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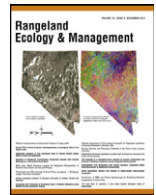
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Original Research

Productivity and CO₂ Exchange of Great Plains Ecoregions. I. Shortgrass Steppe: Flux Tower Estimates[☆]

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ABSTRACT

The shortgrass steppe (SGS) occupies the southwestern part of the Great Plains. Half of the land is cultivated, but significant areas remain under natural vegetation. Despite previous studies of the SGS carbon cycle, not all aspects have been completely addressed, including gross productivity, ecosystem respiration, and ecophysiological parameters. Our analysis of 1998–2007 flux tower measurements at five Bowen ratio–energy balance (BREB) and three eddy covariance (EC) sites characterized seasonal and interannual variability of gross photosynthesis and ecosystem respiration. Identification of the nonrectangular hyperbolic equation for the diurnal CO₂ exchange, with vapor pressure deficit (VPD) limitation and exponential temperature response, quantified quantum yield α , photosynthetic capacity A_{max} , and respiration rate r_d with variation ranges ($19 < \alpha < 51 \text{ mmol mol}^{-1}$, $0.48 < A_{max} < 2.1 \text{ mg CO}_2 \text{ m}^{-2} \text{ s}^{-1}$, $0.15 < r_d < 0.49 \text{ mg CO}_2 \text{ m}^{-2} \text{ s}^{-1}$). Gross photosynthesis varied from 1 100 to 2 700 g CO₂ m^{−2} yr^{−1}, respiration from 900 to 3,000 g CO₂ m^{−2} yr^{−1}, and net ecosystem production from −900 to +700 g CO₂ m^{−2} yr^{−1}, indicating that SGS may switch from a sink to a source depending on weather. Comparison of the 2004–2006 measurements at two BREB and two parallel EC flux towers located at comparable SGS sites showed moderately higher photosynthesis, lower respiration, and higher net production at the BREB than EC sites. However, the difference was not related only to methodologies, as the normalized difference vegetation index at the BREB sites was higher than at the EC sites. Overall magnitudes and seasonal patterns at the BREB and the EC sites during the 3-yr period were similar, with trajectories within the ± 1.5 standard deviation around the mean of the four sites and mostly reflecting the effects of meteorology.

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Introduction

The shortgrass steppe (SGS) ecoregion occupies the southwestern part of the Great Plains of North America, covering approximately 0.34 10⁶ km² (Fig. 1). Half of the land is cultivated, but significant areas

remain under natural vegetation dominated by the Grama-buffalo grass (*Bouteloua-Buchloe*) association (Lauenroth, 2008). As an important resource for agricultural production (cereals and animal products) and ecosystem services (Burke et al., 2008), the SGS ecoregion has been a focus of comprehensive systems—ecological studies, including the “Grassland Biome” project of the US IBP Program (Van Dyne, 1971). As a result, many aspects of the structure and functioning of the shortgrass steppe ecosystems have been thoroughly described (Lauenroth and Burke, 2008). In particular, biological productivity and element cycling of the shortgrass steppe have received special attention and eventually led to construction of dynamic ecosystem simulation models such as ELM and Century (Innis, 1978 and Parton et al., 1987, respectively). Several studies of CO₂ exchange in ecosystems of the shortgrass ecoregion of North America and similar ecoregions of Europe and Asia

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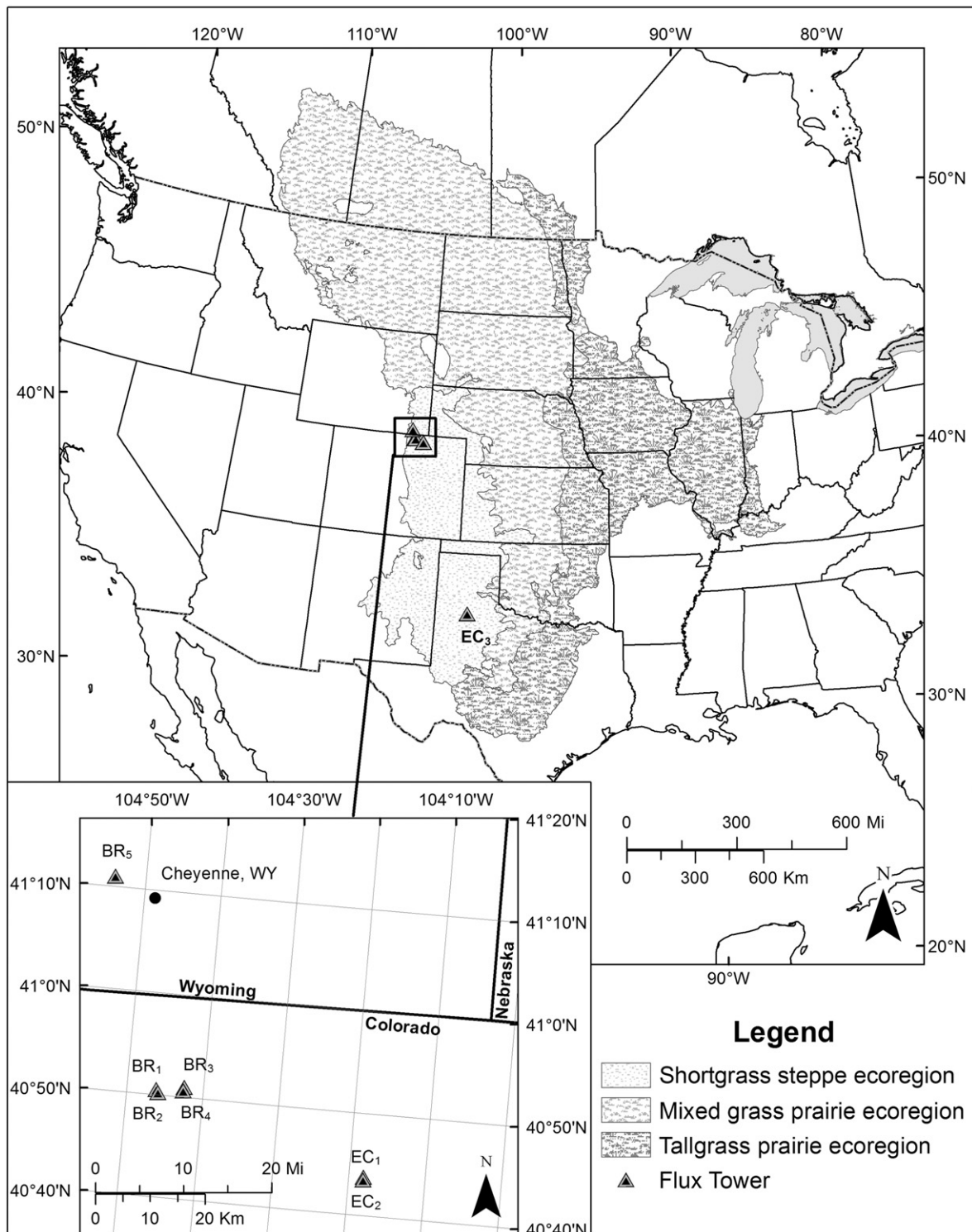


Figure 1. Major grassland ecoregions of the Great Plains (Omernik, 1987; Homer et al., 2015) and location of the study sites.

were conducted using the chamber, flux tower, and remote sensing techniques (Brown and Trlica, 1977a; LeCain et al., 2000; Li et al., 2000; Gilmanov et al., 2004b, 2005, 2007, 2010; Fu et al., 2006, 2009; Beletti-Marchesini et al., 2007; Wu et al., 2008; Alfieri et al., 2009; Rey et al., 2012; Zhang et al., 2012; Rajan et al., 2013; Shao et al., 2013; Gao et al., 2014). Nevertheless, due to anticipated changes in climate and anthropogenic management, certain aspects of the SGS carbon cycle require additional scrutiny. Particularly, we do not have sufficient data on the ecosystem-scale estimates of fundamental characteristics of gross primary productivity (GPP), total ecosystem respiration (RE), and

the resulting net ecosystem carbon budget, as the few available GPP and RE estimates of the North American SGS (Andrews et al., 1974; Brown and Trlica, 1977b; Detling, 1979; Risser et al., 1981; Klopatek and Risser, 1982) were based on extrapolating and modeling data from physiological studies at the leaf, plant, or chamber scales. Long-term measurements of ecosystem-scale CO_2 exchange of SGS communities using the Bowen ratio–energy balance (BREB) and later the eddy covariance (EC) techniques began in the mid-1990s (Svejcar et al., 1997; Alfieri et al., 2009). There are different opinions concerning the evaluation and comparison of BREB and EC flux tower measurements of CO_2 exchange

of nonforest, particularly grassland ecosystems. Dugas et al. (1997) recognized BREB as an adequate tool for CO₂-exchange measurements on grasslands, and the method was used by the US Department of Agriculture—Agricultural Research Service (USDA-ARS) Rangeland CO₂ Flux project (Angell et al., 2001; Frank and Dugas, 2001; Sims and Bradford, 2001; Emmerich, 2003; Gilmanov et al., 2003b, 2006; Mielenick et al., 2005; Svejcar et al., 2008), as well as in other studies of grasslands and croplands (Dugas et al., 1991, 1999; Ham and Knapp, 1998; Asseng and Hsiao, 2000; Ansley et al., 2002; Gilmanov et al., 2003a; Baron et al., 2005; Scott et al., 2006; Irmak, 2010; Jamiyansharav et al., 2011; O'Dell et al., 2014). Phillips and Beeri (2008) have summarized long-term BREB measurements in the mixed grassland of North Dakota and established consistent relationships with remote sensing indices. Comparison of parallel BREB and EC system measurements demonstrated that although there are certain differences in energy and water fluxes, CO₂ fluxes recorded by the two systems did not differ significantly (Dugas et al., 2001; Wolf et al., 2008). Hipps et al. (2002) observed reasonable agreement for water vapor fluxes measured by parallel running BREB and EC systems in a crested wheatgrass (*Agropyron desertorum*) field, wherein the EC CO₂ fluxes were always larger than those measured with BREB. Nevertheless, results from both systems indicated that during the period of study the crested wheatgrass ecosystem was a net source of carbon. In contrast, Alfieri et al. (2009) found that BREB overestimates the magnitudes of carbon dioxide fluxes. Skinner and Wagner-Riddle summarized the problem: "Currently missing are studies comparing EC and BR flux estimates for the entire season or for complete annual cycles to determine how differences between systems affect long-term estimates of net C exchange" (Skinner and Wagner-Riddle, 2012, p. 377). Clearly, there is a need to compare seasonal patterns, annual totals, and ecophysiological parameters obtained from the two methods to evaluate opportunities to integrate the legacy BREB data accumulated from grassland and crops with the growing datasets from the EC networks.

The objectives of this study, using all the available BREB and EC datasets of flux tower net CO₂ exchange (F) measurements in ecosystems of the North American SGS ecoregion, are to 1) partition net CO₂ fluxes into gross photosynthesis (P_g) and total ecosystem respiration (R_e) components and estimate major ecophysiological parameters; 2) gap fill the data, describe seasonal patterns of CO₂ exchange components and parameters, and estimate weekly and annual totals of gross primary production, total ecosystem respiration, and net ecosystem production (NEP); 3) compare CO₂ exchange components and parameter estimates from BREB and EC flux towers, and 4) compare source/sink activity of ecosystems of the SGS ecoregion to mixed-grass and tallgrass ecoregions.

Materials and Methods

Study Sites

Five BREB and three EC tower sites considered in this paper (see Fig. 1) represent fundamental properties of grassland ecosystems of the SGS ecoregion and reflect features of the dominant management regimes: ungrazed, moderately grazed, and heavily grazed.

Measurements at the BREB towers were conducted during 1998–2006 at several locations in northeastern Colorado and southeastern Wyoming. During 1998–1999, BREB tower measurements (location BR₁, Table 1) were conducted at an ungrazed site on native shortgrass steppe at the Central Plains Experimental Range (CPER), administered by the USDA-ARS. In 2000 the BREB system was moved to a second ungrazed site at the CPER (location BR₂). Both of these ungrazed sites were on an Olney fine-loamy soil (mesic Ustic Haplargids). The BREB station was moved a third time in 2004 to a heavy continuously grazed plot (BR₃) located on a Remmit coarse-loamy soil (mesic Ustic Haplocambids) at the CPER, and a second BREB tower was installed in a nearby moderate continuously grazed plot (location BR₄) on a Zigweid fine-loamy soil (mesic Ustic Haplocambids). Characterized by a rather flat relief, the BREB sites have a spatially homogeneous vegetation of up to 0.3 m in height dominated by the C₄ grasses blue grama (*Bouteloua gracilis* [H. B. K.] Lag. Ex Steud.) and buffalo grass (*Buchloe dactyloides* [Nutt.] Engelm.), accompanied by a mixture of C₃ grasses (western wheat [*Pascopyrum smithii* {Rydb.} A.], needle-and-thread [*Stipa comata* Trin and Rupr.], and others); cacti; and shrubs. The grasses constitute > 70% of the total vegetation (Milchunas et al., 1989). According to both on-site measurement (LeCain et al., 2002) and remote sensing data (see corresponding section later), leaf area index at the site seldom exceeds 1.5 m² m⁻².

The BREB site in southeastern Wyoming (BR₅), located in the northwestern part of the SGS ecoregion, represents a true mixed-grass prairie dominated by the midheight cool-season grasses western wheatgrass and needle-and-thread grass. The site also contains a warm-season grass, blue grama, which is characteristic of shortgrass steppe. The soil of the site is an Ascalon sandy loam classified as a mixed, mesic, Aridic Argiustoll (LeCain et al., 2000). CO₂ exchange measurements at BR₅ were conducted during 1997–1998.

Ecological similarity of all sites is emphasized by the fact that in terms of the most detailed Level 4 Ecoregion taxonomy of the United States (Omernik and Griffith, 2008), sites BR₁–BR₅, EC₁, and EC₂ were classified as "High Plains. 25c. Moderate Relief Plains," and site EC₃ as "High Plains. 25i. Llano Estacado."

The three sites using the EC methodology were within the native shortgrass steppe region but had been converted to seeded pastures or, at one time, row cropping. Measurements were conducted on ungrazed (location EC₁) and moderately grazed (location EC₂) USDA Conservation Reserve Program (CRP) pastures at the Curtis Ranch near Briggsdale, Colorado, from 2004–2007, and on a pasture seeded to warm-season C₄ perennial bunch grass Caucasian bluestem (*Bothriochloa bladhii* [Retz] S. T. Blake) in the Texas High Plains (location EC₃) from 2010–2011 (Rajan et al., 2013). Before 1987, the EC₁ (and EC₂) site had been in a wheat/fallow production rotation. In 1987, the site was placed in the CRP. It had no livestock grazing and a well-established cover of the cool season C₃ grasses western wheat and needle-and-thread and the native warm season C₄ grasses buffalo grass, sideoats grama (*Bouteloua curtipendula* [Michx.] Torr.), and little

Table 1
Location of the flux towers in the SGS ecoregion of the North American Great Plains

Code	Site	Latitude	Longitude	Elevation	Tower type	Yr	Management	Principal investigator
BR ₁	CPER ungrazed	40.8424	−104.7697	1648	BREB	1998–1999	Ungrazed	J. Morgan
BR ₂	CPER ungrazed	40.8373	−104.7654	1672	BREB	2000–2003	ungrazed	J. Morgan
BR ₃	CPER heavy continuous	40.8501	−104.7086	1655	BREB	2004–2006	14 AU per 160 ac during 5 mo.	J. Morgan
BR ₄	CPER moderate continuous	40.8444	−104.7107	1650	BREB	2004–2006	7 AU per 160 ac during 5 mo.	J. Morgan
BR ₅	Cheyenne ungrazed	41.1835	−104.9017	1910	BREB	1997–1998	Ungrazed	J. Morgan
EC ₁	Curtis Ranch ungrazed	40.7297	−104.3013	1520	EC	2004–2007	CRP ungrazed	N. Hanan
EC ₂	Curtis Ranch moderate	40.7251	−104.3014	1504	EC	2004–2007	CRP, spring and fall grazing while forage ≥ 250 kg/ha	N. Hanan
EC ₃	Lockney pasture	34.1384	−101.4799	1100	EC	2010–2011	Grazed (2010) Ungrazed (2011)	N. Rajan

bluestem (*Schizachyrium scoparium* [Michx.] Nash). The soil of the site was classified as an Ascalon fine sandy loam (mixed, mesic, Aridic, Arjiustoll). In October 2003, a grazing treatment (EC₂) was opened to spring and fall grazing at moderate intensity. Moderate grazing intensity was cattle grazing during two main periods in the spring and fall, until such time as approximately 250 kg/ha of forage remained in the

pasture. Site EC₃, located in the High Plains of Texas, was seeded to Caucasian bluestem in May 2007 and was grazed three times (May, July, August) in 2010, which was a high-productive year, but not grazed in 2011 due to extreme drought. The soil is a Pullman clay loam (fine, mixed, superactive, thermic Torrertic Paleustoll) with a flat relief (0%–1% slope) (Rajan et al., 2013).

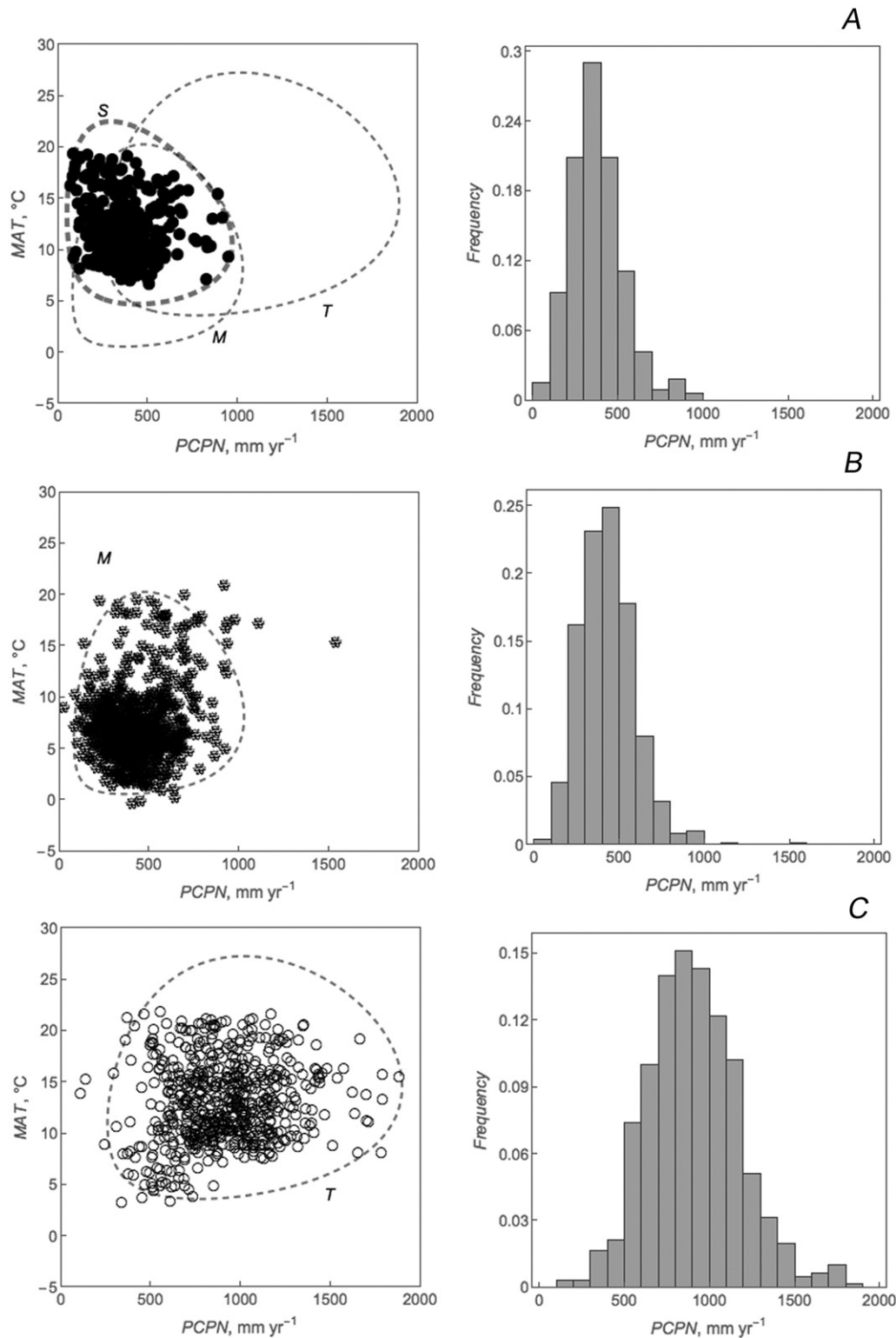


Figure 2. Mean annual temperature versus annual precipitation scatterplots (left) and precipitation histograms (right) for airport meteorological stations in (A) shortgrass steppe (95% confidence contour S), (B) mixed-prairie (contour M), and (C) tallgrass-prairie (contour T) ecoregions. Data were taken from the meteorological forecasting site Tutiempo Network S. L. Available at: <http://en.tutiempo.net/climate/united-states.html>.

The shortgrass steppe ecoregion has higher temperatures and lower precipitation (Fig. 2) than the mixed and tallgrass ecoregions, and most rains (~70%) occur during May–September.

Meteorological conditions during the years of the study demonstrated a wide variety of weather patterns (Fig. 3). For example, at the ungrazed SGS Colorado sites (1998–2007), conditions varied from the wettest and coolest in 1999 (hydrologic year precipitation $PCPN_h = 545$ mm, sum of temperatures above 5°C $T_{sum5} = 2123$ -degree days) to the hottest and driest days in 2002 ($PCPN_h = 160$ mm, $T_{sum5} = 2344$ -degree days). At the Wyoming site, lower temperatures ($MAT = 6.7^\circ\text{C}$) and higher precipitation ($PCPN_h = 437$ mm) in 1997 resulted in higher net production ($NEP = 436$ g CO_2 m^{-2} yr^{-1}) than in 1998 ($MAT = 7.8^\circ\text{C}$, $PCPN_h = 247$ mm, $NEP = 142$ g CO_2 m^{-2} yr^{-1}). Weather conditions during the first (2010) year at the Lockney site were also marked by lower temperatures ($MAT = 14.9^\circ\text{C}$) and high precipitation ($PCPN_h = 680$ mm), followed by hot and dry conditions during the second (2011) year ($MAT = 15.4^\circ\text{C}$, $PCPN_h = 187$ mm). Such a wide range of meteorological factors provides a good opportunity to model climatic response of carbon dioxide exchange.

Instrumentation and Data Processing

Flux tower sites analyzed in this study were equipped with modern field instrumentation corresponding to the BREB or EC method described in Dugas (1993) and Campbell Scientific (1998) for BREB and Dugas et al. (2001), Meyers (2001), Alfieri et al. (2009), and Rajan et al. (2013) for EC. Standard correction procedures and outlier detection algorithms recommended for grassland ecosystems were applied to the raw datasets (Dugas et al., 1997, 2001; Falge et al., 2001; Wolf et al., 2008). In particular, during periods when the BREB algorithm was not valid for calculating turbulent diffusivity, it was estimated

using atmospheric parameters as described by Dugas et al. (1999). Resulting “Level 2” (using Ameriflux terminology) files containing aggregated subhourly (20 min for BREB and 30 min for EC) values of the net CO_2 fluxes (F) and the ancillary variables (incoming photosynthetically active radiation [Q], air temperature [T_a], soil temperature [5-cm depth, T_s], air relative humidity [RH], VPD , and others) served as inputs for the procedure of partitioning F into gross photosynthesis (P_g) and ecosystem respiration (R_e) components using the “light–soil temperature– VPD ” response method (Gilmanov et al., 2013, 2014; Morgan et al., 2016). Following are the basic equations of the method:

$$P_g(Q, VPD) = \varphi(VPD, VPD_{cr}, \sigma_{VPD}) \left(\alpha Q + A_{max} - \sqrt{(\alpha Q + A_{max})^2 - 4\alpha A_{max} \theta Q} \right) / (2\theta) \quad [1]$$

$$\varphi(VPD, VPD_{cr}, \sigma_{VPD}) = \begin{cases} \exp\left[-(VPD - VPD_{cr})^2 / \sigma_{VPD}^2\right], & VPD > VPD_{cr} \\ 1, & VPD \leq VPD_{cr} \end{cases} \quad [2]$$

$$R_e(T_s) = \begin{cases} r_0 \exp(k_T T_s), & \text{exponential temperature response} \\ r_d, & \text{no temperature response} \end{cases} \quad [3]$$

$$F(Q, T_s, VPD) = P_g(Q, VPD) - R_e(T_s) \quad [4]$$

where α is the initial slope (apparent quantum yield), A_{max} is the plateau (photosynthetic capacity) of the light-response, θ is the convexity parameter (Thornley and Johnson, 2000), r_d is respiration rate when no temperature response was observed, r_0 and k_T are the coefficients of the exponential temperature response ($r_0 = R_e[0]$), and the normalized VPD -response function $\varphi(VPD, VPD_{cr}, \sigma_{VPD})$ depends on two parameters: the critical value VPD_{cr} , below which water deficit doesn't affect photosynthesis ($\varphi = 1$ for $VPD \leq VPD_{cr}$), and the VPD -response curvature

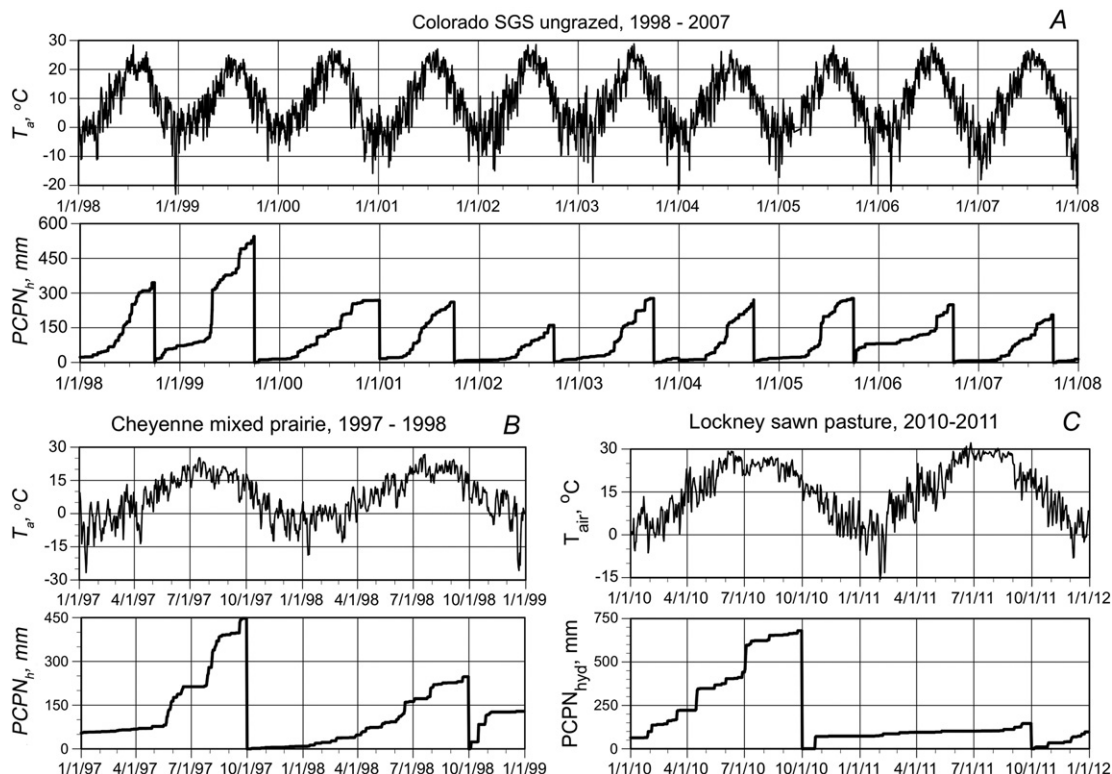


Figure 3. Seasonal and interannual dynamics of mean daily air temperature (T_a) and hydrologic year precipitation ($PCPN_h$) at the ungrazed shortgrass steppe site (A), Cheyenne site (B), and Lockney site (C).

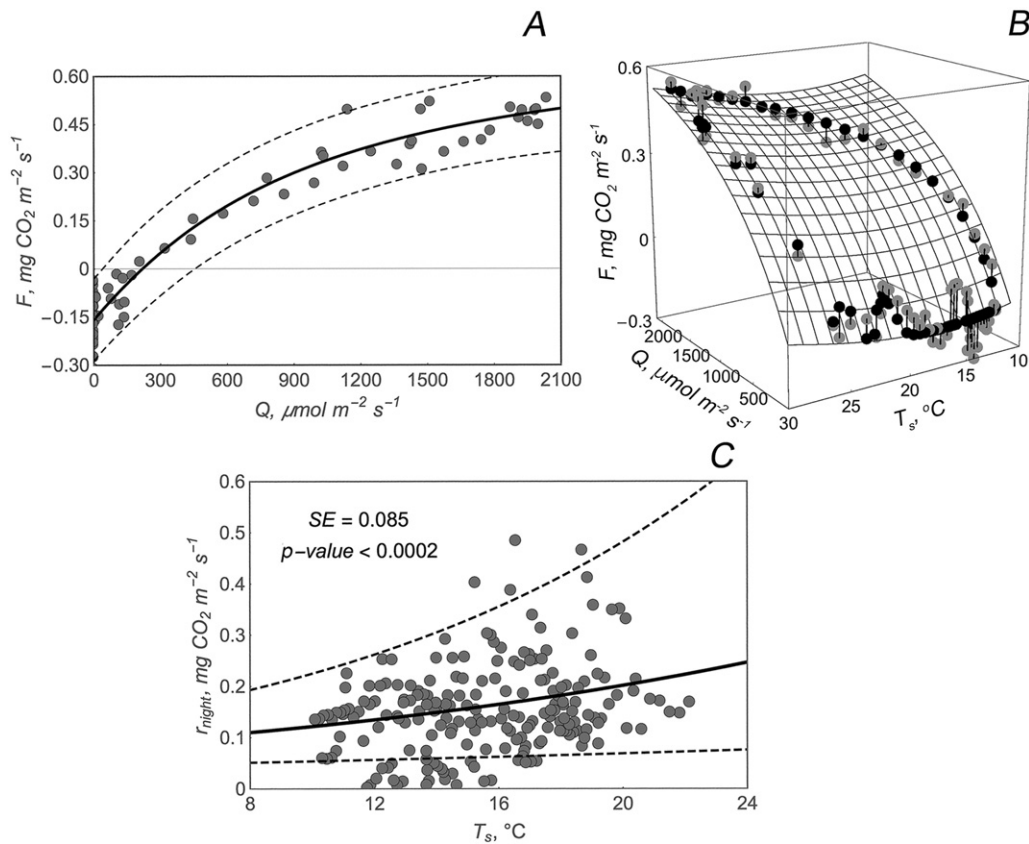


Figure 4. CO₂ uptake of the ungrazed shortgrass steppe ecosystem of the BR₂ site for 2 June, 2003 (DOY 153): (A) light-response with the 95% confidence band (Eq. [6], $\alpha = 0.00082 \text{ mg CO}_2 \mu\text{mol}^{-1}$; $A_{\max} = 0.898 \text{ mg CO}_2 \text{ m}^{-2} \text{ s}^{-1}$; $\theta = 0.409$; $r_d = 0.158 \text{ mg CO}_2 \text{ m}^{-2} \text{ s}^{-1}$); (B) light-soil temperature-VPD response (Eqs. [1–5], $\alpha = 0.00077 \text{ mg CO}_2 \mu\text{mol}^{-1}$; $A_{\max} = 1.07 \text{ mg CO}_2 \text{ m}^{-2} \text{ s}^{-1}$; $\theta = 0.05$; $r_0 = 0.134 \text{ mg CO}_2 \text{ m}^{-2} \text{ s}^{-1}$; $k_T = 0.009 \text{ }^\circ\text{C}^{-1}$; $\text{VPD}_{\text{cr}} = 1.0 \text{ kPa}$; $\sigma_{\text{VPD}} = 4.3 \text{ kPa}$; gray dots—measurement data; black dots—model data; surface—response with mean daily VPD); (C) exponential regression of the nighttime respiration on soil temperature for DOY 149–157 with the 95% confidence band, $r_{\text{night}} = 0.074 \exp(0.05 T_s)$.

parameter, σ_{VPD} ($1 \leq \sigma_{\text{VPD}} \leq 30$), with lower values describing a strong water-stress effect and higher values describing a weak effect (Gilmanov et al., 2013). The average daytime respiration rate r_d was calculated as:

$$r_d = \frac{r_0}{(t_2 - t_1)} \int_{t_1}^{t_2} \exp(k_T T_s(t)) dt \quad [5]$$

where t_1 and t_2 denote moments of sunrise and sunset, correspondingly. In addition, for days with identifiable Eqs. [1–4] parameters, exponential Eq. [3] was also fitted for the n -day window of the nighttime subhourly data centered at the day under consideration ($n \approx 9$).

Parameters of Eqs. [1–5] characterizing diurnal dynamics of the CO₂ exchange, such as apparent quantum yield α , photosynthetic capacity $A_{\max}$, convexity of the light-response θ , and others, were numerically fitted to the datasets of individual measurement days of each measurement site-year. Interpolation and extrapolation of the seasonal patterns demonstrated by these parameters to days with missing measurements were used as major tools of gap filling (in addition to statistical interpolation of missing data for short subhourly intervals). Diurnal rates ($\text{mg CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) of P_g , R_e , and F were calculated by Eqs. [1–5] using the diurnal data for meteorological drivers (Q , T_a , T_s , RH , VPD). The 24-hr integration of these rates provided the year-round daily ($\text{g CO}_2 \text{ m}^{-2} \text{ d}^{-1}$) series of $P_g(t)$, $R_e(t)$, and $F(t)$ values ($t = 1, 2, \dots, 365$) for corresponding years of study. Daily estimates of the ecosystem-scale ecophysiological parameters of photosynthesis and respiration in Eq. [1–5] were also obtained. For days when identification of Eq. [1–2] parameters was not possible (mostly outside the growing season), the net CO₂ exchange for the day j was described as $F(T_s) = -R_e(T_s)$ with parameters of

Eq. [3] estimated from the subhourly (F , T_s) data for the n -day window centered in day j (depending on data availability, n varied from 9 to 14).

Light-Use Efficiency

Among the diversity of light-use efficiency (LUE) coefficients (cf. Bonhomme, 2000), the two most frequently used are the physiological LUE coefficient calculated as a ratio of gross primary production P_g to absorbed photosynthetically active radiation Q_a , $\epsilon_{\text{phys}} = P_g/Q_a$, and the ecological LUE coefficient, $\epsilon_{\text{ecol}} = P_g/Q$, where Q is incident photosynthetically active radiation ($Q_a < Q$). In contrast to ϵ_{phys} characterizing mostly physiological abilities of plants to assimilate atmospheric CO₂, ϵ_{ecol} also incorporates ecologically significant information about structure and architecture of the plant canopy, as $\epsilon_{\text{ecol}} = \epsilon_{\text{phys}} \cdot f_{\text{PARa}}(\text{LAI})$, where $f_{\text{PARa}}(\text{LAI}) = Q_a/Q$ is the fraction of PAR absorbed by the plant canopy. While physiological LUE coefficient ϵ_{phys} is considered a valuable characteristic providing a basis for the rapidly growing wave of publications on “production efficiency models” (PEMs) pioneered by Monteith (1972) and Sellers (1987), it was demonstrated that ϵ_{phys} poorly (Garbulsky et al., 2010) or even negatively (Runyon et al., 1994; Polley et al., 2011) correlates with LAI, evapotranspiration, and ecosystem productivity. In contrast, since the early period of plant production studies, ϵ_{ecol} is known as a positive correlate of ecosystem productivity (Odum, 1959; Ničiporovič, 1968; Duvigneaud, 1974; Runyon et al., 1994). A particularly close relationship between ecological light-use efficiency and productivity was demonstrated for grasslands: Data by Sims and Singh (1978) for $n = 36$ site-yr of production measurements in grasslands of the western United States show highly significant correlation ($r = 0.91$, $P < 10^{-6}$) between net primary production and ecological LUE. On these ecosystem comparisons, we are using the

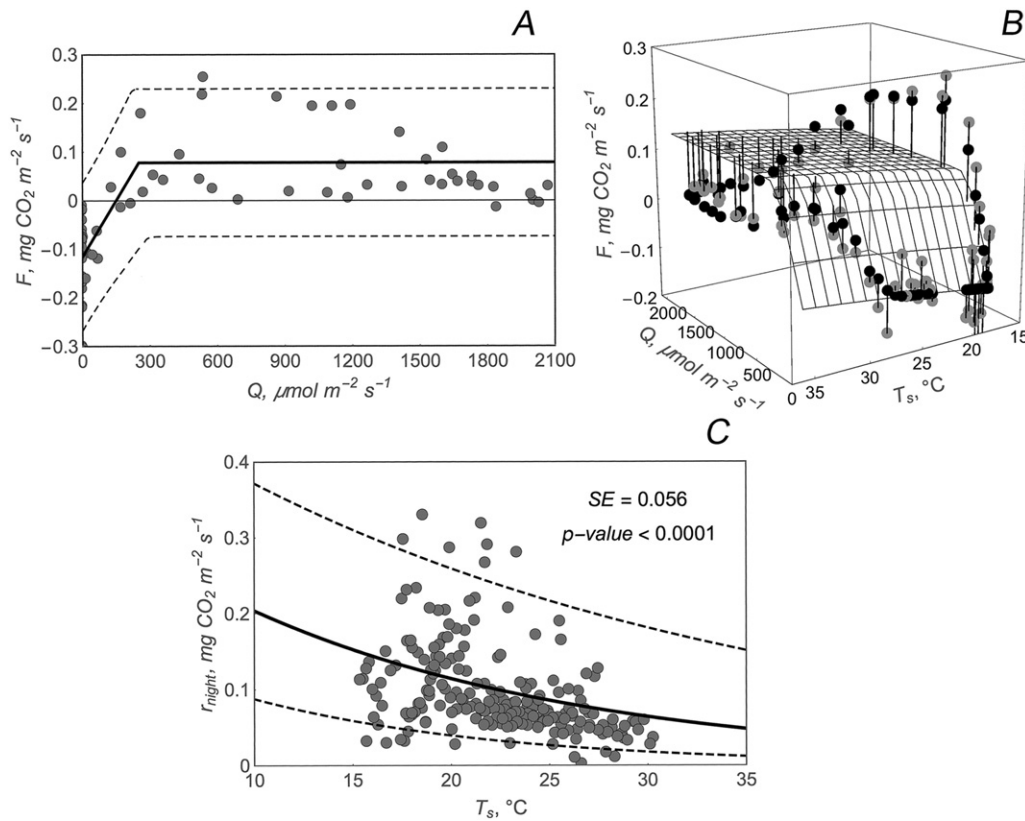


Figure 5. CO₂ uptake of the ungrazed shortgrass steppe ecosystem of the BR₂ site for 1 July, 2003 (DOY 182): (A) light-response with the 95% confidence band (Eq. [6], $\alpha = 0.00077 \text{ mg CO}_2 \mu\text{mol}^{-1}$; $A_{\max} = 0.193 \text{ mg CO}_2 \text{ m}^{-2} \text{ s}^{-1}$; $\theta = 1.0$; $r_d = 0.115 \text{ mg CO}_2 \text{ m}^{-2} \text{ s}^{-1}$); (B) light-soil temperature-VPD response (Eqs. [1–5], $\alpha = 0.00082 \text{ mg CO}_2 \mu\text{mol}^{-1}$; $A_{\max} = 0.332 \text{ mg CO}_2 \text{ m}^{-2} \text{ s}^{-1}$; $\theta = 1.0$; $r_0 = 0.254 \text{ mg CO}_2 \text{ m}^{-2} \text{ s}^{-1}$; $k_T = -0.041^\circ\text{C}^{-1}$, $\text{VPD}_{\text{cr}} = 1.0 \text{ kPa}$; $\text{VPD}_{\text{VPD}} = 3.16 \text{ kPa}$; gray dots—measurement data; black dots—model data; surface—response with mean daily VPD); (C) exponential regression of the nighttime respiration on soil temperature for DOY 178–186 with the 95% confidence band, $r_{\text{night}} = 0.364 \exp(-0.058 T_s)$.

ecological LUE coefficient for brevity denoted below as $\varepsilon = P_g/Q$ based on the incident PAR, Q .

Remote Sensing Data

Remote sensing indices were recognized as useful tools for interpretation and scaling-up of on-site ecosystem-scale measurements (Wylie et al., 2004, 2007; Gilmanov et al., 2005; Heinsch et al., 2006). In this paper we used the 250-m data of the normalized difference vegetation index (NDVI) derived from the Moderate Resolution Imaging Spectroradiometer (MODIS) sensor for all sites and the 1-km resolution MODIS LAI estimates (DAAC/ORN, 2015) as supplemental tools for comparing productivity of the BREB and EC sites in Colorado. More specifically, for the 2004–2006 period, we used the 7-day NDVI composites from the expedited MODIS (eMODIS) product (Jenkinson et al., 2010). The 8-d MODIS LAI values were recalibrated to match the 7-d eMODIS NDVI time step.

Results and Discussion

Light-Response Functions

Variability of the meteorological drivers affecting the SGS ecosystems during the measurement period (see Fig. 3) is translated into variability of their CO₂ exchange. Two major environmental response patterns of the CO₂ exchange may be distinguished as illustrated in Figures 4 and 5. The case of photosynthetic response dominated by incoming radiation Q is illustrated in Figure 4 by the BR₂ site data for 2 June, 2003, which was a clear, sunny day (daily PAR total $51.9 \text{ mol m}^{-2} \text{ d}^{-1}$, $Q_{\max} = 2038 \mu\text{mol m}^{-2} \text{ s}^{-1}$) with moderate temperatures (mean $T_a = 14.2^\circ\text{C}$, $T_{a,\max} = 21.6^\circ\text{C}$) and low evaporative

demand ($\text{VPD}_{\text{avg}} = 0.95 \text{ kPa}$, $\text{VPD}_{\max} = 1.68 \text{ kPa}$). As shown in Figure 4A, on this day, CO₂ uptake of the SGS community is fairly well described by the univariate nonrectangular light-response function:

$$F(Q) = \frac{1}{2\theta} \left(\alpha Q + A_{\max} - \sqrt{(\alpha Q + A_{\max})^2 - 4\alpha A_{\max} \theta Q} \right) - r_d, \quad [6]$$

where r_d is average daily ecosystem respiration rate and other parameters as described above in Eqs. [1–5]. This equation describes a significant part of the variance of the CO₂ exchange ($R^2 = 0.95$) achieving the mean squared error of $SE = 0.063 \text{ mg CO}_2 \text{ m}^{-2} \text{ s}^{-1}$. There is little difference between the morning and afternoon branches of the light-response curve for day of the year (DOY) 153. Predominance of radiation as a major driver of CO₂ uptake in this case is confirmed by the fact that switching from a univariate light-dependent function (Eq. [6], Fig. 4A) to the multivariate light-soil temperature-VPD response function (Eqs. [1–4], Fig. 4B) results in only a small additional improvement of the data fit ($SE = 0.060 \text{ mg CO}_2 \text{ m}^{-2} \text{ s}^{-1}$, $R^2 = 0.96$). Visually, the low significance of the VPD as a factor controlling the CO₂ uptake in this case is reflected by the fact that the black dots in Fig. 4B, displaying model-generated $F(Q, T_s, \text{VPD})$ values, remain close to the response surface $F(Q, T_s; \text{VPD}_{\text{avg}})$ (shown by a mesh) corresponding to the model [1]–[4] applied with a constant value $\text{VPD}_{\text{avg}} = 0.95 \text{ kPa}$ equal to the average daily VPD for DOY 153. This surface, in its turn, is close to the gray dots indicating original data.

In contrast to day 153, day 182 (1 July, 2003) was a sunny, hot summer day (daily PAR total $53.68 \text{ mol m}^{-2} \text{ d}^{-1}$, $Q_{\max} = 2071.9 \mu\text{mol m}^{-2} \text{ s}^{-1}$) with high temperatures (mean $T_a = 20.1^\circ\text{C}$, $T_{a,\max} = 35.6^\circ\text{C}$) and strong evaporative demand ($\text{VPD}_{\text{avg}} = 2.5 \text{ kPa}$, $\text{VPD}_{\max} = 5.2 \text{ kPa}$). The diurnal $F(Q)$ plot for this day exhibited a strong hysteresis-like pattern with the afternoon branch significantly lower than the morning branch (see Fig. 5A).

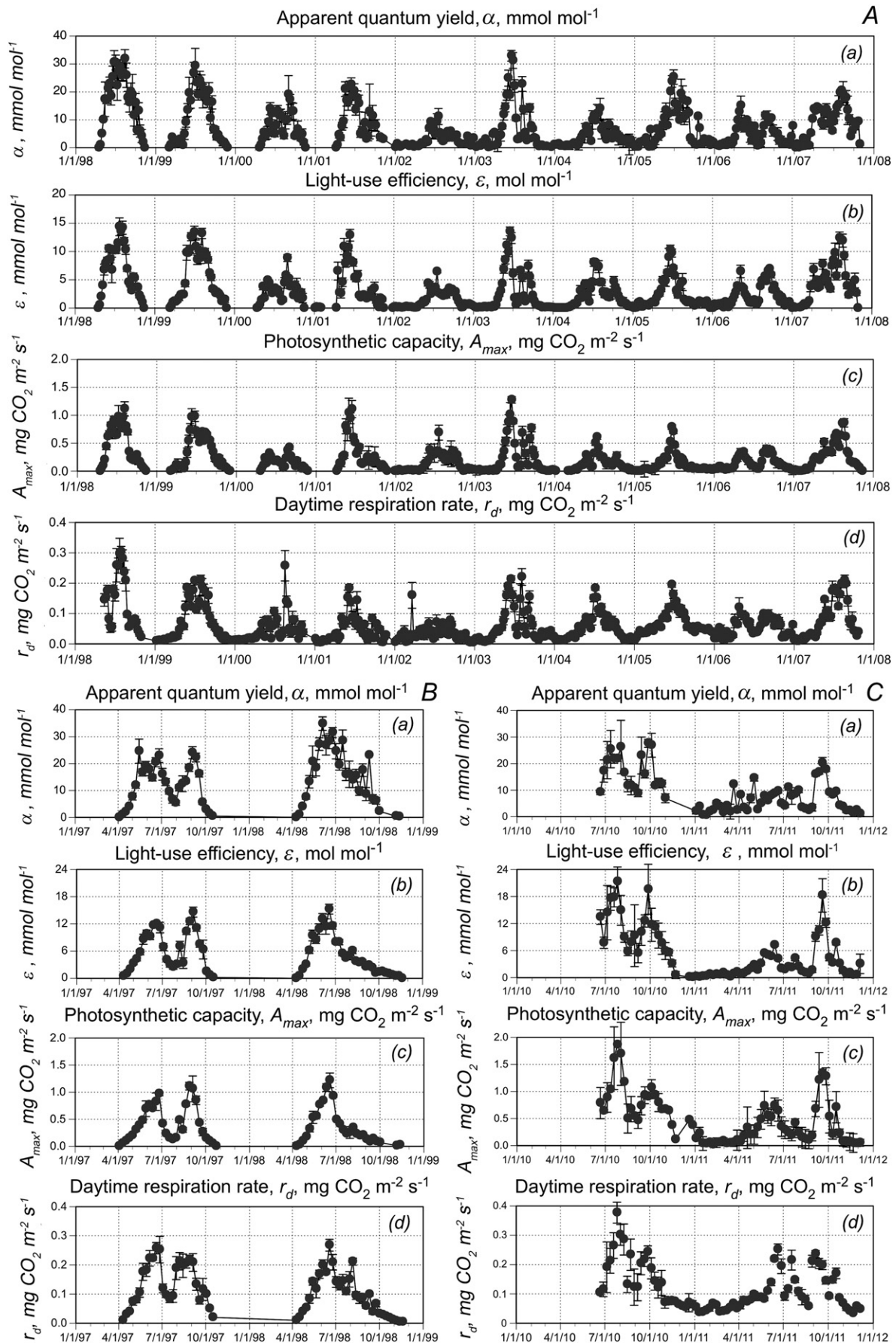


Figure 6. Seasonal and interannual variation of the ecosystem-scale parameters at ungrazed shortgrass steppe (A), Cheyenne mixed prairie (B), and Lockney pasture (C): (a)—apparent quantum yield, α ; (b)—light-use efficiency, ϵ ; (c)—photosynthetic capacity, A_{max} ; (d)—daytime respiration rate, r_d . Dots indicate weekly means; bars, errors of the means.

Table 2

Ecosystem-scale ecophysiological parameters of the SGS ecoregion ecosystems resulting from flux tower data processing by the light-soil temperature—vapor pressure deficit response method, Eqs. [1]–[6]

Site	Yr		PCPN _h	A _{max}	A _{max,wk}	α _{max}	α _{max,wk}	ε _{max}	ε _{max,wk}	r _{d,max}	r _{d,max,wk}
		Grazing		mg CO ₂ m ⁻² d ⁻¹				mmol mol ⁻¹		mg CO ₂ m ⁻² d ⁻¹	
BR ₁	1998	No	345	1.32	1.11	38.8	31.9	21.2	14.6	0.49	0.32
BR ₁	1999	No	545	1.15	1.03	41.7	28.9	19.3	13.8	0.31	0.21
BR ₁	2000	No	254	0.57	0.42	30.4	19.0	13.9	8.2	0.42	0.27
BR ₁	2001	No	261	1.63	1.11	36.4	22.7	16.7	12.9	0.25	0.18
BR ₂	2002	No	161	0.89	0.69	22.4	11.2	6.8	5.7	0.28	0.16
BR ₂	2003	No	277	1.39	1.27	45.1	32.9	16.4	13.6	0.30	0.22
BR ₃	2004	Yes	271	0.99	0.69	18.7	15.2	13.2	8.9	0.25	0.17
BR ₃	2005	Yes	254	1.25	0.99	26.0	21.9	18.4	11.3	0.38	0.16
BR ₃	2006	Yes	214	0.52	0.36	22.9	14.2	10.0	6.8	0.21	0.15
BR ₄	2004	Yes	212	0.92	0.74	22.1	17.8	15.8	11.8	0.36	0.17
BR ₄	2005	Yes	270	1.29	1.04	35.3	24.9	20.5	12.3	0.32	0.15
BR ₄	2006	Yes	178	0.87	0.60	27.3	17.1	17.7	11.1	0.31	0.19
BR ₅	1997	No	447	1.28	1.11	41.5	24.6	20.8	14.6	0.45	0.26
BR ₅	1998	No	247	1.50	1.22	42.0	34.9	17.9	15.2	0.35	0.27
EC ₁	2004	No	271	0.66	0.44	25.4	14.0	16.7	8.7	0.20	0.15
EC ₁	2005	No	277	1.00	0.79	33.7	25.3	15.5	10.2	0.22	0.20
EC ₁	2006	No	249	0.50	0.39	19.7	16.5	8.8	6.0	0.15	0.12
EC ₁	2007	No	206	0.94	0.86	24.1	20.5	19.3	12.8	0.29	0.21
EC ₂	2004	Yes	264	0.68	0.61	27.8	19.9	12.3	8.1	0.20	0.18
EC ₂	2005	Yes	262	0.90	0.70	36.8	20.5	11.2	8.9	0.25	0.18
EC ₂	2006	Yes	234	0.48	0.39	21.8	13.1	10.99	6.9	0.24	0.12
EC ₃	2010	Yes	680	2.10	1.86	51.0	27.7	28.7	21.3	0.43	0.38
EC ₃	2011	No	187	1.99	1.34	27.9	20.2	19.6	15.2	0.39	0.25
Mean (SD)			286 (125)	1.08 (0.41)	0.85 (0.36)	31.1 (9.2)	21.3 (6.3)	16.2 (4.9)	11.3 (3.7)	0.31 (0.09)	0.20 (0.06)
Min–Max			146–680	0.48–2.10	0.36–1.86	18.7–51.0	11.2–32.9	6.8–28.7	5.7–21.3	0.15–0.49	0.12–0.38

A_{max} indicates maximum daily photosynthetic capacity; A_{max,wk}, maximum mean weekly photosynthetic capacity; α_{max}, maximum daily apparent quantum yield; α_{max,wk}, maximum mean weekly apparent quantum yield; ε_{max}, maximum daily light-use efficiency; ε_{max,wk}, maximum mean weekly light-use efficiency; r_{d,max}, maximum daytime respiration rate; r_{d,max,wk}, maximum mean weekly daytime respiration rate.

Temperature response of the nighttime ecosystem respiration r_{night} for the 9-d window centered on DOY 182 showed a decrease of respiration with negative exponential temperature response coefficient $k_T = -0.058$ (°C)⁻¹ (see Fig. 5C). This suggests that the decrease of the CO₂ uptake in the afternoon likely cannot be explained by an increase of respiration with temperature, making vapor pressure deficit the most significant factor controlling F under drought stress conditions.

This conclusion is strongly supported by results of the nonlinear regression, which showed a highly significant ($P < 0.0001$) magnitude $\sigma_{VPD} = 3.16$ kPa of the curvature parameter in Eq. [2], at the same time indicating a negative value exponential respiration coefficient $k_T = -0.041$ (°C)⁻¹ of low significance (P value 0.076). VPD-response function [2] with parameters $VPD_{cr} = 1.0$ kPa and $\sigma_{VPD} = 3.16$ kPa indicates a nearly fivefold decrease of CO₂ uptake rate as the VPD on 1 July, 2003, increased from near 0 in the morning to 5.2 kPa in the afternoon.

Illustrating the VPD control of CO₂ exchange on DOY 182, 2003, the black dots in Fig. 5B (full model [1]–[4] predictions) markedly deviate from the VPD-independent surface $F_{VPDavg}(Q, T_s) = F(Q, T_s, VPD_{avg})$ generated from Eq. [2], with the VPD fixed at its mean value for day 182, $VPD_{avg} = 2.5$ kPa, and are close to the gray dots (measurement data).

While plots similar to Figs. 4–5 were available for most of the days with valid $\{(Q, T_s, VPD, F)\}$ datasets at a subhourly time step, a better way to demonstrate applicability of the [1]–[6] modeling scheme is to examine and evaluate the seasonal and interannual dynamics of the parameters resulting from identification of these equations (Fig. 6).

Seasonal and Interannual Dynamics of Ecophysiological Parameters

Time plots of the weekly aggregated ecophysiological parameters of the SGS ecosystems (Fig. 6) revealed both seasonal patterns of their change within the annual cycle and significant year-to-year variability.

Table 3

Maximum gross ecological light-use efficiency ϵ_{max} (mmol CO₂ mol incident quanta⁻¹) of various types of semiarid grasslands

Ecosystem	Photosynthetic pathway	Annual precipitation, mm	Annual temperature, °C	ε _{max} mmol mol ⁻¹	Reference
Kendall semidesert grassland, AZ	C ₄	356	17.0	5.4	Nouvellon et al., 2000
Sevilleta semidesert grassland, NM	C ₄	242	13.3	6.5	Turner et al., 2006; Litvak, 2013
Kherlebayan-Ulaan, semiarid pasture, Mongolia	C ₃ /C ₄	196	1.2	7.0	Li et al., 2008
Xilinhot semiarid grassland, Inner Mongolia, China	C ₃	290	2.0	8.4	Wang and Zhou, 2012
Jasper Ridge serpentine grassland, California	C ₃	616	14.0	13.0	Valentini et al., 1995
Shira small-tussock steppe, Khakasia, Russia	C ₃	301	0.4	14.7	Belelli-Marchesini et al., 2007; Gilmanov et al., 2010; McCallum et al., 2013
Great Plains shortgrass steppes	C₄	286 (146–680)	9.3 (6.7–15.4)	16.2 (6.8–28.7)	This study
Jornada semidesert grassland, NM	C ₄	272	14.4	17.2	Mielnick et al., 2005; Gilmanov et al., 2010
Vall d'Alinya semiarid grassland, Spain	C ₃	550	6.5	18.3	Gilmanov et al., 2010; Chang et al., 2013
Raymond short/mixed-grass prairie, Alberta, Canada	C ₃	403	5.92	18.7	Schwalm et al., 2006
Shortand typical steppe, Kazakhstan	C ₃	323	1.6	20.0	Gilmanov et al., 2004
Duolun grassland, Inner Mongolia, China	C ₃	399	3.3	21.95	Zhang et al., 2007
Tojal grassland, Portugal	C ₃ /C ₄	669	15.5	23.0	Aires et al., 2008
Bugacpuszta grassland, Hungary	C ₃	562	10.4	29.5	Nagy et al., 2007; Gilmanov et al., 2010

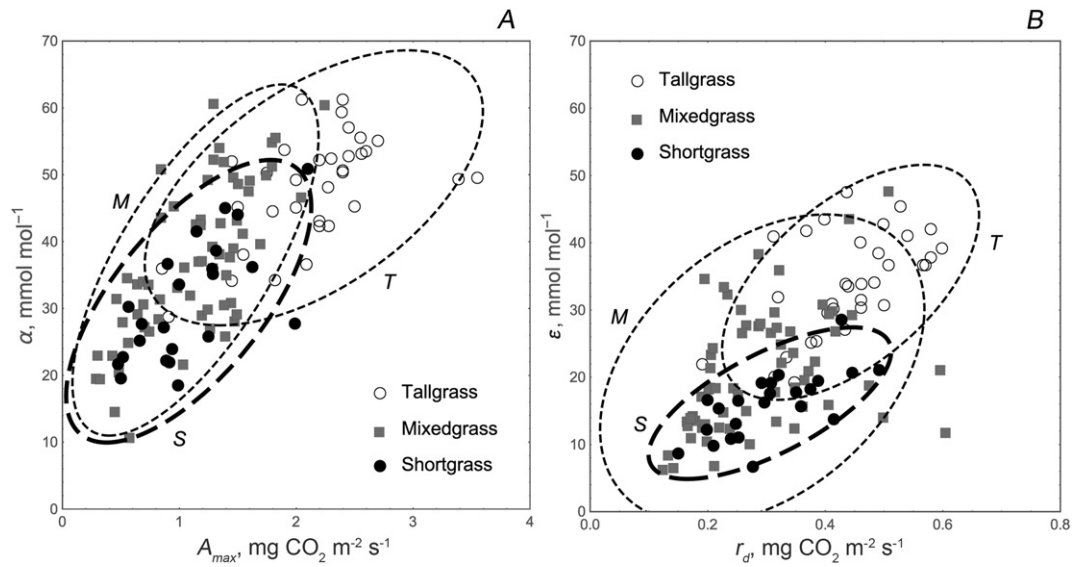


Figure 7. Scatter diagrams of (A) apparent quantum yield versus photosynthetic capacity ($A_{\max}$) and (B) light-use efficiency versus daytime respiration rate (r_d) for ecosystems of the tallgrass, mixed-grass, and shortgrass ecoregions. The ellipses show the 95% statistical confidence zones for corresponding parameter sets. Data for the mixed-grass and tallgrass prairies are from Gilmanov et al. (2017a, b).

Parameters decrease by 50% to 67% in years with unfavorable conditions (yrs 2000, 2002, 2006, 2011, when either extreme drought occurred, PCPN < 180 mm, or low precipitation PCPN < 255 mm was combined MAT > 8.9°C) compared with years with higher precipitation and less

heat stress (1998, 1999, 2001, 2003, 2010). Table 2 has a summary of parameter estimates of the SGS ecoregion resulting from flux tower data processing by the light-soil temperature-VPD response method. Depending on the meteorological and phenological conditions of the

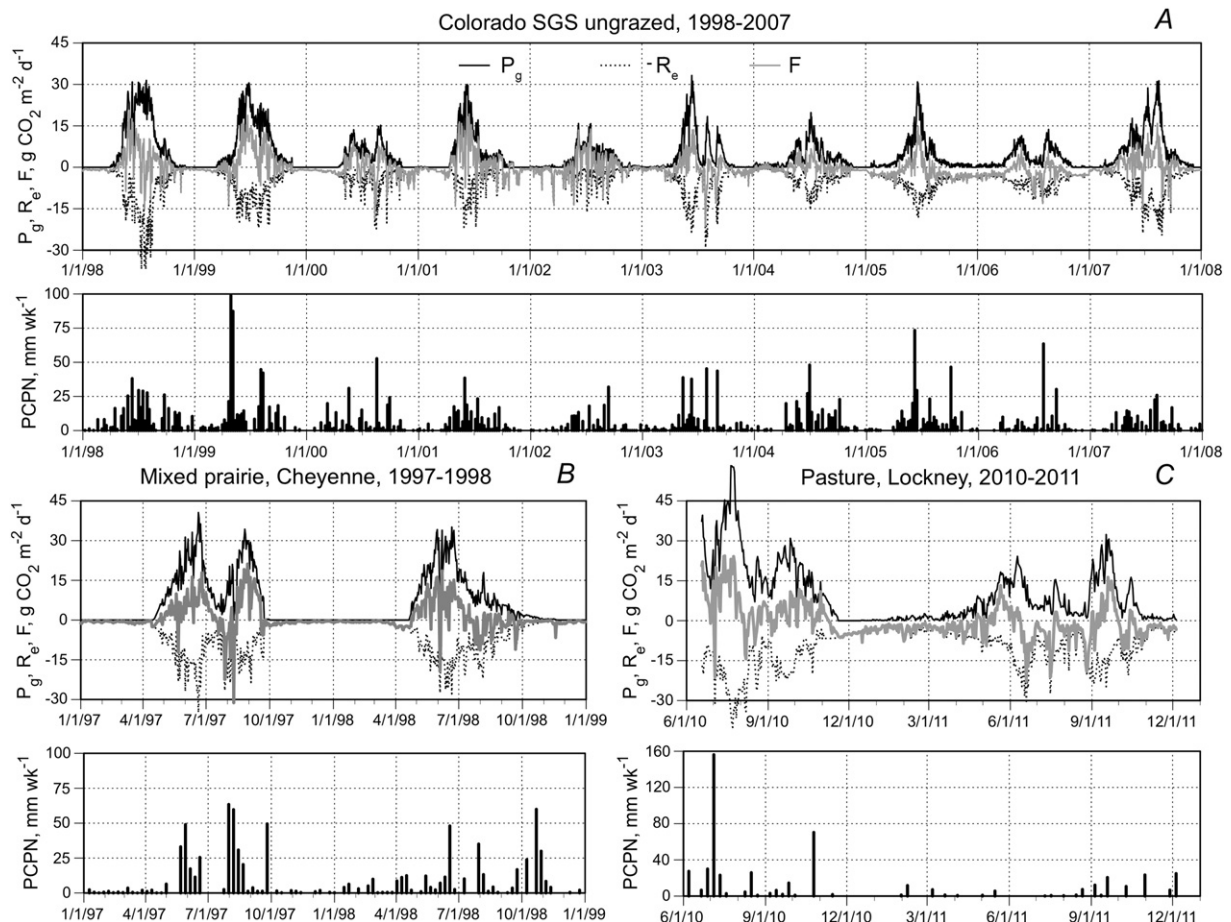


Figure 8. Seasonal and interannual dynamics of gross photosynthesis, ecosystem respiration, net ecosystem production, and precipitation at (A) the Colorado ungrazed shortgrass steppe (1998–2007), (B) mixed-prairie, Cheyenne (1997–1998), and (C) seeded pasture, Lockney (2010–2011).

year, the seasonal curves of parameters may be either unimodal or bimodal (see Fig. 6). In SGS proper sites, maxima of up to 1.63 mg CO₂ m⁻² s⁻¹ for daily photosynthetic capacity A_{max} and 45.1 mmol mol⁻¹ for the apparent daily quantum yield α_{max} were observed. In the bimodal years the midseason values may be as low as 0.2–0.5 mg CO₂ m⁻² s⁻¹ for A_{max} and 5–20 mmol mol⁻¹ for α_{max} . These numbers compare well with the results by LeCain et al. (2003), who used steady-state leaf gas exchange measurements to evaluate native plants enclosed in open-top chambers. For the dominant C₄ species *Bouteloua gracilis*, they estimated maximum A_{max} values of 1.14 for early season, 0.56 for midseason, and 1.05 mg CO₂ m⁻² s⁻¹ for late season. For the subdominant C₃ species *Pascopyrum smithii*, they obtained A_{max} values of 1.44, 0.35, and 1.08 mg CO₂ m⁻² s⁻¹ for early season, midseason, and late season. According to Polley et al. (2010) maximum weekly A_{max} for the SGS site near Nunn, Colorado, is 0.81 mg CO₂ m⁻² s⁻¹, close to our mean maximum of weekly A_{max} of 0.85 mg CO₂ m⁻² s⁻¹ in Table 2. Tower-based A_{max} estimates are also available for two semiarid grasslands in Inner Mongolia, China, with annual precipitation range comparable with shortgrass steppe but lower annual temperature. For the Xilin Gol site (PCPN = 350 mm, MAT = -1°C) maximum A_{max} parameter was estimated at 0.55 mg CO₂ m⁻² s⁻¹ (Fu et al., 2009), whereas for warmer Duolun site (PCPN = 386 mm, MAT = 2.1°C) A_{max} = 1.32 mg CO₂ m⁻² s⁻¹ (Zhang et al., 2012).

Ecosystem-scale quantum yield parameters in Table 2 are also in agreement with literature data for comparable communities. For an SGS site near Nunn, Colorado, Polley et al. (2010) found α_{max} = 55 mmol mol⁻¹. Taking into account a certain overestimation of the slope of the light response by the rectangular hyperbolic model used by these authors (cf. Marshall and Biscoe 1980), this estimate matches well our maximum α_{max} = 51 mmol mol⁻¹ (see Table 2). For dominant SGS species *Bouteloua gracilis*, (C₄) and subdominant *Pascopyrum smithii* (C₃) maximum quantum yield values for the early-season, midseason, and late-season periods were estimated as 47, 46, 66 mmol mol⁻¹ and 66, 51, 53 mmol mol⁻¹, respectively (LeCain et al., 2003). As the community-level value of α is usually lower than that of individual dominant species, these estimates compare well with α_{max} = 51 mmol.

Using eddy covariance measurements at two dry steppe sites in Inner Mongolia with climatic conditions within the shortgrass steppe range (see Fig. 2A), John et al. (2013) estimated maximum initial slopes of the ecosystem-scale light response as 32.7 mmol mol⁻¹ (Doulun site) and 41.6 mmol mol⁻¹ (Xilinhot site), which is practically within the ± 1 standard deviation (SD) interval around the mean of 31.2 mmol mol⁻¹ of daily α_{max} values in Table 2.

The range of the ecological light-use efficiency ϵ in Table 2, calculated as a ratio of total daily photosynthesis to total daily PAR, is approximately half that of α (Figs. 6A–6B), apparently due to the combination of light-saturation and water stress limitations of photosynthesis. Comparative data in Table 3 show that maximum ecological light-use efficiency values, ϵ_{max} , of the Great Plains shortgrass steppe ecosystems fall within the range of estimates for semiarid grasslands of North America, Europe, and Asia, with shortgrass ϵ_{max} being higher than in ecosystems from cooler and drier climates but lower than in communities with temperate climates (see Table 3).

The range of respiration rates, $0 \leq r_d \leq 0.5$ mg CO₂ m⁻² s⁻¹, is approximately one-third the range of A_{max} (Fig. 6C–6D). Pulses of intense CO₂ evolution observed in certain years (2000, 2002, 2003, ungrazed SGS, Fig. 6D) indicate occasional high metabolic activity of the SGS biota, often due to precipitation events (Austin et al., 2004; Parton et al., 2011; Fan et al., 2012).

As expected, ecophysiological parameters for the Cheyenne mixed prairie are comparable (A_{max} , ϵ) or higher (α , r_d) than observed at SGS proper sites (see Table 2). The highest values of parameters were achieved at the seeded Caucasian bluestem (C₄) pasture near Lockney, where in the rain-abundant year 2010, α_{max} = 51 mmol mol⁻¹, ϵ_{max} = 28.7 mmol mol⁻¹, and A_{max} = 2.10 mg CO₂ m⁻² s⁻¹.

To identify the place of SGS ecoregion ecosystems within the general ecophysiological parametric space of the Great Plains grasslands, we compared the (A_{max} , α) scatter diagram and the (r_d , ϵ) scatter diagram of the shortgrass ecoregion with the same plots for the mixed-grass and tallgrass ecosystems (Fig. 7A–7B). The ellipses in Figure 7 delineate the 95% confidence zones of each of the parameter sets. At both two-dimensional plots, the shortgrass ellipse is substantially different

Table 4
Summary characteristics of the annual CO₂ exchange of the shortgrass steppe SGS ecoregion ecosystems during 1998–2007

Site	Yr	Grazing	T_a	PCPN	PCPN _h	$P_{g,max}$	$P_{g,max,wk}$	$R_{e,max}$	$R_{e,max,wk}$	F_{max}	$F_{max,wk}$	GPP	RE	NEP
			°C	mm yr ⁻¹				g CO ₂ m ⁻² d ⁻¹					g CO ₂ m ⁻² yr ⁻¹	
BR ₁	1998	No	8.9	394	345	37.0	27.2	35.17	28.4	20.6	13.0	2683	2334	349
BR ₁	1999	No	9.1	487	545	30.5	29.0	22.12	17.9	18.8	13.1	2592	1917	675
BR ₁	2000	No	9.2	254	254	15.2	13.2	22.25	17.2	9.3	7.0	1113	1293	-181
BR ₁	2001	No	9.2	255	260	30.0	25.0	21.73	15.8	19.2	13.2	1659	1308	350
BR ₂	2002	No	8.9	169	161	17.4	15.1	16.39	8.5	14.2	10.9	968	923	45
BR ₂	2003	No	9.5	278	277	33.2	28.6	29.1	20.4	13.6	9.2	1666	1841	-174
BR ₃	2004	Yes	8.8	271	271	19.8	17.3	16.9	12.5	15.5	6.5	1074	1168	-94
BR ₃	2005	Yes	9.3	285	254	32.1	28.0	26.4	13.7	22.4	20.2	1717	1509	208
BR ₃	2006	Yes	9.2	166	214	14.9	10.3	18.7	12.0	10.0	8.1	1315	1115	200
BR ₄	2004	Yes	8.7	234	212	28.8	23.0	28.1	14.2	16.3	9.3	1134	1252	-118
BR ₄	2005	Yes	9.0	304	270	34.6	30.6	26.6	12.2	22.3	19.3	1754	1307	447
BR ₄	2006	Yes	8.3	122	178	22.3	15.9	24.0	15.0	13.9	8.8	1700	1230	470
BR ₅	1997	No	6.76	401	447	40.6	31.2	34.9	20.7	21.2	14.5	2514	2078	436
BR ₅	1998	No	7.8	367	247	35.1	30.1	27.9	22.0	18.6	11.9	2100	1958	142
EC ₁	2004	No	8.85	282	271	19.8	18.7	15.5	12.3	8.0	9.9	1127	1167	-40
EC ₁	2005	No	8.7	339	277	30.9	25.7	18.0	16.7	14.7	9.0	1439	2009	-570
EC ₁	2006	No	9.1	176	249	13.7	11.0	13.1	10.2	8.4	2.9	1110	1794	-684
EC ₁	2007	No	8.7	206	206	31.3	27.7	24.7	17.9	15.9	10.4	2367	2082	285
EC ₂	2004	Yes	8.7	275	263	22.1	19.5	16.4	14.8	9.4	7.5	1241	1466	-225
EC ₂	2005	Yes	9.0	318	262	26.5	22.1	19.7	16.0	11.1	6.1	1395	1770	-375
EC ₂	2006	Yes	9.1	172	233	12.7	11.2	13.1	8.7	7.5	5.5	1346	1684	-338
EC ₃	2010	Yes	14.9	690	680	58.2	53.1	40.1	34.5	24.3	18.6	4059	3703	256
EC ₃	2011	No	15.4	173	187	32.5	23.7	30.3	23.2	15.8	11.7	2061	2990	-929
Mean (standard deviation)			9.3 (1.9)	288 (125)	285 (120)	27.8 (10.5)	23.4 (9.4)	23.5 (7.3)	16.7 (6.2)	15.3 (5.1)	10.7 (4.4)	1745 (726)	1735 (643)	6 (405)
Min–Max			6.7–15.4	122–690	161–680	12.7–58.2	10.3–53.1	13.1–40.1	8.5–34.5	7.5–24.3	2.9–20.2	968–4059	923–3703	-929–675

T_a indicates mean annual temperature; PCPN, annual precipitation; PCPN_h, hydrologic year precipitation (October previous year to September current year); $P_{g,max}$, maximum daily gross photosynthesis; $P_{g,max,wk}$, maximum mean weekly gross photosynthesis; $R_{e,max}$, maximum daily ecosystem respiration; $R_{e,max,wk}$, maximum mean weekly ecosystem respiration; F_{max} , maximum daily net CO₂ exchange; $F_{max,wk}$, maximum mean weekly net CO₂ exchange; GPP, gross primary production; RE, ecosystem respiration; NEP, net ecosystem production.

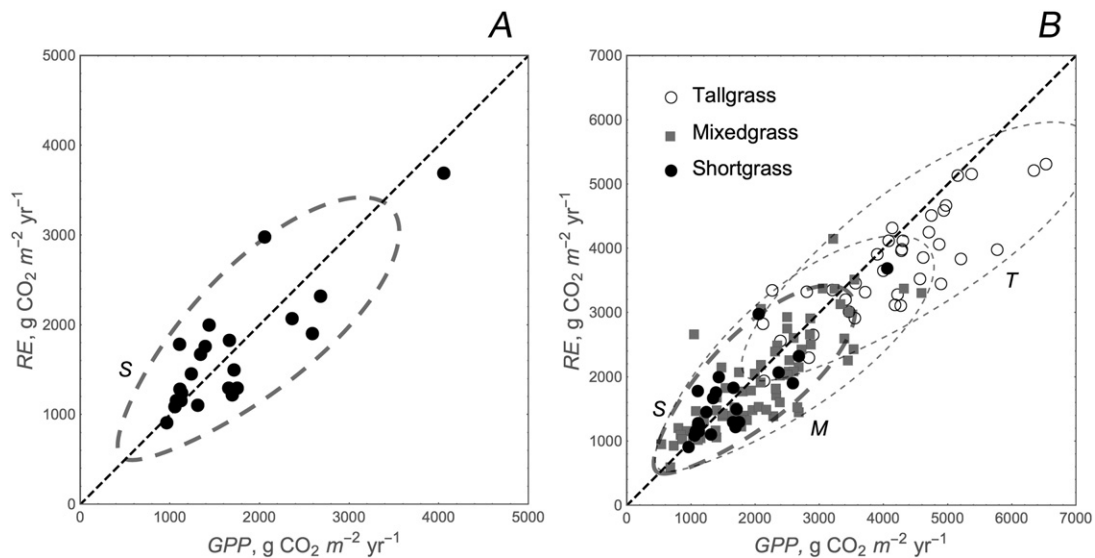


Figure 9. Gross primary productivity-ecosystem respiration (GPP-RE) scatter diagram for shortgrass steppe ecoregion sites (A) compared with the mixed-grass and tallgrass ecoregion sites (B). Contours S, M, and T denote 95% confidence zones for (GPP, RE) pairs for sites in the shortgrass, mixed-grass, and tallgrass ecoregions, respectively. Data for mixed and tallgrass sites are from Gilmanov et al. (2017a, b).

from the tallgrass ellipse, but it overlaps with the mixed-grass ellipse. Considering distribution of the parameters in the entire four-dimensional parametric space $\{(\alpha, \varepsilon, A_{max}, r_d)\}$, it turns out that the centroids $\mu_S = (31.1, 16.2, 1.08, 0.31)$, $\mu_M = (37.8, 21.4, 1.14, 0.29)$, $\mu_T = (48.1, 34.0, 2.14, 0.44)$ of the parametric clusters of the shortgrass, mixed-grass, and tallgrass ecoregions, correspondingly, are significantly different according to the Mahalanobis criterion (Rao, 1965) ($P = 0.0057 < 0.01$ for short – mixed-grass comparison and much lower for comparison of short-tall and mixed-tall clusters).

Parameters α , ε , A_{max} , r_d in Figure 7 using Eqs. [1]–[5] are ecosystem-scale (= canopy-scale as opposed to leaf-scale) parameters, meaning that, e.g., α is calculated per unit of incoming (not absorbed) PAR, and A_{max} measures photosynthetic capacity per m^2 of ground surface (not per m^2 of leaf area, as in physiological photosynthetic capacity $A_{max,L}$). While there are complex interrelations among ecophysiological characteristics of plant canopies aimed at optimization of the photosynthetic uptake (Terashima and Hikosaka, 1995; Wright et al., 2005), leaf area, community diversity, and soil fertility are some of the major factors explaining lower α and A_{max} values in SGS compared with tallgrass prairies, with intermediate values in mixed-prairie communities (see Fig. 7). In particular, as demonstrated by Thornley and France (2007), the canopy A_{max} is proportional (directly proportional in monocultures) to the leaf area index LAI and $A_{max,L}$, resulting in lower photosynthetic capacity of SGS canopies with $LAI \leq 1.5 m^2 m^{-2}$ compared with tallgrass canopies with $LAI \geq 3 m^2 m^{-2}$. Similarly, lower LAI values in the SGS entail lower absorption of solar radiation, explaining lower canopy-level quantum efficiency α . Lower species diversity (including fewer N-fixers) and lower soil fertility in SGS compared with tallgrass prairies leads to lower N content, which also contributes to lower values of α and A_{max} (Evans, 1989; Peterson et al., 1999; Lee et al., 2003).

Seasonal and Interannual Dynamics of Photosynthesis, Respiration, and Net Ecosystem Production

The time series of daily photosynthesis $P_g(t)$, respiration $R_e(t)$, and net CO_2 flux $F(t)$ obtained by application of Eqs. [1]–[5] to the raw tower flux data and their gap filling (Fig. 8A) confirms strong variability of the functioning of the SGS ecosystems emphasized earlier by production ecologists based on biomass and chamber CO_2 exchange measurements and by range scientists based on forage and animal production studies. For example, LeCain et al. (2002) observed a twofold difference in maximum daily rate of CO_2 uptake F in favorable and unfavorable

years in the SGS at CPER in Colorado, with a sixfold difference between the midsummer minimum and the annual maximum during the year with bimodal pattern of F . Field estimates of the aboveground net primary production (ANPP) for the SGS at the CPER for the 1939–2014 period showed a broad range from $14 g m^{-2} yr^{-1}$ in 2002 (year of extreme drought) to $175 g m^{-2} yr^{-1}$ in 2009 ($PCPN_{yr} = 437 mm yr^{-1}$), with a mean of $93 g m^{-2} yr^{-1}$ and a coefficient of variation of 31% (data compiled from Lauenroth and Sala, 1992; SGS-LTER, 2013; Lauenroth, 2013; and Parton 2015, personal communication). The grazing season gain of yearling steers in the same ecosystem during 19 yr of study (1995–2003) was found to vary 2.4 times from 72 to 172 kg/head/season (Hart and Derner, 2008). Data for the Cheyenne (Fig. 8B) and Lockney (Fig. 8C) sites demonstrate the same pattern of strong dependence of grassland CO_2 exchange on the precipitation regime and are in agreement with the observation by Derner and Hart (2007), who reported a fivefold difference in the peak standing crop in years with dry and wet springs in the mixed-grass prairie near Cheyenne, Wyoming, during 1991–2006.

Source-Sink Activity of Ecosystems of the SGS Ecoregion

Table 4 and Figure 9A show summary characteristics of the CO_2 exchange in ecosystems of the SGS ecoregion. Annual photosynthetic uptake varied greatly from < 1000 to nearly $4000 g CO_2 m^{-2} yr^{-1}$, driven mostly by moisture availability. At the same time, respiration losses varied from 900 to $3700 g CO_2 m^{-2} yr^{-1}$, resulting in performance variation from sinks with NEP up to $700 g CO_2 m^{-2} yr^{-1}$, to sources with NEP as low as $-900 g CO_2 m^{-2} yr^{-1}$. As seen in Fig. 9A, the 95% confidence zone for the (GPP, RE) points of the SGS ecoregion is evenly bisected by the 1:1 diagonal. For comparison, a similar zone for the mixed-grass sites (Fig. 9B, contour M) has a larger portion of the confidence zone below the diagonal, indicating higher sink activity. The greatest proportion of data points falling below the diagonal occurs for the tallgrass sites (Fig. 9B, contour T), emphasizing a predominance of the preburn CO_2 -sink performance of the tallgrass prairies and the potential for significant accumulation of soil organic matter (Derner and Schuman, 2007).

Effects of Grazing

Statistical comparison between ungrazed and grazed site-years using the entire dataset (see Table 2) shows no significant difference

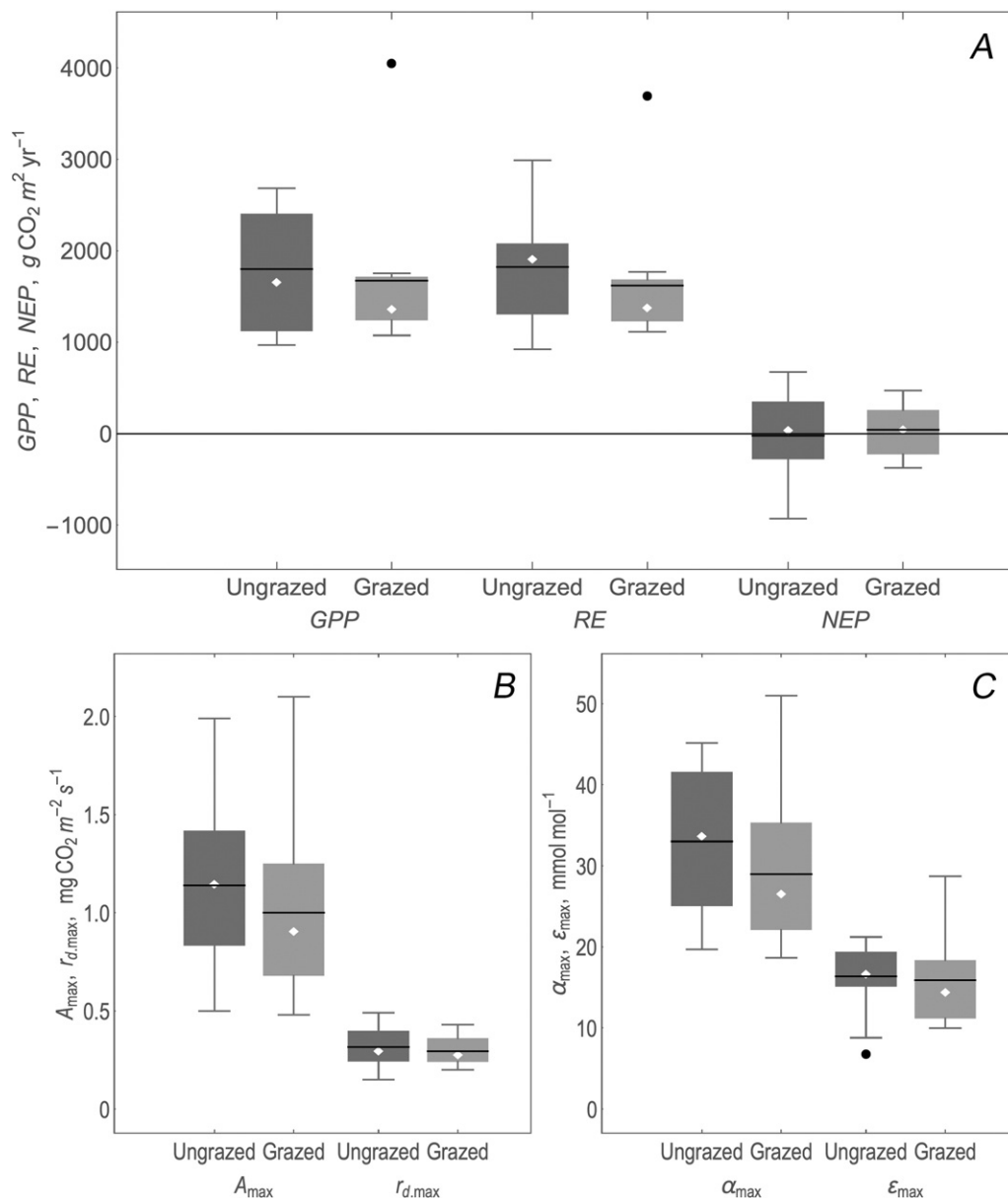


Figure 10. Box-whisker plots of the CO₂ exchange components (A), maximum parameters of photosynthetic capacity and respiration rate (B), and quantum yield and light-use efficiency (C) of the ungrazed and grazed SGS ecosystems.

(two-sided P values > 0.05) between means of gross primary production, ecosystem respiration, net ecosystem production, and ecophysiological parameters A_{max} , $r_{d,max}$, α_{max} , and ϵ_{max} (Fig. 10). The result remains unchanged (P values for differences of the means > 0.05) after exclusion of the outlier (site EC₃ grazed in 2010 with high annual PCPN = 690 mm yr⁻¹), except for annual respiration RE for which the difference between means of ungrazed and grazed site-years becomes significant (two-sided P value = 0.02 < 0.05). This result is in agreement with the observation by Milchunas et al. (2008) that 50-yr-long studies of grazing intensity treatments at SGS sites at the Central Plains Experimental Range (Colorado) showed average forage production rates of 75, 71, 68, and 57 g DM m⁻² yr⁻¹ for ungrazed, lightly grazed, moderately grazed, and heavily grazed treatments, respectively, range productivity being most sensitive to precipitation and soil fertility and, only last, grazing intensity. A comprehensive analysis of the effects of grazing and weather presented in Morgan et al. (2016) demonstrated that weather affects CO₂ fluxes more than grazing practices in SGS ecosystems.

BREB–EC Comparison

The SGS ecoregion dataset provides an opportunity to compare BREB and EC systems data by either examining the concurrent flux measurements or comparing models and parameters on the basis of tower data from the two methods. Measurements from BREB and EC towers parallel and independently collected at similar (37-km distance) heavily and moderately grazed BR₃ and BR₄ sites and the ungrazed and moderately grazed EC₁ and EC₂ sites from 2004 to 2006, postprocessed by the same algorithm, [1]–[5], allow a comparison of the magnitudes and flux patterns at all four sites (Fig. 11). The curves of gross photosynthesis P_g for BREB sites are consistently higher than the curves for EC sites, while the respiration curve R_e for the BREB sites is lower than for the EC sites. Corresponding 3-yr GPP cumulates for the BREB sites (4 106 and 4 588 g CO₂ m⁻²) are higher than for the EC sites (3 676 and 3 982 g CO₂ m⁻²), while the 3-yr RE cumulates for the BREB sites (3 792 and 3 789 g CO₂ m⁻²) are lower than for the EC sites (4 970 and 4 920 g CO₂ m⁻²) (Table 5). This might be considered as

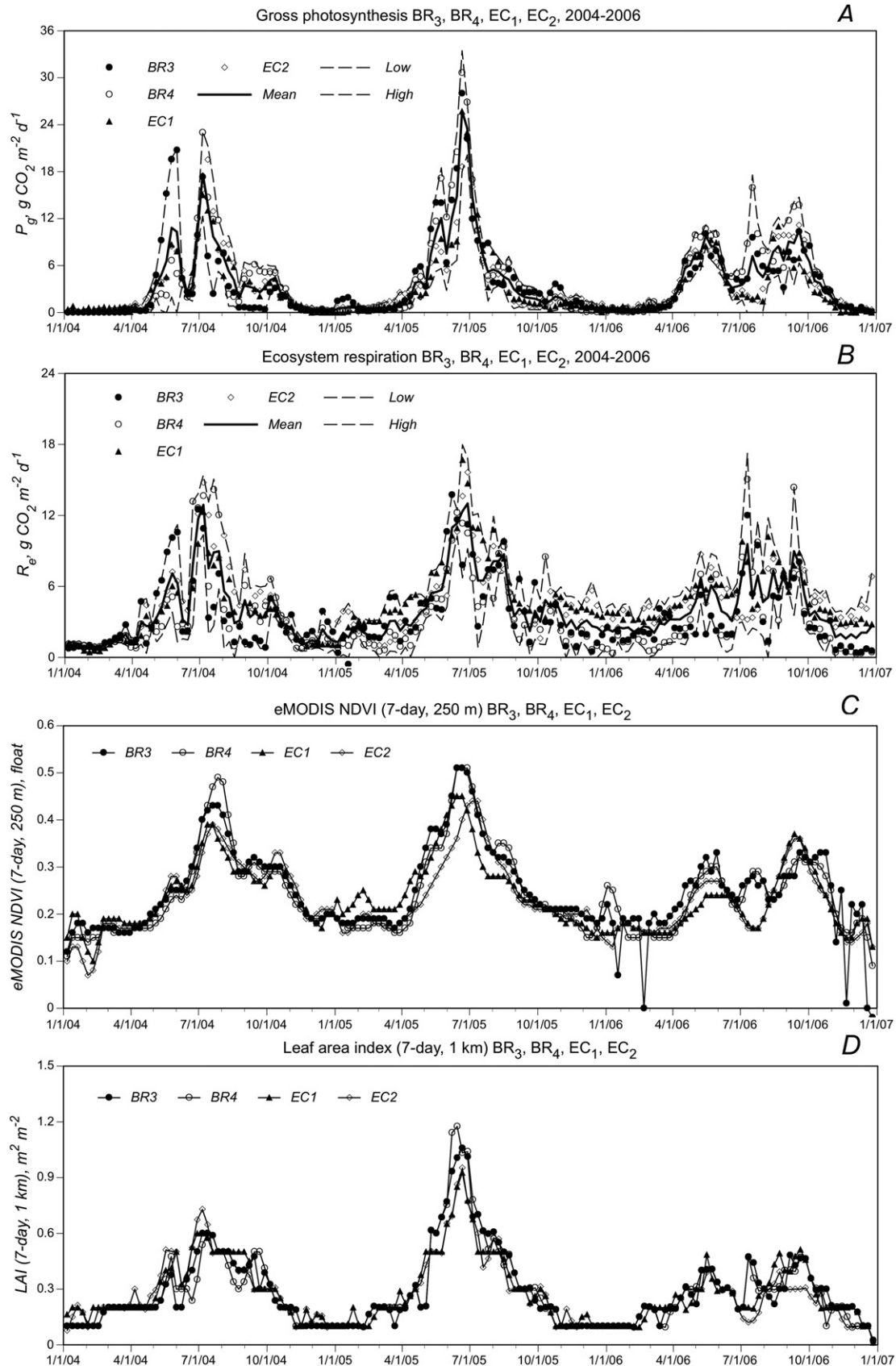


Figure 11. Seasonal and interannual dynamics of the weekly photosynthetic uptake (A), ecosystem respiration (B), eMODIS normalized difference vegetation index (C), and leaf area index (D) at the high continuously grazed (BR₃), moderate continuously grazed (BR₄), ungrazed (EC₁), and moderately grazed (EC₂) shortgrass steppe sites during 2004–2006. “Mean” indicates mean of the four sites (solid line); “Low,” mean is 1.5 standard deviation; “High,” mean is + 1.5 standard deviation.

Table 5

3-yr integrals of the normalized difference vegetation index (NDVI) truncated at the background, leaf area duration, gross primary production, ecosystem respiration, and net ecosystem production at the Bowen ratio–energy balance and eddy covariance sites, 2004–2006.

Grazing regime	Integrated NDVI (3 yr), NDVI d	Leaf area duration (3 yr), d	GPP (3 yr), g CO ₂ m ⁻²	RE (3 yr), g CO ₂ m ⁻²	NEP (3 yr), g CO ₂ m ⁻²
BR ₃ high continuous grazing	76.5	310	4106	3792	314
BR ₄ moderate continuous	85.2	308	4588	3789	799
EC ₁ ungrazed	61.3	309	3676	4970	–1294
EC ₂ moderate spring/fall grazing	45.2	300	3982	4920	–983

GPP indicates gross primary productivity; RE, total ecosystem respiration; NEP, net ecosystem production.

confirmation observations by Alferi et al. (2009) that the BREB method generates higher net CO₂ fluxes than the EC method. However, NDVI curves for the BREB sites lay mostly above the ones for the EC sites (particularly during the periods of maximum NDVI), emphasizing higher general greenness of the BREB compared with the EC sites (Fig. 11C). Comparing the seasonal NDVI integrals showed that 3-yr iNDVI totals for the BREB sites during 2004–2006 are significantly higher than for the EC sites for the same years (76.5 and 85.2 NDVI days vs. 61.3 and 45.2 NDVI d, $P < 0.004$ for all BREB–EC pairs). The LAI data (1-km resolution) demonstrate rather close dynamics at both tower types (Fig. 11D), with the 3-yr totals of leaf area duration (310 and 308 d) at the BREB sites somewhat higher than at the EC sites (309 and 300 d).

Comparison of ecophysiological parameters estimated from BREB and EC towers (Fig. 12) showed no significant differences between centers of the scatter ellipsoids of the BREB-derived and EC-derived parametric clusters.

In the entire four-dimensional space of maximum parameter values $\{(\alpha, \varepsilon, A_{\max}, r_d)\}$, Mahalanobis criterion (Rao, 1965) showed no statistical grounds to reject the hypothesis that the BREB-based centroid $\mu_{\text{BREB}} = (31.9, 16.3, 1.11, 0.33)$, is not significantly different from the EC-based centroid $\mu_{\text{EC}} = (29.8, 15.9, 1.03, 0.26)$ ($P = 0.243 \gg 0.05$). In other words, there is no significant difference between parameter means obtained from BREB or EC towers.

Thus, our study confirms the conclusion by Wolf et al. (2008) that the results of flux measurements by the BREB and EC towers are not essentially different, though “subtle differences” do occur under certain conditions. Figures 11A and 11B show that during the growing season, both of the $P_g(t)$ and $R_e(t)$ curves remain within the $\{\pm 1.5 \text{ SD}\}$ band around the mean of the combined BREB and EC data. The EC respiration estimates outside the growing period are definitely higher than the BREB-based estimates (see Fig. 11B), contributing to the overall higher RE totals for EC compared with BREB (see Table 5). Both methods have technical difficulties during the cold season (Dugas et al., 1999; Gilmanov et al., 2004a; Burba et al., 2008). The higher gross and net CO₂ uptakes recorded at the BREB sites compared with the EC sites

might be related to not only the higher respiration losses at the EC sites but also possibly higher general productivity of the BREB sites (BR₃ and BR₄), as indicated by their higher NDVI and LAI indices (Fig. 11C and D) and higher integrated NDVI and leaf area duration (see Table 5).

Summarizing the rates and patterns of the CO₂ exchange in rangelands of the SGS ecoregion, this study demonstrated considerable variability of daily maxima for photosynthesis $13 < P_g < 41 \text{ g CO}_2 \text{ m}^{-2} \text{ d}^{-1}$, for respiration $13 < R_e < 35 \text{ g CO}_2 \text{ m}^{-2} \text{ d}^{-1}$, and for net daily CO₂ flux $9 < F < 24 \text{ g CO}_2 \text{ m}^{-2} \text{ d}^{-1}$. The annual ranges were $1100 < \text{GPP} < 2700 \text{ g CO}_2 \text{ m}^{-2} \text{ yr}^{-1}$, $900 < \text{RE} < 3000 \text{ g CO}_2 \text{ m}^{-2} \text{ yr}^{-1}$, and $-900 < \text{NEP} < 700 \text{ g CO}_2 \text{ m}^{-2} \text{ yr}^{-1}$. Both the daily and the annual fluxes in shortgrass steppe were lower than in mixed-grass and tallgrass ecosystems. Depending on meteorological conditions each year (chiefly precipitation amount and distribution), either unimodal or bimodal patterns of the seasonal CO₂ exchange were observed and the ecosystem response varied from a net sink of $+700 \text{ g CO}_2 \text{ m}^{-2} \text{ yr}^{-1}$ in 1999 with 487 mm precipitation to a net source of $-900 \text{ g CO}_2 \text{ m}^{-2} \text{ yr}^{-1}$ in 2010 with 173-mm precipitation. Parameterization of CO₂ exchange using the nonrectangular hyperbolic equation with VPD limitation and exponential dependence of respiration on soil temperature provided quantitative estimation of apparent quantum yield α , ecological light-use efficiency ε , photosynthetic capacity $A_{\max}$, and ecosystem respiration rate r_d , including determination of the ranges of variation ($19 < \alpha < 51 \text{ mmol mol}^{-1}$, $7 < \varepsilon < 29 \text{ mmol mol}^{-1}$, $0.48 < A_{\max} < 2.1 \text{ mg CO}_2 \text{ m}^{-2} \text{ s}^{-1}$, $0.15 < r_d < 0.49 \text{ mg CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) and patterns of their seasonal and interannual dynamics. Numerical values of parameters derived from flux tower measurements agree with estimates obtained by other authors using the open-top chamber technique. While there were certain differences in CO₂ fluxes between BREB and EC towers, they occurred occasionally and were not related only to different methodologies because the NDVI and LAI values at the BREB sites were generally higher than at the EC sites. These differences were not significant in terms of the overall seasonal patterns (both systems responded to meteorological drivers with trajectories of P_g and R_e remaining within a $\pm 1.5 \text{ SD}$ band around

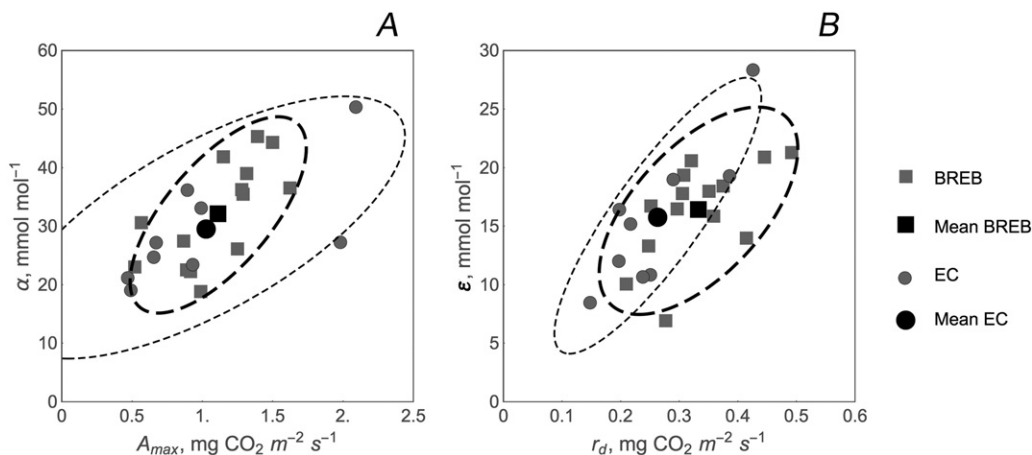


Figure 12. Comparison of parameters estimated from Bowen ratio–energy balance and eddy covariance (EC) towers: (A) maximum apparent quantum yield α versus maximum photosynthetic capacity $A_{\max}$; (B) maximum light-use efficiency ε versus maximum daytime respiration rate r_d . Black quadrat indicates mean of the BREB estimates; black circle, mean of the EC estimates; thick ellipses, 95% zone for BREB parameters; thin ellipses, 95% zone for EC parameters.

the mean) or in the magnitudes of annual GPP, RE, and NEP totals and parameters. This suggests a need to consider including the legacy of the BREB CO₂ exchange data, especially extensive in rangeland ecosystems, in ongoing efforts of comparative analysis, synthesis, and upscaling of rangeland CO₂ exchange and productivity.

Implications

The light-soil temperature-VPD-based method provides consistent partitioning of the raw flux tower CO₂ exchange measurements in shortgrass rangelands into gross primary production (GPP) and total ecosystem respiration (RE) components while avoiding overestimation of respiration inherent to the widely used partitioning method based on nighttime temperature response functions. No significant differences were observed between GPP and RE estimates and CO₂ exchange parameters at SGS sites with BREB and EC tower types. Long-term flux tower measurements demonstrate that SGS ecosystems switch from a sink to a source for atmospheric CO₂ depending on weather. Quantification of the rates and parameters of the SGS ecosystems presented in this paper strengthens the empirical basis for sB60patiotemporal modeling and upscaling as tools for forage production and carbon management of the SGS rangelands.

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