grateful to Mr. Fernando Jara for help in statistical analyses, Drs. J.L. Bodkin and Glenn R. VanBlaricom for reviewing the manuscript and making important contributions. We thank Paula Emily Neill for proofreading the manuscript and Mr. René Monsalve for his help during the field work.

References

- Castilla, J.C. 1982: Nuevas observaciones sobre conducta, ecología y densidad de *Lutra felina* (Molina 1782) (Carnívora: Mustelidae) en Chile. - Publicación Ocasional Museo Nacional de Historia Natural, Santiago, Chile, 38: 197-206. (In Spanish).
- Castilla, J.C. & Bahamondes, I. 1979: Observaciones conductuales y ecológicas sobre *Lutra felina* (Molina) 1782 (Carnívora: Mustelidae) en las zonas central y centro-norte de Chile. - Archivos de Biología y Medicina Experimental, Santiago, 12:119-132. (In Spanish).
- Ebensperger, L. & Castilla, J.C. 1991: Conducta y densidad poblacional de *Lutra felina* en Isla Pan de Azúcar (III Región), Chile. - Medio Ambiente 11: 79-83. (In Spanish).
- Ebensperger, L. & Castilla, J.C. 1992: Selección de hábitat en tierra por la nutria marina *Lutra felina*, en Isla Pan de Azúcar, Chile. - Revista Chilena de Historia Natural 65: 429-434. (In Spanish).
- Estes, J.A. 1989: Adaptations for aquatic living by carnivores. - In: Gittleman, J.L. (Ed.); Carnivore Behavior Ecology, and Evolution. Cornell University Press, New York, pp. 242-282.
- Estes, J.A., Jameson, R.J. & Rhode, E.B. 1982: Activity and prey election in the sea otter: influence of population status on community structure. - The American Naturalist 120: 242-258.
- Gese, E.M. 2001: Monitoring of terrestrial carnivore populations. - In: Gittleman, J.L., Strehman, M.F., Macdonald, D. & Wayne, R.K. (Eds.); Carnivore Conservation. Cambridge University Press, Cambridge, pp. 272-396.
- Hutchings, M.R. & White, P.C.L. 2000: Mustelid scent-marking in managed ecosystems: implications for population management. - Mammal Review 30: 157-169.
- Kruuk, H. 1992: Scent marking by otters (*Lutra lutra*): signalling the use of resources. - Behavioural Ecology 3: 133-140.
- Kruuk, H. 1995: Wild Otters Predation and Population. - Oxford University Press, Oxford, 290 pp.
- Kruuk, H., Moorhouse, A., Conroy, J.W.H., Durbin, L. & Frears, S. 1989: An estimate of numbers and habitat preferences of otters (*Lutra lutra*) in Shetland, UK. - Biological Conservation 49: 241-254.
- Larivière, S. 1998: Lontra felina. - Mammalian Species 575: 1-5.
- Medina, G. 1995a: Activity budget and social behaviour of marine otter (*Lutra felina*) in southern Chile. - Proceedings of the VI International Otter Colloquium Pietermaritzburg, Habitat 11: 62-64.
- Medina, G. 1995b: Feeding habits of marine otter (*Lutra felina*) in southern Chile. - Proceedings VI International Otter Colloquium Pietermaritzburg, Habitat 11: 65-68.
- Medina-Vogel, G., Delgado, C.R., Alvarez, R.E.P. & Bartheld, J.L.V. 2004: Feeding ecology of the marine otter (*Lontra felina*) in a rocky seashore of the south of Chile. - Marine Mammal Science 20: 134-144.
- Melquist, W.E. & Hornocker, M.G. 1979: Development and use of a telemetry technique for the study river otter. - Conference on Wildlife Biotelemetry 2, Laramie, Wyoming, pp. 104-114.
- Moreno, C., Sutherland, J. & Jara, F. 1984: Man as predator in the intertidal zone of southern of Chile. - Oikos 42: 155-160.
- Ostfeld, R., Ebensperger, L., Klosterman, L. & Castilla, J.C. 1989: Foraging, activity budget and social behavior of the South American marine otter *Lutra felina* (Molina 1782). - National Geographic Research 5: 422-438.
- Parera, A. 2002: Los mamíferos de la Argentina y la región austral de Sudamérica. - Editorial El Ateneo, Buenos Aires, República Argentina, 453 pp. (In Spanish).
- Reuther, C., Dolch, D., Green, R., Jahrl, J., Jefferies, D., Krekemyer, A., Kucerova, M., Madsen, A.B., Romanowski, J., Roche, K., Ruiz-Olmo, J., Teubner, J. & Trinidad, A. 2000: Surveying and monitoring distribution and population trends of the Eurasian otter (*Lutra lutra*). Guidelines and evaluation of the standard methods for surveys as recommended by the European Section of the IUCN/SSC Otter Specialist Group. - Habitat 12: 1-148.
- Rozzi, C. & Torres-Mura, J.C. 1990: Observaciones del chungungo (*Lutra felina*) al sur de la Isla Grande de Chiloé: Antecedentes para su conservación. - Medio Ambiente 11: 24-28. (In Spanish).
- Sauer, T.M., Ben-David, M. & Bowyer, R.T. 1999: A new application of the adaptive-kernel method: estimating linear home ranges of river otters, *Lutra canadensis*. - Canadian Field-Naturalist 113(3): 419-424.
- Shimek, S. & Monk, A. 1977: Daily activity of sea otter off the Monterrey Peninsula, California. - Journal of Wildlife Management 41: 277-283.
- Sielfeld, W. 1990: Características del hábitat de *Lutra felina* (Molina) y *L. provocax* (Thomas) en Tierra del Fuego-Patagonia. - Investigación Científicas y Técnicas, Serie: Ciencias de Mar 1: 30-36. (In Spanish).
- Sielfeld, W.K. 1992: Abundancias relativas de *Lutra felina* (Molina, 1782) y *L. provocax* (Thomas), 1908 en el litoral

- de Chile austral. - Investigaciones Científicas y Técnicas, Serie Ciencias del Mar 2: 3-11. (In Spanish).
- Sokal, R.R. & Rohlf, F.J. 1981: Biometry, the principles and practice of statistical in biological research. 2nd edition. - W.H. Freeman, New York, 887 pp.
- van der Zee, D. 1981: Density of cape clawless otter *Aonyx capensis* (Schinz, 1821) in the Tsitsikama coastal National Park. - South African Journal of Wildlife Research 12: 8-13.
- Zar, J. 1996: Biostatistical Analysis. 3rd edition. - New Jersey, Prentice-Hall International, Inc., 662 pp.