



Environmental and Biotic Determinants of Growth and Height of Arctic Willow Shrubs along a Latitudinal Gradient

Author: Pajunen, Anu Marjukka

Source: Arctic, Antarctic, and Alpine Research, 41(4) : 478-485

Published By: Institute of Arctic and Alpine Research (INSTAAR),
University of Colorado

URL: <https://doi.org/10.1657/1938-4246-41.4.478>

BioOne Complete (complete.BioOne.org) is a full-text database of 200 subscribed and open-access titles in the biological, ecological, and environmental sciences published by nonprofit societies, associations, museums, institutions, and presses.

Your use of this PDF, the BioOne Complete website, and all posted and associated content indicates your acceptance of BioOne's Terms of Use, available at www.bioone.org/terms-of-use.

Usage of BioOne Complete content is strictly limited to personal, educational, and non - commercial use. Commercial inquiries or rights and permissions requests should be directed to the individual publisher as copyright holder.

BioOne sees sustainable scholarly publishing as an inherently collaborative enterprise connecting authors, nonprofit publishers, academic institutions, research libraries, and research funders in the common goal of maximizing access to critical research.

Environmental and Biotic Determinants of Growth and Height of Arctic Willow Shrubs along a Latitudinal Gradient

Anu Marjukka Pajunen*

*Arctic Centre, University of Lapland,
Box 122, FI-96101 Rovaniemi, Finland
anu.pajunen@oulu.fi

Abstract

Ecological factors determining the growth of arctic shrubs remain poorly understood, thereby obscuring the current predictions about climate change effects. I conducted a study on the Yamal Peninsula, West Siberia, to find out which factors determine the growth and height of upright willows (*Salix glauca* and *S. lanata*). I sampled willow thickets at different slope positions at 13 sites along a 300-km-long north-south transect. The measurements included the height of willow shrubs and the length and diameter of shoots.

The length and diameter of willow shoots increased southwards and with increasing distance from the sea. At the top of the slope and in areas with shallow thaw, the height and growth of willow were low compared to other slope positions and to areas with deep thaw. An increasing level of reindeer (*Rangifer tarandus*) grazing intensity was associated with decreased height and growth of willow and also associated with reduced foraging activity of willow grouse (*Lagopus lagopus*).

The results show that even though increasing summer warmth is likely to enhance willow growth, there are other factors such as distance from the sea that also affect the growth and height of willow. The results suggest that reindeer grazing may locally counteract the effects of climate change.

DOI: 10.1657/1938-4246-41.4.478

Introduction

In the southern parts of arctic environments, shrub communities are one of the most prominent types of vegetation and therefore play an important role in ecosystem functioning (Chernov, 1985; Bliss and Matveyeva, 1992; Blanken and Rouse, 1994). According to recent aerial photo surveys, the abundance of tundra shrubs in northern Alaska has considerably increased in recent decades apparently due to climate change (Sturm et al., 2001; Tape et al., 2006). There is also evidence showing that the experimental warming of air temperature favors growth of shrubs at the expense of other tundra plants (Chapin et al., 1995; Walker et al., 2006). The ongoing and predicted climatic warming in the Arctic (ACIA, 2005; IPCC, 2007) is likely to affect growth patterns of shrubs directly and indirectly through interactions with other plants. However, in the current predictions, other environmental and biotic growth determinants of shrub growth have not sufficiently been taken into account, and they still remain poorly understood (see also Post and Pedersen, 2008).

A reduction of woody plant communities along increasing altitude and latitude is determined by summer temperatures and the length of the growing season. The radial growth of trees and shrubs correlate with mean summer temperatures (Walker, 1987; Knorre et al., 2006), even though other factors such as delayed snowmelt in the spring may interfere with this relationship (Kirdyanov et al., 2003; Schmidt et al., 2006). Warm air and soil temperature directly enhance the metabolism rate and the photosynthetic activity of plants, and have indirect effects through changes in soil resources such as water and nutrient availability and the depth of annual thaw (Nadelhoffer et al., 1991; Jarvis and Linder, 2000). These direct and indirect effects of temperature on vegetation are highly species and site specific (Hobbie and Chapin,

1998; Bret-Harte et al., 2002) and in most cases still remain inadequately known (Jarvis and Linder, 2000). In the Eurasian Arctic, the most common species in the shrub communities are grey-leaved willows, *Salix glauca* and *S. lanata*, the growth patterns of which remain unstudied (see, however, Walker, 1987).

Permafrost in tundra areas affects plants not only through changes in available rooting space but also through changes in hydrological, chemical, and thermal conditions of soils. Willows and other deciduous shrubs may effectively utilize areas with discontinuous or deep permafrost, where they have plenty of rooting space, available nutrients, and good drainage (Schickhoff et al., 2002; Lloyd et al., 2003; Walker et al., 2003; Schuur et al., 2007). Therefore, if the observed warming of permafrost (Osterkamp and Romanovsky, 1999; Pavlov and Moskalenko, 2002) will lead to degradation of permafrost and deepening of the active layer as predicted (ACIA 2005; Anisimov and Reneva, 2006), then shrub abundance is likely to increase. However, in predicting vegetation changes in relation to permafrost, local conditions including vegetation type, small-scale topography, snow conditions, and soil characteristics should be taken into account as they are likely to complicate the relationship (Smith and Riseborough, 1996; Walker et al., 2003).

The vertical structure of arctic willow vegetation along toposequences remains largely undocumented (see, however, Edlund and Egginton, 1984; Schickhoff et al., 2002). Despite the latitude of the overall area, wind and snow patterns create microclimate variability along slopes which is likely to affect the growth of shrubs (Seppälä, 2004). In addition, soil conditions and nutrient availability vary along toposequences thereby affecting the growth of plants. In their turn, shrubs may also change the conditions of slopes, for instance by increasing the trapping of

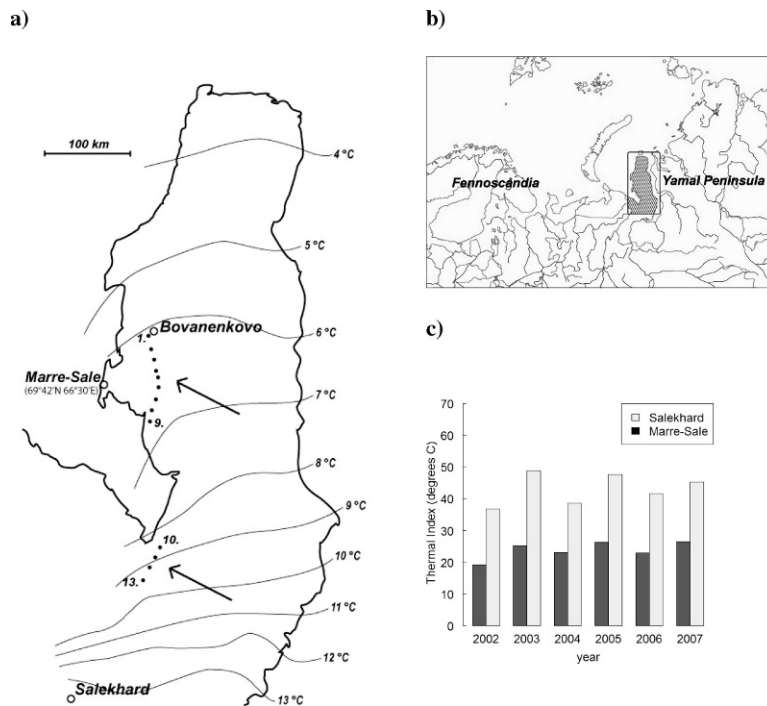


FIGURE 1. (a and b) The study was conducted along a latitudinal transect across the Yamal Peninsula, northwestern Russia. The isotherms of July temperatures are redrawn from Shiyatov and Mazepa (1995). The arrows indicate the direction of the migration routes of Nenets reindeer herders. (c) The thermal sum indexes (sum of positive monthly mean temperatures May–September, °C) of the weather stations in the vicinity of the study transects are presented during the years 2002–2007.

snow (Liston et al., 2002; Sturm et al., 2005) and enhancing the turnover of nutrients (Epstein et al., 2004). On a broader geographical scale, the distance from the sea is likely to affect the growth of shrubs due to strong winds often associated with areas near arctic seas (Seppälä, 2004).

In the Eurasian Arctic, the reindeer (*Rangifer tarandus*) is the most important herbivore affecting the growth of willow (Chernov, 1985). Willows are of special importance for reindeer at the beginning of the growing season when herbs and graminoids are scarce and reindeer suffer from protein and mineral deficiency after the winter (Vakhtina, 1964). Reindeer have a tendency to suppress the growth of their forage shrubs and fragment their distribution in a landscape (den Herder et al., 2004, 2008; Manseau et al., 1996; Pajunen et al., 2008). This may have effects on other herbivores such as willow grouse (*Lagopus lagopus*) that feeds on willow during the winter (Chernov, 1985; Ims et al., 2007; den Herder et al., 2008). The effects of reindeer grazing on the growth of willow are likely to depend on abiotic factors such as climate and soil nutrient availability (Eskelinen, 2008; Post and Pedersen, 2008), and yet, these interactions remain to a large extent unknown.

I conducted a field study in the central and southern Yamal Peninsula in order to find out which environmental factors determine the growth and height of willow shrubs in a tundra region of West Siberia that is extensively grazed by large herds of reindeer. Particularly, I addressed the following questions: (1) What is the relative importance and mutual dependency between climate, soil thaw, slope position, and distance from the sea for the height and growth of willow shrubs in Arctic conditions? (2) How does reindeer grazing affect the height and growth of willow and how do the effects of reindeer interact with the above-mentioned abiotic factors and herbivory by willow grouse?

Study Area

I sampled willow thickets along a latitudinal transect spanning from the central Yamal Peninsula to the northern Polar Urals (Fig. 1). The total length of the transect was over 300 km.

The sampling transect includes a gap in the southern Yamal Peninsula, because field work had to be restricted to sites that were accessible during the summers of 2006 and 2007. Therefore, the transect consists of two parts: (1) an approximately 125-km-long section in the more northern area (sites 1–9), and (2) an approximately 30-km-long section in the more southern area (sites 10–13) (Fig. 1). The transects followed the only two available access routes into the region, an all-terrain vehicle trail in the more northern area and a road in the more southern area.

The mean July temperature at the northernmost study site is 6 °C and increases approximately 1 °C/100 km in a southward direction. Along the western coast of the peninsula, the July isotherms curve southwards due to the cooling effect of the sea (Shiyatov and Mazepa, 1995) (Fig. 1). The mean annual temperature in the more northern area (Marre-Sale weather station) is –8.4 °C and in the more southern area (Salekhard weather station) –6.0 °C. The mean annual precipitation is 289.0 mm and 494.8 mm, and the mean snow depth is 21.4 cm and 35.6 cm, respectively, at more northern and more southern areas (<http://meteo.infospace.ru>). The thermal sum indexes (sum of positive monthly mean temperatures May–September, °C) of the study areas during the years 2002–2007 are presented in Figure 1c.

The study areas have undergone Pleistocene glaciation, although details of glaciation history are unknown (Svendsen et al., 1999). In the western Yamal Peninsula, the oldest exposed sediments are laminated clayed silts and silty clays (Astakhov et al., 1996), which are overlain by the Kara till which is up to 20 m thick (Gataullin, 1988). These sediments are mantled by Baydarata sandy silt. This can be partly regarded as loess, analogous to the Yedoma formation of East Siberia and Alaska (Astakhov et al., 1996). This sedimentary structure is modified by various Holocene formations, such as limnic fills of thermokarst sinkholes and fluvial sediments. Podzols are normally nonexistent, but some signs of leaching occur on well-drained hills. These sites can have fairly acidic soils and frost boils also occur.

The entire sampling area is underlain by continuous permafrost; the annual depth of thaw varies between 0.3 and 1.7 m depending on temperature regimes at variable levels above

and below ground. Therefore the depth of annual thaw is also affected by soil and snow conditions and vegetation (Pavlov and Moskalenko, 2002). In the Russian classification scheme (Matveyeva, 1998), Sites 1–9 belong to the typical tundra zone, whereas sites 10–13 belong to the southern tundra (Fig. 1). In another classification scheme, both of the tundra zones belong to the Low Arctic vegetation zone (Walker et al., 2005, and sources cited there). Typically, the abundance and height of shrubs is greater in the southern than in the typical tundra zone (Walker et al., 2005).

The number of semi-domesticated reindeer has increased dramatically during the past decades on the Yamal Peninsula (Baskin, 2005). The number of reindeer in the whole Yamal region (500,000 km²) was approximately 630,000 animals in 2007 (http://www.uralpolit.ru/regions/jamal/28-06-2007/page_63081.html). Nenets nomads migrate with their animals through the study areas in several herding units (brigades) twice a summer, moving north in mid-June and back south in mid-August). When a reindeer herd moves through an area the number of animals (approximately 3000 animals occupying 100 km²) may for several days approach 30 animals km⁻² in a location (Stammler, 2005). The direction of the migration routes of Nenets reindeer herders is indicated in Figure 1.

Data Collection

The data was collected in the more northern area (sites 1–9) during the period 18 July–1 August 2006 and in the more southern area (sites 10–13) during the period 19–25 July 2007. The criteria for selection of sampling locations were easily met along the only two available access routes into the region. In both of the areas, sampling was restricted to locations that had a slope of a watershed growing willow shrubs within walking distance from a road/all-terrain vehicle trail. In the more northern area, the distance between sampling locations ranged from 5 to 27 km, whereas in the more southern area, the distance between sampling locations ranged from 7 to 15 km. The study site was the nearest slope of a watershed from a road/all terrain vehicle trail, where willow shrubs were growing.

At each study site three distinct topographical parts based on vertical position and the steepness of the slope were selected. (1) The uppermost position (hereafter called top) was at the lowest flat part of the top, under which the angle of the slope began to grow. This part of the slope was characterized by tundra heath vegetation types (Andreyashkina and Peshkova, 1995). The species composition of these tundra heath vegetation types varied depending on the moisture regime, but typically bryophytes, lichens, graminoids, and dwarf shrubs dominated. Low shrubs (mainly *Salix glauca* and *Betula nana*) occurred sporadically. The cover of willow was on average 16%. (2) The mid-slope position, (hereafter called middle), represented the steepest part of the slope. This part of the slope was characterized by shrub vegetation types (Andreyashkina and Peshkova, 1995) with ubiquitous willow cover (average cover of willow 52%). At this slope position, the understorey of vegetation was relatively sparse constituting mainly of horsetails (*Equisetum arvense*) and forbs (e.g., *Polemonium acutiflorum*, *Rubus arcticus*, *Stellaria peduncularis*) and sporadically occurring bryophytes (e.g. *Sanionia uncinata*). Towards the more southern study sites, willows formed a thicket with *Betula nana*. (3) The down slope position (hereafter called bottom) was near the end of the steep part of the slope, but on level or nearly level ground. At this part of the slope the vegetation type was either a willow thicket, as in the middle slope, or a mire (Andreyashkina and Peshkova, 1995) that was dominated by tall willows in the upper layer of vegetation and by *Sphagnum* mosses

or sedges (*Carex* spp.) in the understorey of vegetation. The cover of willow in the down slope was on average 58%. This part of the slope was restricted by a river, lake, or mire in the down slope.

In each part of the study site (top, middle and bottom), 1–3 study plots 10 × 10 m in size were chosen by using a randomization procedure. To ensure that willow vegetation was present at the top sites, the sampling was restricted to sites with an overall cover of willow of at least 10%. In the middle and bottom sites, the cover of willow was always high and sampling was performed at all sites.

In each study plot, the height of 10 random willow ramets per species (*Salix glauca* or *S. lanata*) representing the prevalent uppermost canopy was measured from the ground. The top shoots with at least four years of growth increments of the measured individuals were collected for later retrospective growth measurements.

The top shoots collected in 2006 were dried and measured retrospectively in the laboratory during winter 2007. The top shoots collected in 2007 were measured in the field. In the retrospective measurements, the length of shoots was measured starting from the leading shoot and proceeding annually back to the shoot that had been growing in 2003. The living shoot and total length of the shoots were measured separately and the frost- or herbivore-damaged shoots were marked. For the analyses of length of shoots, only the measurements from non-damaged shoots were used. If the damage had occurred in later years and the actual growth for a specific year could be seen from a dead, but undamaged shoot, the result was included in the analyses. The shoot diameter was measured using cross measurements in 2003. The number of feeding marks by willow grouse (*Lagopus lagopus*) was counted annually between years 2003 and 2005.

The number of reindeer faeces was counted along a side of one randomly chosen plot (10 × 2 m) at each part of a study site (top, middle, bottom), and the mean value of these three transects was used in the analyses as a representative figure for the grazing intensity for that site. Faeces from previous years could not be differentiated. In addition, the depth of thaw (five measurements) was measured within each study plot.

Statistical Analyses

Prior to statistical analyses, the mean value of each response variable—the height of willow shrubs and the length and diameter of willow shoots—was calculated for each plot in order to avoid pseudoreplication. The explanatory variables were latitude (distance from the Equator [km]), distance from the sea (km), position on the slope (“top,” “middle,” “bottom”), depth of thaw (cm), the mean number of reindeer faeces on the study slope, the number of willow grouse feeding marks per ramet, and the identity of *Salix* species (*S. lanata* or *S. glauca*). In addition, a year (2003, 2004, 2005, or 2006) was included as an explanatory variable in the model for the length of willow shoots.

Linear mixed effects (LME) models (Pinheiro and Bates, 2000; Crawley, 2002) were used to investigate the relationships between response and explanatory variables. Since the plots within a given site were expected to be more similar to each other than to random plots drawn from the entire data set, a study site (1–13) was set as a random factor, whereas all the explanatory variables mentioned above were set as fixed factors. The quadratic terms of all the continuous explanatory variables were included as explanatory variables in the initial models. The modeling procedure followed the suggestion for multiple regression models by Crawley (2002). All the possible one-way interactions were

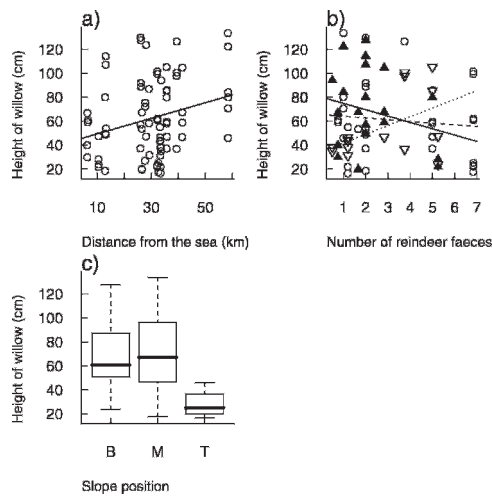


FIGURE 2. The height of willow ramets (cm) in relation to (a) the distance from the sea (km), (b) the number of reindeer faeces, and (c) the position on the slope. The lines (a)–(b) represent a fitted value for each relationship. The relationships between all these illustrated explanatory variables (and their interactions) and willow height were significant ($p < 0.05$) in the LME models (see text for details). In Figure 2b, the black triangle marks and the solid line indicate the plots that are characterized by deep thaw (thaw depth above or equal to 76.8 cm), and the open circles and the dashed line indicate the plots with middle thaw depth ($53.3 \text{ cm} \leq \text{thaw depth} \leq 76.7 \text{ cm}$). The open triangles and the dotted line indicate the plots with low thaw (thaw depth less than or equal to 53.2 cm). In the boxplot Figure 2c, the middle line represents the median, and the other two lines the lower and upper quartiles. The extreme values are marked as separate dots if they go beyond 1.5 times the interquartile, which is the maximum length of the line. The abbreviation B indicates bottom; M middle; and T the top part of the slope.

included in the initial models, whereas two-way interactions were excluded. To obtain minimum adequate models (Crawley, 2002), non-significant terms were removed, and final models contained only significant explanatory variables or their interaction terms. In order to obtain satisfactory residual plots of the models, the data of length and diameter of willow shoots were log-transformed prior to the modeling. To graphically illustrate the results, linear or curvilinear regression models with only one variable were used to draw the prediction lines for each variable. In order to graphically illustrate the interaction between two continuous explanatory variables, some of the variables were split into three groups by subtracting the smallest value from the largest and dividing the figure by three. In this way, the response of a continuous variable could be plotted at different levels of a grouped variable. The grouping classes are described in the legends for Figures 2–4. The statistical analyses were carried out using the R statistical environment (R Development Core Team, 2005).

Results

The **height of willow ramets** linearly increased as a response to increasing distance from the sea (sea: $F_{1,46} = 4.9$, $p = 0.03$; Fig. 2a). The effect of reindeer grazing on the height of willow ramets depended on the depth of thaw (grazing \times thaw: $F_{1,46} = 4.6$, $p = 0.04$). In the areas with deep or middle thaw, the increase in the grazing intensity decreased the height of willow ramets, whereas at low thaw depths, the increase in the grazing intensity increased the height of willow ramets (Fig. 2b). Willow ramets

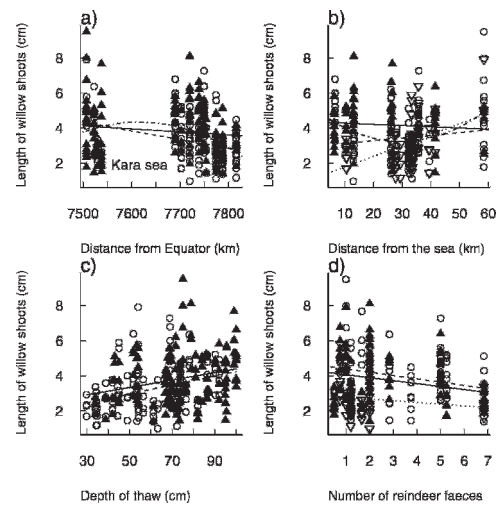


FIGURE 3. The length of willow shoots (cm) in relation to (a) the distance from the Equator (km), (b) the distance from the sea (km), (c) the depth of thaw (cm), and (d) the number of reindeer faeces. The lines (a)–(d) represent fitted value for each relationship. The relationships between all these illustrated explanatory variables (and their interactions) and the length of willow shoots were significant ($p < 0.05$) in the LME models (see text for details). In Figures 3a and 3c, the triangle plots and the solid line indicate *Salix lanata* and the open circles and the dashed line indicate *S. glauca*. The dot-dash line indicates the quadratic term of distance from the Equator. In Figure 3b, the black triangle marks and the solid line indicate the plots that have deep thaw (thaw depth above or equal to 76.8 cm), and the open circles and the dashed line indicate the plots with middle thaw depth ($53.3 \text{ cm} \leq \text{thaw depth} \leq 76.7 \text{ cm}$). The open triangles and the dotted line indicate the plots with low thaw (thaw depth less than or equal to 53.2 cm). The dot-dash line indicates the quadratic effect of distance from the sea. In the Figure 3d, the black triangle marks and the solid line indicate the plots at the bottom of the slope, the open circles and the dashed line plots at the middle of the slope, and the open triangles and the dotted line plots at the top of the slope.

were highest in the middle of the slope and lowest at the top of the slope (slope: $F_{2,46} = 12.8$, $p < 0.0001$, Fig. 2c). The effects of explanatory variables did not interact with the identity of *Salix* species, but the difference of average ramet length of *Salix lanata* ($74.1 \pm 32.6 \text{ cm}$) and *S. glauca* ($45.2 \pm 22.3 \text{ cm}$) was consistent (species: $F_{1,46} = 19.6$, $p < 0.0001$).

The **length of willow shoots** curvilinearly decreased northwards (latitude $\wedge 2$: $F_{1,8} = 6.4$, $p = 0.04$; Fig. 3a). The effect of latitude on the length of willow shoots also varied between *Salix* species (latitude \times species: $F_{1,226} = 10.0$, $p = 0.002$): the length shoot of *Salix glauca* declined more strongly northwards than the length shoot of *S. lanata* (Fig. 3a). There was a significant concave curvilinear relationship between the distance from the sea and the length of willow shoots (sea $\wedge 2$: $F_{1,226} = 7.0$, $p = 0.009$). From the coast until approximately 30 km inland, the curve showed a decline in the length shoot as a response to increasing distance from the sea, whereas from 30 km onwards, the curve showed an increase in length shoot as a response to increasing distance from the sea (Fig. 3b). The effect of distance from the sea on the length of willow shoots depended on the depth of thaw (sea \times thaw: $F_{1,226} = 5.7$, $p = 0.02$). At shallower thaw depths, where the shoot length was smaller than at the greater thaw depths, there was a strong increase in shoot length as a response to increasing distance from the sea (Fig. 3b). In the areas of middle thaw depth, the length of willow shoots slightly increased as a response to increasing distance from the sea, and in the areas with deep thaw,

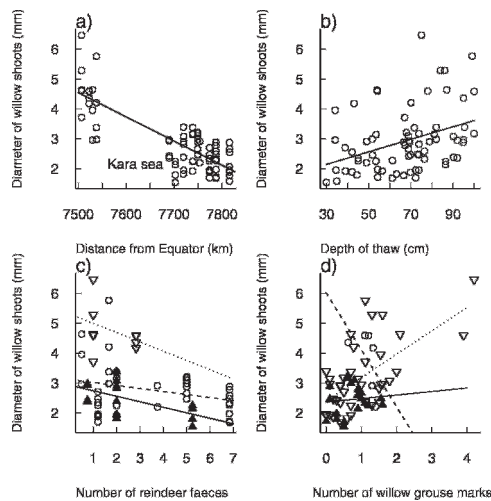


FIGURE 4. The diameter of willow shoots (mm) in relation to (a) the distance from the Equator (km), (b) the depth of thaw (cm), (c) the number of reindeer faeces, and (d) the number of willow grouse feeding marks. The lines (a)–(d) represent the fitted value for each relationship. The relationships between all these illustrated explanatory variables (and their interactions) and the diameter of willow shoots were significant ($p < 0.05$) in the LME models (see text for details). In Figure 4c, the black triangles and the solid line indicate the plots close to the sea (23.7 km or less from the sea) and the open circles and the dashed line indicate the plots with middle distance from the sea ($23.8 \text{ km} \leq \text{distance from the sea} \leq 41.2 \text{ km}$). The open triangles and the dotted line indicate the plots far away from the sea (41.3 km or more from the sea). In Figure 4d, the black triangle marks and the solid line indicate high reindeer pressure (the mean number of faeces greater than or equal to 4.9), the circles and the dashed line indicate medium reindeer pressure ($2.7 \leq \text{the mean number of faeces on the slope} \leq 4.8$), and the open triangles and the dotted line indicate low reindeer pressure (the mean number of faeces on the slope less than or equal to 2.6).

the shoot length remained nearly constant irrespective of the distance from the sea (Fig. 3b). The length of willow shoots linearly increased as a response to increasing depth of thaw. For *Salix glauca* this effect was slightly stronger than for *S. lanata* (thaw \times species: $F_{1,226} = 3.8$, $p = 0.05$) (Fig. 3c). The length of willow shoots was negatively related to reindeer density, and this effect depended on the position on the slope (grazing \times slope: $F_{2,226} = 6.0$, $p = 0.003$). At the middle and the bottom of the slope, where the shoot length was greater than at the top of the slope, the negative effect of grazing was stronger than at the top of the slope (Fig. 3d). In 2005, the length of willow shoots was smaller ($2.8 \pm 1.3 \text{ cm}$) than in the other investigated years (2003: $3.8 \pm 1.3 \text{ cm}$; 2004: $3.8 \pm 1.4 \text{ cm}$; 2006: $3.9 \pm 1.5 \text{ cm}$) (year: $F_{3,226} = 22.5$, $p = 0.0001$).

The diameter of willow shoots in 2003 decreased northwards (latitude: $F_{1,9} = 451.0$, $p < 0.0001$) (Fig. 4a) and increased as a response to increasing depth of thaw (thaw: $F_{1,46} = 24.3$, $p < 0.0001$) (Fig. 4b). The increasing intensity of reindeer grazing decreased the diameter of willow shoots and this effect depended on the distance from the sea (grazing \times sea: $F_{1,46} = 22.5$, $p < 0.0001$). In the areas far away from the sea, where the diameter of willow shoots was greatest, the response of shoot diameter to increasing grazing pressure was more clear than in the areas closer to the sea. The effect of willow grouse on the shoot diameter of willow depended on the intensity of reindeer grazing (grazing \times willow grouse: $F_{1,46} = 6.8.3$, $p = 0.01$). In the areas of high or low intensity of reindeer grazing, shoot diameter increased with

increasing number of willow grouse feeding marks, whereas in the areas of medium intensity of grazing, there was a negative association between shoot diameter and willow grouse feeding (Fig. 4d). The diameter of willow shoots was smaller at the top of the slope ($2.3 \pm 0.7 \text{ mm}$) than at the other two slope positions (middle: $3.0 \pm 1.1 \text{ mm}$; bottom: $3.1 \pm 1.0 \text{ mm}$) (slope: $F_{2,46} = 27.6$, $p < 0.0001$). The effects of explanatory variables did not interact with the identity of *Salix* species, but the difference between shoot diameter of *Salix lanata* ($3.4 \pm 1.0 \text{ cm}$) and *S. glauca* ($2.3 \pm 0.7 \text{ cm}$) was constant (species: $F_{1,46} = 167.3$, $p < 0.0001$).

Discussion

ENVIRONMENTAL DETERMINANTS OF WILLOW GROWTH

The observed associations between latitude and distance from the sea and the primary (the length of shoots) and the secondary (the diameter of shoots) growth of willow suggest a strong correlation between willow growth and the July isotherm. This result is in line with dendrochronological and experimental studies showing that an increase in mean summer temperatures causes an increase in shrub abundance and radial growth (Walker, 1987; Knorre et al., 2006; Walker et al., 2006). The positive correlation between the height of willow and increasing distance from the sea supports earlier observations according to which most abundant arctic willow communities are found in continental areas (Walker, 1987; Bliss and Matveyeva, 1992). As the suppressing effect of the cold sea on the primary growth of willow was stronger in the areas with shallow thaw, the results suggest that several factors may simultaneously reduce willow growth. The low primary growth in 2005 compared to the other investigated years possibly resulted from a relatively small thermal index value in 2004 compared to the other investigated years (Fig. 1c).

The primary and secondary growth of willow was strongly associated with increasing depth of thaw, which is determined by temperature regimes at variable levels and is also affected by soil and snow conditions and vegetation (Smith and Riseborough, 1996). This result conforms to earlier reports according to which most of the arctic willow communities are found in southern tundra areas, where there is either discontinuous permafrost or the active layer is deep (Bliss and Matveyeva, 1992). Also, in Alaskan riparian habitats, the height of willows decreased along topographical gradients, along which the depth of thaw decreased (Schickhoff et al., 2002). According to the results found here the depth of thaw explains the growth of willow more clearly than the direct effect of latitude. This suggests that willow is strongly dependent on rooting space and on the nutrients and water that are more available in soils characterized by deep thaw.

The height of willow, as well as the length and diameter of willow shoots, were lowest at the top of the slope. This may be a result of wind-driven snow accumulation at the bottom parts of slopes, which would support the traditional view according to which snow depth determines the height of willows in arctic and subarctic environments (Chernov, 1985). On the other hand, at the top part of slopes there tends to be low nutrient availability and a reduction in soil moisture that may reduce the height and growth of willow and favor other plants in competition with willow. In this case, the snow trapping by shrubs would cause the accumulation of snow at the bottom parts of slopes instead of the other way around (Seppälä, 2004). The positive feedback between shrub growth and accumulation of snow (Liston et al., 2002; Sturm et al., 2005) increases the insulation of soils in winter,

thereby retarding the freezing process and leading to higher winter ground temperatures than in adjacent shallow snow sites (Seppälä, 2004). This, in addition to the steepness of slope that enhances drainage, promotes the development of deep thaws, which further promotes the growth of willows (Liston et al., 2002; Sturm et al., 2005).

THE ROLES OF REINDEER GRAZING AND WILLOW GROUSE

The observed negative association between the height and growth of willow and intensity of reindeer grazing is consistent with earlier studies in Fennoscandia (den Herder et al., 2004, 2008; Pajunen et al., 2008; Kitti et al., 2009) and in Canada (Ouellet and Boutin, 1994; Manseau et al., 1996). It is likely that if the sampling had not been restricted to mature willows the effects would have been stronger still since willow seedlings tend to be especially vulnerable to grazing (den Herder and Niemelä, 2003).

The observed interactions between grazing and abiotic variables suggest that the effects of reindeer grazing are strongest in the areas where abiotic factors favor willow growth. In the areas of deep thaw the increase in the grazing intensity decreased the height of willow, whereas at the shallower thaw depth, the opposite held true. Similarly, at the top position of the slope and near the coast the effects of grazing on the length and diameter of willow shoots were smaller than at the two lower slope positions and in the areas further away from the coast. The results found here resemble experimental studies, where the effects of reindeer have been greatest in the warmed conditions (Post and Pedersen, 2008) and in nutrient-increased sites (Gough et al., 2007; Eskelinen, 2008). Similarly as found experimentally by Post and Pedersen (2008), the results found here imply that the effects of reindeer grazing may locally counteract the effects of climate change on the abundance of shrubs (Walker et al., 2006).

The grazing by willow grouse for the most part did not reduce the growth or abundance of willow. However, in the areas with low or medium intensity of reindeer grazing, the feeding activity of willow grouse was generally higher than in the areas of intensive grazing, where shoots were thinner. This is in line with earlier findings according to which intensive reindeer grazing limits the foraging of willow grouse by reducing the height of willow and length of shoots (Ims et al., 2007; den Herder et al., 2008). In the areas of medium intensity of grazing there was an indication that willow grouse may reduce the growth of willow as found in an earlier study (Hakkarainen et al., 2007).

IMPLICATIONS FOR PREDICTING THE EFFECTS OF CLIMATE CHANGE

The results support the aerial photo surveys (Sturm et al., 2001; Tape et al., 2006) and experimental evidence (Jones et al., 1997; Walker et al., 2006) according to which the growth of willow is likely to increase as a consequence of climatic warming. Future temperature scenarios predict that the average projected increase in the mean annual temperature in the Central Arctic will be more than 5 °C in the next decades (ACIA, 2005; IPCC, 2007). Even though the warming will be more pronounced during the winter than during the summer, this could theoretically increase the range limits of willow hundreds of kilometers northward. However, the results also show that the growth and height of willow is likely to be affected more by a variety of factors than by the regional temperature regime alone. For example, the distance from the sea,

local topography, reindeer grazing, and depth of the annual thaw were shown to be important for the height and growth of willow shrubs.

The height and growth of willow were strongly associated with the depth of thaw. The results imply that if the observed permafrost warming (Osterkamp and Romanovsky, 1999; Serreze et al., 2000; Pavlov and Moskalenko, 2002) leads to degradation of permafrost and deepening of thaw depth as predicted (ACIA, 2005; Anisimov and Reneva, 2006), then the effects of thaw depth may exceed the direct effects of climate warming on the growth of willow. However, the relationship between summer warmth and the depth of thaw is not linear (Smith and Riseborough, 1996). If climatic warming increases the biomass of vegetation or the depth of snow cover, the insulative effects may counteract the effects of temperature on thaw depth (Walker et al., 2003). In addition, variations in small-scale topography are likely to affect the thawing patterns of soils.

It is noteworthy that an increase in the nutrient status of soils—a predicted effect of climate change (ACIA 2005) that was not studied here—is also likely to increase the abundance of deciduous shrubs (Bret-Harte et al., 2008). For instance, the observed exceptionally high and dense willow thickets in the central Yamal Peninsula (Avramchik, 1969) have been proposed to have developed on old landslides after leaching of salts from clayey marine sediments (Ukrainitseva et al., 2003). Also, it was shown that an artificial increase in the nutrient status of soils increased the growth of dwarf birch (*Betula nana*), albeit not the growth of willow (*Salix pulchra*) (Bret-Harte et al., 2002).

According to observations from high-latitude areas, precipitation has generally increased, while the duration of snow cover has generally declined in recent decades due to increases in spring temperatures (Serreze et al., 2000; ACIA, 2005; IPCC, 2007). If the winter precipitation increases in the next decades as expected (ACIA, 2005; IPCC, 2007), this may however result in delayed snowmelt, which is likely to decrease the growth of woody plants (Vaganov et al., 1999; Kirdeyanov et al., 2003; Schmidt et al., 2006).

Taking into consideration the importance of willow-dominated shrub vegetation to carbon balance (Blanken and Rouse, 1994), wildlife (Chernov, 1985), and to the peoples of the Arctic (Prokof'yeva, 1964), more efforts should be directed to understanding the mechanisms that determine the growth patterns of them.

Acknowledgments

I am grateful to Risto Virtanen for valuable help and advice during all the stages of this work. Risto Virtanen, Bruce Forbes, Anu Eskelinen and two anonymous referees gave constructive comments on earlier versions of the manuscript. Elina and Juha Kaarlejärvi were most valuable assistants in the fieldwork as was Karoliina Räsänen in the laboratory measurements. Jari Oksanen is thanked for statistical advice. The logistical support in 2006 was provided by a group of scientists from the Institute of Plant and Animal Ecology, Ural Division of the Russian Academy of Science in Jekaterinburg. The discussions with the leader of this group (Margarita A. Megadova), now sadly deceased, and Mikhail G. Golovatin helped me to deepen my understanding of the ecology of willows. Lassi Kalleinen helped me with drawing the figures and Aaron Bergdahl revised the language. I was funded by the Academy of Finland financed ENSINOR project (decision number 208147) and Arktis graduate school (University of Lapland) during this work.

References Cited

- ACIA, 2005: *Arctic Climate Impacts Assessment*. Cambridge: Cambridge University Press.
- Andreyashkina, N. I., and Peshkova, N. V., 1995: Vegetation. In: Dobrinskii, L. N. (ed.), *The Nature of Jamal*. Jekaterinburg: Jekaterinburg Nauka, 174–217 (in Russian).
- Anisimov, O., and Reneva, S., 2006: Permafrost and changing climate: the Russian perspective. *Ambio*, 35: 169–175.
- Astakhov, V. I., Kaplyanskaya, F. A., and Tarnogradsky, V. D., 1996: Pleistocene permafrost of West Siberia as a deformable glacier bed. *Permafrost and Periglacial Processes*, 7: 165–191.
- Avramchik, I. A., 1969: Subzonal characteristic of a vegetative cover of tundra, forest-tundra and taiga of West-Siberian Lowland. *Botanicheski Zhurnal*, 3: 410–421 (in Russian).
- Baskin, L. M., 2005: Number of wild and domesticated reindeer in Russia in the late 20th century. *Rangifer*, 25(1): 51–57.
- Blanken, P. D., and Rouse, W. R., 1994: The role of willow-birch forest in the surface energy balance at Arctic treeline. *Arctic and Alpine Research*, 26(4): 403–411.
- Bliss, L. C., and Matveyeva, N. V., 1992: Circumpolar arctic vegetation. In: Chapin, F. S., III, Jefferies, R., Reynolds, J., Shaver, G., and Svoboda, J. (eds.), *Arctic Ecosystems in a Changing Climate: An Ecophysiological Perspective*. Orlando, Florida: Academic Press, 59–89.
- Bret-Harte, M. S., Shaver, G. R., and Chapin, F. S., III, 2002: Primary and secondary stem growth in arctic shrubs: implications for community response to environmental change. *Journal of Ecology*, 90: 251–267.
- Bret-Harte, M. S., Mack, M. C., Goldsmith, G. R., Sloan, D. B., DeMarco, J., Shaver, G. R., Ray, P. M., Biesinger, Z., and Chapin, F. S., 2008: Plant functional types do not predict biomass responses to removal and fertilization in Alaskan tussock tundra. *Journal of Ecology*, 96: 713–726.
- Chapin, F. S., Shaver, G. R., Giblin, A. E., Naderhoffer, K. J., and Laundre, J. A., 1995: Responses of arctic tundra to experimental and observed changes in climate. *Ecology*, 76: 694–711.
- Chernov, Y. I., 1985: *The Living Tundra*. Cambridge: Cambridge University Press.
- Crawley, M. J., 2002: *Statistical Computing: an Introduction to Data Analysis Using S-Plus*. Chichester: John Wiley & Sons.
- den Herder, M., and Niemelä, P., 2003: Effects of reindeer on the re-establishment of *Betula pubescens* subsp. *czerepanowii* and *Salix phylicifolia* in a subarctic meadow. *Rangifer*, 23(1): 3–11.
- den Herder, M., Virtanen, R., and Roininen, H., 2004: Effects of reindeer browsing on tundra willow and its associated insect herbivores. *Journal of Applied Ecology*, 41: 870–879.
- den Herder, M., Virtanen, R., and Roininen, H., 2008: Reindeer herbivory reduces willow growth and grouse forage in a forest-tundra ecotone. *Basic and Applied Ecology*, 9: 324–331.
- Edlund, S. A., and Egginton, P. A., 1984: Morphology and description of an outlier population of tree-sized willows on western Victoria Island, District of Franklin. In: *Current Research, Part A, Geological Survey of Canada, Paper 84-1A*, 279–285.
- Epstein, H., Beringer, J., Gould, W. A., Lloyd, A. H., Thompson, C. D., Chapin, F. S., III, Michaelson, G. J., Ping, C. L., Rupp, T. S., and Walker, D. A., 2004: The nature of spatial transitions in the Arctic. *Journal of Biogeography*, 31: 1917–1933.
- Eskelinen, A., 2008: Herbivore and neighbour effects on tundra plants depend on species identity, nutrient availability and local environmental conditions. *Journal of Ecology*, 96: 155–165.
- Gataullin, V. N., 1988: *Upper Quaternary Deposits of the Western Coast of the Yamal Peninsula*. PhD thesis. Leningrad: VSEGEI (in Russian).
- Gough, L., Ramsey, E. A., and Johnson, D. R., 2007: Plant-herbivore interactions in Alaskan arctic tundra change with soil nutrient availability. *Oikos*, 116: 407–418.
- Hakkarainen, H., Virtanen, R., Honkanen, J. O., and Roininen, H., 2007: Willow bud and shoot foraging by grouse in relation to snow level in NW Finnish Lapland. *Polar Biology*, 3(5): 619–624.
- Hobbie, S. E., and Chapin, F. S., III, 1998: An experimental test of limits to tree establishment in Arctic tundra. *Journal of Ecology*, 86: 449–461.
- Ims, R. A., Yoccoz, N. G., Bråthen, K. A., Fauchald, P., Tveraa, T., and Hausner, V., 2007: Can reindeer overabundance cause a trophic cascade? *Ecosystems*, 10: 607–622.
- IPCC, 2007: *Climate Change 2007: Impacts, Adaptation and Vulnerability. Contribution of Working Group II to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change*. Cambridge: Cambridge University Press.
- Jarvis, P., and Linder, S., 2000: Constraints to growth of boreal forest. *Nature*, 405: 904–905.
- Jones, M. H., Bay, C., and Nordenhäll, U., 1997: Effects of experimental warming on arctic willows (*Salix* spp.): a comparison of responses from the Canadian High Arctic, Alaskan Arctic, and Swedish subarctic. *Global Change Biology*, 3(Suppl. 1): 55–60.
- Kirdyanov, A., Hughes, M., Vaganov, E., Schweingruber, F., and Silkin, P., 2003: The importance of early summer temperature and date of snow melt for tree growth in the Siberian subarctic. *Trees*, 17: 61–69.
- Kitti, H., Forbes, B. C., and Oksanen, J., 2009: Long- and short-term effects of reindeer grazing on tundra wetland vegetation. *Polar Biology*, 32: 253–261.
- Knorre, A. A., Kirdyanov, A. V., and Vaganov, E. A., 2006: Climatically induced variability in aboveground production in forest-tundra and northern taiga of central Siberia. *Oecologia*, 147: 86–95.
- Liston, G. E., McFadden, J. P., Sturm, M., and Pielke, R. A., Sr, 2002: Modelled changes in arctic tundra snow, energy and moisture fluxes due to increased shrubs. *Global Change Biology*, 8: 17–32.
- Lloyd, A. H., Yoshikawa, K., Fastie, C. L., Hinzman, L., and Fraver, M., 2003: Effects of permafrost degradation on woody plant vegetation at Arctic treeline on the Seward Peninsula, Alaska. *Permafrost and Periglacial Processes*, 14: 93–101.
- Manseau, M., Huot, J., and Crête, M., 1996: Effects of summer grazing by caribou on composition and productivity of vegetation: community and landscape level. *Journal of Ecology*, 84: 503–513.
- Matveyeva, N. V., 1998: *Zonation in Plant Cover of the Arctic*. Proceedings of the Komarov Botanical Institute, No. 21, St. Petersburg: Russian Academy of Sciences (in Russian).
- Nadelhoffer, K. J., Giblin, A. E., Shaver, G. R., and Laundre, J. A., 1991: Effects of temperature and substrate quality on element mineralization in six arctic soils. *Ecology*, 72(1): 242–253.
- Osterkamp, T. E., and Romanovsky, V. E., 1999: Evidence for warming and thawing of discontinuous permafrost in Alaska. *Permafrost and Periglacial Processes*, 10: 17–37.
- Ouellet, J.-P., and Boutin, S., 1994: Responses to simulated grazing and browsing of vegetation available to caribou in the Arctic. *Canadian Journal of Zoology*, 72: 1426–1435.
- Pajunen, A., Virtanen, R., and Roininen, H., 2008: The effect of reindeer grazing on the composition and species richness of vegetation in forest-tundra ecotone. *Polar Biology*, 31: 1233–1244.
- Pavlov, A. V., and Moskalenko, N. G., 2002: The thermal regime of soils in the north of Western Siberia. *Permafrost and Periglacial Processes*, 13: 43–51.
- Pinheiro, J. C., and Bates, D. M., 2000: *Mixed-Effects Models in S and S-PLUS*. New York: Springer.
- Post, E., and Pedersen, C., 2008: Opposing plant community responses to warming with and without herbivores. *PNAS*, 105(34): 12353–12358.

- Prokof'yeva, E. D., 1964: The Nentsy. In: Levin, M. G., and Potapov, L. P. (eds.), *The Peoples of Siberia*. Chicago, London: University of Chicago Press, 547–570.
- R Development Core Team, 2005: *R: a Language and Environment for Statistical Computing*. <http://www.R-project.org>.
- Schickhoff, U., Walker, M. D., and Walker, D. A., 2002: Riparian willow communities on the Arctic Slope of Alaska and their environmental relationships: a classification and ordination analysis. *Phytocoenologia*, 32(2): 145–204.
- Schmidt, N. M., Baittinger, C., and Forchhammer, M. C., 2006: Reconstructing century-long snow regimes using estimates of High Arctic *Salix arctica* radial growth. *Arctic, Antarctic, and Alpine Research*, 38(2): 257–262.
- Schuur, E. A. G., Grummer, K. G., Vogel, J. G., and Mack, M. C., 2007: Plant species composition and productivity following permafrost thaw and thermokarst in Alaskan tundra. *Ecosystems*, 10: 280–292.
- Seppälä, M., 2004: *Wind as a Geomorphic Agent in Cold Climates*. Cambridge: Cambridge University Press, 56–79 and 236–248.
- Serreze, M. C., Walsh, J. E., Chapin, F. S., III, Osterkamp, T., Dyurgerov, M., Romanovsky, V., Oechel, W. C., Morison, J., Zhang, T., and Barry, R. G., 2000: Observational evidence of recent change in the northern high-latitude environment. *Climatic Change*, 46: 159–207.
- Shiyatov, S. G., and Mazepa, V. S., 1995: Climate. In: Dobrinski, L. N. (ed.), *The Nature of Yamal*. Yekaterinburg: Yekaterinburg Nauka, 174–216 (in Russian).
- Smith, M. W., and Riseborough, D. W., 1996: Permafrost monitoring and detection of climate change. *Permafrost and Periglacial Processes*, 7: 301–309.
- Stammler, F., 2005: *Reindeer Nomads Meet the Market—Culture, Property and Globalisation at the “End of the Land.”* New Brunswick, New Jersey/London, U.K.: Transaction Publishers, 45–120.
- Sturm, M., Racine, C. R., and Tape, K., 2001: Increasing shrub abundance in the Arctic. *Nature*, 411: 546–547.
- Sturm, M., Schimel, J., Michaelson, G., Welker, J. M., Oberbauer, S. F., Liston, G. E., Fahnestock, J., and Romanovsky, E., 2005: Winter biological processes could help convert arctic tundra to shrubland. *BioScience*, 55(1): 17–26.
- Svendsen, J. I., Astakhov, V. I., Bolshiyarov, D. Yu., Demidov, I., Dowdeswell, J. A., Gataullin, V., Hjort, C., Hubberten, H. W., Larsen, E., Mangerud, J., Melles, M., Möller, P., Saarnisto, M., and Siegert, M. J., 1999: Maximum extent of the Eurasian ice sheets in the Barents and Kara Sea region during the Weichselian. *Boreas*, 28: 234–242.
- Tape, K., Sturm, M., and Racine, C., 2006: The evidence for shrub expansion in northern Alaska and the Pan-Arctic. *Global Change Biology*, 12: 686–702.
- Ukrainseva, N. G., Streletskaya, K. A., Ermokhina, K. A., and Yermakov, S. Y., 2003: Chemical properties of plant-soil permafrost system on landslide slopes, Yamal, Russia. In: Phillips, M., Springman, S. M., Arenson, L. U., and Balkema, A. A. (eds.), *Proceedings of the 8th International Conference on Permafrost, Zurich, Switzerland, 21–25 July 2003*, Vol. 1. Lisse: Swets & Zeitlinger, 1149–1154.
- Vaganov, E. A., Hughes, M. K., Kirilyanov, A. V., Schweingruber, F. H., and Silkin, P. P., 1999: Influence of snowfall and melt timing on tree growth in subarctic Eurasia. *Nature*, 400: 149–151.
- Vakhtina, T. V., 1964: The yield of leaves of some fodder shrubs in the tundra and their utilization for deer breeding. *Problemy Severa*, 8: 290–296.
- Walker, D. A., 1987: Height and growth rings of *Salix lanata* ssp. *richardsonii* along the coastal temperature gradient of northern Alaska. *Canadian Journal of Botany*, 65: 988–993.
- Walker, D. A., Jia, G. J., Epstein, H. E., Raynolds, M. K., Chapin, F. S., III, Copass, C., Hinzman, L. D., Knudson, J. A., Maier, H. A., Michaelson, G. J., Nelson, F., Ping, C. L., Romanovsky, V. E., and Shiklomanov, N., 2003: Vegetation-soil-thaw-depth relationships along a Low-Arctic bioclimate gradient, Alaska: synthesis of information from the ATLAS studies. *Permafrost and Periglacial Processes*, 14: 103–123.
- Walker, D. A., Raynolds, M. K., Daniëls, F. J. A., Einarsson, E., Elvebakk, A., Gould, W. A., Katenin, A. E., Kholod, S. S., Markon, C. J., Melnikov, E. S., Moskalenko, N. G., Talbot, S. S., and Yurtsev, B. A., 2005: The circumpolar Arctic vegetation map. *Journal of Vegetation Science*, 16: 267–282.
- Walker, M. D., Wahren, C. H., Hollister, R. D., Henry, G. H. R., Ahlquist, L. E., Alatalo, J. M., Bret-Harte, M. S., Calef, M. P., Callaghan, T. V., Carroll, A. B., Epstein, H. E., Jónsdóttir, I. S., Klein, J. A., Magnússon, B., Molau, U., Oberbauer, S. F., Rewa, S. P., Robinson, C. H., Shaver, G. R., Suding, K. N., Thompson, C. C., Tolvanen, A., Totland, O., Turner, P. L., Tweedie, C. E., Webber, P. J., and Wookey, P. A., 2006: Plant community responses to experimental warming across the tundra biome. *PNAS*, 103(5): 1342–1346.

MS accepted April 2009