



Playing Into the Hands of the Powerful: Extracting “Success” by Mining for Evidence in a Payments for Environmental Services Project in Matiguás-Río Blanco, Nicaragua

Authors: Van Hecken, Gert, Kolinjivadi, Vijay, Huybrechs, Frédéric, Bastiaensen, Johan, and Merlet, Pierre

Source: Tropical Conservation Science, 14(1)

Published By: SAGE Publishing

URL: <https://doi.org/10.1177/19400829211020191>


BioOne Complete (complete.BioOne.org) is a full-text database of 200 subscribed and open-access titles in the biological, ecological, and environmental sciences published by nonprofit societies, associations, museums, institutions, and presses.

Your use of this PDF, the BioOne Complete website, and all posted and associated content indicates your acceptance of BioOne's Terms of Use, available at www.bioone.org/terms-of-use.

Usage of BioOne Complete content is strictly limited to personal, educational, and non - commercial use. Commercial inquiries or rights and permissions requests should be directed to the individual publisher as copyright holder.

BioOne sees sustainable scholarly publishing as an inherently collaborative enterprise connecting authors, nonprofit publishers, academic institutions, research libraries, and research funders in the common goal of maximizing access to critical research.

Playing Into the Hands of the Powerful: Extracting “Success” by Mining for Evidence in a Payments for Environmental Services Project in Matiguás-Río Blanco, Nicaragua

Tropical Conservation Science
 Volume 14: 1–8
 © The Author(s) 2021
 Article reuse guidelines:
sagepub.com/journals-permissions
 DOI: 10.1177/19400829211020191
journals.sagepub.com/home/trc


Gert Van Hecken , Vijay Kolinjivadi, Frédéric Huybrechs, Johan Bastiaensen, and Pierre Merlet

Abstract

Payments for Environmental Services (PES) are premised upon the provision of monetary incentives to induce land-use practices viewed to be beneficial for advancing tropical conservation. A recent article published by Pagiola et al. in this journal claims that PES successfully transitioned land-use from agricultural use in Matiguás-Río Blanco, Nicaragua to silvopastoralism through afforestation and hence associated improvements in carbon sequestration and biodiversity conservation. Building on contrasting perspectives from peasants and local organizations in the region for more than a decade, we illustrate why viewing relations like payment provision and adoption of land-use outcomes that disregard parallel voices of implicated actors is not only analytically imprecise, but risks being anti-ecological if such a decontextualized connection is used to show evidence that tropical conservation is being advanced. We argue that the effect of payments must be contextualized with: a) increasingly globalized and expanding commodity frontiers for which PES programs may actually further advance to the detriment of tropical conservation; and b) the assumptions made in the methodological approaches adopted to determine causality. In sum, we highlight the dangers of uncritically portraying narratives of “success” to scale up investment to further proliferate decontextualized conservation projects that may not ensure long-term outcomes. We propose responding to these potential dangers through more open, horizontal, and long-term engagement on both the criteria and the consequences of defining success in tropical conservation interventions with actors whose lives are directly affected by them.

Keywords

power, Central America, Nicaragua, political ecology, human–nature relations, conservation projects, social-ecological systems, payments for ecosystem services

In a recent contribution to this journal, Stefano Pagiola, Jordi Honey-Rosés and Jaume Freire-González (Pagiola et al., 2020) assess the permanence of restoration-oriented silvopastoral land use changes in relation to a short-term Payment for Environmental Services (PES) flagship project: the Regional Integrated Ecosystem Management Project (RISEMP) funded by the GEF and implemented by the World Bank between 2003 and 2008 in Colombia, Costa Rica and Nicaragua. They revisit the Nicaraguan project site in Matiguás-Río Blanco in 2012, i.e. four years after project completion, and assess how payments have induced farmers to adopt and sustain the promoted silvopastoral practices. By surveying a subset of former project participants, the authors narrow down 5 hypotheses that, as we describe

below, are disengaged from actual grounded evidence of this particular project. In addition to *a priori* establishing core assumptions of human behaviour and ecological relationships, they selectively analyse data to assemble evidence claiming that payments were successful in inducing both short-term and longer-term land use

Institute of Development Policy and Management, University of Antwerp, Belgium

Received 18 December 2020; Accepted 6 May 2021

Corresponding author:

Institute of Development Policy and Management, University of Antwerp, Prinsstraat 13, B-2000 Antwerp, Belgium.
 Email: gert.vanhecken@uantwerpen.be



Creative Commons Non Commercial CC BY-NC: This article is distributed under the terms of the Creative Commons Attribution-NonCommercial 4.0 License (<https://creativecommons.org/licenses/by-nc/4.0/>) which permits non-commercial use, reproduction and distribution of the work without further permission provided the original work is attributed as specified on the SAGE and Open Access pages (<https://us.sagepub.com/en-us/nam/open-access-at-sage>)

changes and to “rule out alternative explanations” (p. 1). As a group of scholars who have been conducting extensive research on PES policies and projects, including on the above-mentioned project, we have serious concerns about the findings of the article, as well as the broader implications of evaluations of conservation projects oriented to showcase “success stories” for replication and scale-up to other contexts.

In the following sections we reassess the authors’ evidence and conclusions in view of alternative explanations and findings that shed a different light on the project’s short and longer-term results. We thereby highlight two dangers of the original assessment, which range from structural and historical patterns of land-use change in the region that have been largely ignored to methodological flaws of the research. We argue that being transparent about underlying knowledge and value frameworks upheld by researchers and actors implicated in conservation projects is a fundamental aspect of conservation science and practice and therefore should be a crucial element in any attempt to evaluate PES projects. We further caution against attempts to showcase success stories through analyses that tend to disregard many relevant contextual elements and contrasting evidence. The failure to be transparent about diverging value frameworks that underpin conservation research directly impacts how success gets defined and evaluated (Bromley, 2012; Joslin, 2019; Svarstad & Benjaminsen, 2017). As Nelson and Bigger (2021) recently argue, interventions like PES that encourage tropical conservation do not necessarily supplant extractivist projects that are otherwise counterproductive to conservation; instead, they may simply build on underlying drivers for intensive commodity production like timber, or securitize the ecological soil conditions for sustaining intensive and ecologically-destructive forms of agriculture. In this sense, generating success stories is akin to a site of extraction where specific evidence is literally mined, sometimes unconsciously, to reflect a particular epistemological framing that is blind to context and which reinserts decontextualized evidence to justify and expand a global conservation value chain. This value chain emerges when conservation success stories beget the endless growth of bankable conservation projects worthy of investment, while lacking the capacity to confront the longer term historical drivers of biodiversity loss and even work to encourage the political and economic systems that exploit both humans and non-human natures.

Mining for ‘Success’ in PES Evaluation

The primary claim of Pagiola et al. (2020) is that the provision of payments is the core reason for greening land uses. In other words, ‘success’ (and permanence

of this ‘success’) hinges on establishing a linear causal relationship between financial incentives and ensuing land use changes. However, we show that in the RISEMP case, project payments alone are an insufficient explanation of the underlying socio-cultural and economic drivers that were actually influencing land use changes.

To corroborate the potential link between the PES incentives and desired land use change, we have reassessed the project’s results (Vaessen & Van Hecken, 2009; Van Hecken et al., 2015; Van Hecken & Bastiaensen, 2010) and contextualized them within a broader series of (ongoing) studies in the region (Bastiaensen et al., 2015; Polvorosa, 2013). This included an analysis of the original project surveys, as well as the results of its randomized control trial (RCT) experiment, complemented with a two-month qualitative field study in 2009 encompassing in-depth interviews with thirty-five participant and non-participant farmers and with local project staff. Additionally, in 2013 follow-up interviews were conducted with thirty-two project participants, which provided insights in the multiple motivational underpinnings of land-use changes and social-ecological dynamics in the region since project termination. The main findings of these studies were also validated with former local project staff during workshops in 2013 and 2014. Finally, our analysis is also complemented by in-depth studies into processes of agrarian change in the context of cattle-based agricultural frontier expansion in the broader region, encompassing the area in which the RISEMP project took place (Bastiaensen et al., 2015, 2021; Van Hecken et al., 2019). We draw upon these previous studies to illustrate the diverging ways in which success can be interpreted from the RISEMP project, and in doing so highlight why context matters for conservation projects.

Dismissal of Structural and Historical Patterns of Land-Use Change

The first observation that we have demonstrated in previous work is that understanding land-use change dynamics in Matiguás-Rio Blanco (as anywhere else!) cannot be isolated from the broader historical processes in the region that are transforming the landscape. Such inquiry offers important insights on *how the PES project has interplayed with multi-scalar land-use dynamics*, and as such, allows for a more meaningful understanding of farmer motivations and the potential effect and “permanence” of the intervention. As we describe below, attention to history and context is not a mere detail to understanding the success of conservation interventions like PES, but plays a central role in the way such interventions unfold.

Though absent from Pagiola et al. (2020)'s account, historical economic drivers around dairy production are a key element for understanding land-use dynamics in the study region of the RISEMP project. Until the 1970s, most communities in the region were relatively isolated and lacked access to milk storage facilities, which made the commercialisation of fresh milk unattractive (Bastiaensen et al., 2015). However, increasing interest from international dairy companies through the 1980s and 1990s led to a rapid growth of roads and basic milk storage facilities, which opened up the fresh milk market for many farmers and led to an intensification of dairy production. In the late 1990s, Salvadoran cheese traders further commercialised dairy production and offered higher, but seasonally fluctuating milk prices (Levard et al., 2001), and most importantly, generated an additional, more lucrative market niche compared to the local markets. Our previous analyses have also shown that regional fresh milk prices have on average increased by 10% between June 2004 and September 2006, *coinciding* with the RISEMP project implementation (Van Hecken & Bastiaensen, 2009). In order to benefit from these dairy market opportunities, farmers need to keep (and feed) their milking cows close to the roads where fresh milk for the industry is collected. Moreover, since the more productive crossbred dairy cattle breeds in the region (Brown-Swiss or Holstein combined with Brahman cattle) are less resistant to heat, they require the protection of shade from trees (Yamamoto et al., 2007). Collectively, these factors have increasingly encouraged producers to invest in stable milk production year-round, which is made possible by silvopastoralism that ensures continuous cattle feed for secured access to dairy markets (Polvorosa, 2013). These practices that also optimise stocking rates of cattle include those promoted by the RISEMP project (e.g. improved pastures with trees, living fences, fodder banks, increase of tree density within pasture areas).

This underlying historical and economic context substantially calls into question Pagiola et al. (2020)'s exclusive attribution of positive changes in farmers' behaviour to the payments made (further reasons are described in Methodological Mismatches section). Furthermore, growing opportunities in the emerging dairy markets have indirectly influenced a wide range of governmental and non-governmental agencies entering the area to promote silvopastoral systems as efficient technical solutions to enhance livestock production (Van Hecken & Bastiaensen, 2009). The World Bank's RISEMP project intervention aimed to do exactly the same thing, but used conservation as a justification for what is evidently a process of land-use intensification for industrialized dairy markets. As a multilateral development institution, The World Bank should be all too cognizant of these

dynamics. The assumed conservation "success" therefore must be interpreted in this light.

Additionally, a rebound effect becomes likely when any localized ecological gains are compensated by increasing land brought under intensified production systems at a more regional scale (often referred to as 'leakage'). Indeed, as demonstrated in various studies in the study site, the desired land use changes associated with the PES project were mostly being adopted by the largest and most wealthy cattle farmers who were already deeply engaged in the burgeoning dairy markets and intensification of cattle stocking (Mairena et al., 2010; Polvorosa, 2013). Considering PES payments in isolation ignores how land at the regional level has become increasingly scarce due to the consolidation of agricultural land-holdings by a minority of well-resourced cattle farmers engaged in this booming dairy market, whereby opportunity costs of land have become significantly higher. Farmers who do not have sufficient cattle and who are unable to shift from agricultural crops to strictly dairy-oriented operations are literally dispossessed of their existing livelihoods and eventually obliged to sell their lands to the larger and wealthier cattle farmers and migrate to 'new' agricultural frontiers (Van Hecken et al., 2015). Incidentally, it is precisely these larger and wealthier cattle farmers that capture the majority of the PES payments (Van Hecken & Bastiaensen, 2010).

Pagiola et al. (2020) do mention outmigration (p. 7), yet only from the perspective of it being an unfortunate methodological issue of the PES follow-up evaluation rather than as a core element of the agrarian dynamics in the region. Accordingly, a crucial aspect in understanding how such dynamics influence the effectiveness of project payments in inducing long-term ecological benefits is missing. The broader context illustrates how the PES payments can even have the unintended effect of worsening social inequalities and ecological outcomes by *actually encouraging* further intensification of cattle-raising production for the dairy markets, while pushing farmers who are unable to compete in this market to deforest land elsewhere. Our research found that in 2013, i.e. only five years after project termination, twenty-nine (22%) of the original 123 farmer participants in Matiguás-Río Blanco had sold their properties. Most of these emigrants were small or middle-income farmers who had moved eastward to the new agricultural frontier in search of cheaper land (Van Hecken et al., 2015). This is not merely a problem of attrition, but a clear indicator of the underlying, inequitable agrarian dynamics of the dairy-cattle context. Delinking the move to silvopastoralism from these broader land-use dynamics results in an incomplete and potentially misleading assessment of the social and environmental outcomes of the PES project.

Lastly, it is relevant to reflect on why an ahistorical approach to assessing PES effectiveness has important implications for understanding the assumed ‘permanence’ of the induced land use changes. For example, silvopastoral land-use practices within peasant-directed agroecological systems will produce very different ecologies than they would do in intensively managed cattle-stocking environments destined for an increasingly competitive and globalized dairy market (Cochet, 2012). In this regard, reflections originating from historicized systems analysis encompassing alternative explanations based on plural knowledge frameworks are imperative to more accurately make scientific assessments on the robustness of assessing outcomes from conservation interventions. A totalizing lens that treats non-human natures as commodified resources and humans as labourers for capital-intensive cattle production generates its own ecological consequences by cheapening nature and human social systems into the logic of profit orientation *regardless of the land-use system adopted* (e.g., Moore, 2015). We argue that the blindspot of some conservationists and economists’ focus on the ‘permanence’ of maintaining a desired land-use in isolation of these considerations is *fundamentally anti-ecological*. It also reflects a deep-seated flaw in the positivist ‘ideal type’ PES conceptualizations that Pagiola et al. (2020) and other influential PES theorists and scholars (e.g., Engel & Wünscher, 2015; Wunder, 2015) have either not addressed or continue to pursue.

Methodological Mismatches

In addition to the above arguments, the methodology adopted by Pagiola et al. (2020) to evaluate the program’s “success” raises further questions regarding the rigor of the analysis. In order to attribute desired land-use changes to the induced payment, the project initially designed a randomised control trial (RCT) in which 123 farmers were subdivided into three different experimental groups, receiving payments only (28 farmers), payments and technical assistance (TA) (70 farmers), or neither of these in serving as a control group (25 farmers). This setup was intended to separate the effect of PES and TA from other (undefined) contextual variables in order to assess their relative effectiveness. Not

surprisingly, however, and as explained extensively elsewhere (Vaessen & Van Hecken, 2009; Van Hecken & Bastiaensen, 2010), the RCT design proved to be problematic and far from random in practice, making it impossible to distinguish between the different ‘treatment’ and control groups as if the project took place in an experimental laboratory instead of the real world. The follow-up survey, with a subset of the original participants that Pagiola et al. (2020) refer to, and the reference to land use maps (which are absent for the reader to verify) attempt to make claims about the short-term and long-term additionality of payments-induced land use changes. However, without any meaningful ‘counterfactual’ or contextualized analysis, Pagiola et al. (2020) claim that project payments, rather than TA or any other possible motivation, were most likely to have accounted for the observed land use changes. In order to back-up the claim of project success, they rely on “casual observations” that indicate that neighbouring areas not included in the project observed “very few land-use changes” (p.8). We are left wondering how “casual observations” can ever rigorously justify such an important claim, given the context that we have described in the previous sections.

The fact that the authors decided to initially use an RCT¹ to assess the program’s success and then subsequently proceed to selectively disregard elements of the RCT (when the expected outcomes were not forthcoming, as noted below), demonstrates how they carefully cherry-pick the evidence that supports their claims of success. Our comparison of the different treatment/control groups (Tables 1 and 2) indicates that dismissing the control group from the analysis, as the authors have done with the argument that the control group was “highly suspect” (p. 7), denotes a strong selection bias to support their claims and to dismiss other explanations that fall outside of their interpretive framework. Closer examination of the farmers in the control group reveals that these individuals were among the wealthiest and largest land-owners, having on average more land and greater stocking of cattle, and a smaller household size (Table 1). These factors alone invalidate the control group as a reliable counterfactual for the RCT. Yet, it remains relevant to observe that these land-owners had

Table 1. Project Participants by Treatment Group, 2002 (Mean and Standard Deviation).*

Treatment group	Group size	Area (ha)	Household size	Number of cattle
PES only	28	29.5 (25.1)	6.0 (2.7)	34.0 (35.5)
PES+TA	70	31.9 (25.8)	6.3 (2.5)	34.8 (30.9)
Control	25	46.7 (37.1)	5.3 (1.9)	53.6 (39.2)
Total	123	34.4 (28.8)	6.0 (2.5)	38.5 (34.4)

*Standard deviations indicated between parentheses.
Source: Vaessen and Van Hecken (2009).

Table 2. Changes in Land Use, per Treatment Group (*), 2003–2007.

Land use group	Crop	TP	DP	NP-T	IP-T	NP+T	IP+T	FB	Tac	TF	LF**
PES+TA	−6.0	−0.4	−19.4	0.0	0.3	4.2	14.1	5.5	0.5	0.8	213.5
PES	−5.8	0.8	−21.7	−0.9	−0.1	2.0	21.5	4.9	−2.3	2.2	164.4
Control	−0.4	−4.0	−23.0	2.3	0.1	5.5	11.3	4.9	−0.4	−0.5	1364.4

Legend: Crop = annual crops; TP = total pastures; DP = degraded pastures; NP-T = natural pastures without trees; IP-T = improved pastures without trees; NP+T = natural pastures with trees; IP+T = improved pastures with trees; FB = fodder banks; Tac = tacotales (scrub habitats); TF = total forests; LF = living fences.

*Changes in land use: the additional percentage of the selected land use within the total land size of every treatment group.

**The per cent for LF is calculated as an increase in the length, compared to each group's initial LF length in 2003.

Source: Van Hecken and Bastiaensen (2010).

among the most substantial outcomes in terms of enhanced silvopastoral land-use (Table 2), suggesting once again that the authors' claim that observed land-use changes are attributable to the payment incentive alone is problematic.

Table 2 gives an overview of land use changes across the two treatment and the control groups, showing that degraded pastures have decreased in all groups, with the highest reduction paradoxically in the control group (−23 per cent). Furthermore, the implementation of living fences have also increased most in the control group (eight times more than in the PES group). Fodder banks show a very similar pattern across all groups, while the highest increase in natural pastures, with and without trees, was also in the control group. Due to the outstanding results of the control group, it is baffling to us as to why Pagiola et al. (2020) failed to mention these surprising outcomes and then proceeded to dismiss any kind of engagement with the findings of this control group by arguing that its inclusion in the analysis would bias the supposed objectivity of the RCT experiment. This makes their argument a circular one in which the authors disregard or even *purposefully dismiss*² the relevant information these cases provide reflecting the political and economic dynamics that are operating in this landscape.

Implications for Conservation: Missing the Forest for the Trees

Tropical conservation depends on *regenerating* the intimate ecological functions and relations that sustain topsoils, groundwater, biodiversity, and which regulate the climate. These should be among the core aspects that count in assessing the additionality of PES projects. In this sense, the putative purpose of PES interventions is to reduce land degradation in the long-term by generating the conditions that attend to these dynamic relations and motivations and which crucially involve both human-social and non-human-biophysical entanglements. However, in our view, focusing only on the

adoption of particular land-uses on the farm plot disregards the political-economic systems that permit or obstruct restoration and regeneration of ecological functions at a landscape level. In the case of the RISEMP project, this absence of analytical precision on the part of determining “success” is resulting in a PES program that piggybacks on underlying land use dynamics that encourage increased cattle-stocking (and the ecological consequences these entrain at the project site) as well as the expansion of the agricultural frontier elsewhere, further exacerbating tropical deforestation.

Moreover, reducing phenomena to the sum of their parts by concentrating only on the effect of payments on land-use change, in our perspective, ultimately circumscribes both human behaviour and ecological relations in such a narrow way as to effectively render futile and even counterproductive any intervention that claims to be about conservation. To avoid the risk of simplifying complexity to justify donor spending by selectively engaging with the social and political context of conservation interventions, we underscore the importance of inserting alternative knowledge framings to bring rigour to the core of tropical conservation science. Barring these considerations, the PES recipe becomes a vacuous exercise of *mining for evidence of success* with minimal or even counterproductive implications for conservation. We therefore echo others who have warned against the fetishization of PES blueprints (e.g. Ervine, 2010) through the deployment of new bankable PES projects in other locations (e.g. Colombia, Brazil, and elsewhere in Nicaragua, as the authors also hint at in the conclusion of the article). Instead, attentiveness to methodological plurality and conceptual openness is crucial to more responsibly inquire into the multifaceted nature of project outcomes and to respect the legitimate perceptions and experiences of a diverse range of actors involved in these interventions.

The implications of the authors' study amplifies the critique of many other scholars across different vantage points within the field of conservation, ranging from the natural to the social sciences. The conclusion of

conservation success from the evidence provided by Pagiola et al. (2020) echoes deep-seated concerns in terms of how communities and ecological relations become caricatured to match the simplistic relation of ecosystem services (Barnaud & Antona, 2014; Tadaki et al., 2015). The authors' decontextualized and reductionist assessment of conservation intervention outcomes strikes a chord with critiques raised by interdisciplinary scholars on the social asymmetries of power that shape environmental governance (Büscher, 2014; Ferdinand, 2019; Myers et al., 2018) as well as the ecological inconsistencies of overly simplified study designs (e.g. Christie et al., 2019). The lack of engagement with existing studies that offer conflicting scientific evidence is itself contrary to the process of doing rigorous conservation science (e.g. Sutherland et al., 2020). Lastly, scholars have noted that conservation interventions may also securitize against potential social and ecological risks to production schedules that ultimately serve to worsen social and ecological degradation in the long-term (e.g., Battistoni, 2017; Nelson & Bigger, 2021; Pasgaard et al., 2017).

The concerns we have raised are by no means limited to PES implementation in Nicaragua, but also reflect the ways that adverse ecological outcomes result from conservation interventions more generally when attention to land-use drivers, power asymmetries, human-nature relations, and conceptual framing and research methods are poorly considered. In another example of payments for forest conservation in China's Sloping Land Conversion Program for instance, Li et al. (2015) question the long-term ecological outcomes of rural depopulation and urban migration as new forms of cheap labour to produce manufactured goods to global markets. Such transnational forms of 'leakage' are deeply embedded in the broader supply chains of resources and labour that ultimately exacerbate ecological breakdown that conservation is meant to address. Elsewhere in Canada, Kolinjivadi et al. (2020) highlight how the implementation of buffer zone conservation strips ultimately induce farmers to acquire more land to bring into cultivation in order to meet ever tightening quotas for industrial french-fry production within a growing fast-food industry. Failure to generate specified quotas has led to a form of corporate blackmail, obliging producers, and indeed governments across scales, to ensure production is met or risk large-scale potato processors picking up and leaving for greener, more lucrative pastures. Once again, conservation that ignores the political economy of production -paradoxically then- serves to hasten land-use degradation.

There is a greater need for transparency on the value and knowledge frameworks that define how criteria for conservation success are determined and interpreted. Such reflections are particularly relevant when research

is carried out by project implementers themselves. The World Bank in particular plays a commanding role in funding PES projects around the world, yet simultaneously encourages outside investors to turn so-called "underutilized" land into value-added industrial-scale production for integration into global commodity circuits (e.g. Borrás et al., 2007; Li, 2014). As such, more reflection is needed by researchers and project evaluators as to the political and economic interests that underpin their positions and affiliations when carrying out these kinds of studies that directly intervene in the everyday lives and life choices of peasants and smallholders around the world. Great care is required so as to be responsible and attentive to the unintended ripple effects that such projects might generate.

It is these points that we hope will serve as a springboard for more open, horizontal, and long-term dialogue and reflection between participants, project promoters, scholars, activists, and all other directly and indirectly involved actors in PES projects around the world. To this end, we call for a deliberative platform to unpack PES interventions that are interpolated by plural knowledge and value frameworks in order to examine how socio-political and ecological dynamics influence long-term outcomes and their implications for conservation. In short, greater attention to "ePEStemology" is required in exposing how uneven power relations shape the ways PES interventions unfold in practice. In this way, it becomes possible to better understand how conservation "success" is defined and by whom, while fostering much needed dialogue between diverse social actors where diverging positions exist. We believe improved transparency on conservation values will enable a stronger collaborative dimension in identifying the long-term outcomes of conservation interventions. Given the continued and substantial interest in PES for responding to social and ecological challenges around tropical conservation, such an initiative is urgently needed.

Declaration of Conflicting Interests

The author(s) declared no potential conflicts of interest with respect to the research, authorship, and/or publication of this article.

Funding

The author(s) disclosed receipt of the following financial support for the research, authorship, and/or publication of this article: This study received financial support from the Belmont Forum and NORFACE Joint Research Programme on Transformations to Sustainability, co-funded by ANR, FWO, ISSC and the European Commission through Horizon 2020 (TruePATH project, Grant 730211). G. V. H. and V. K. also acknowledge the financial support of an FWO-FRQ bilateral Flanders-Québec research cooperation grant

(ePEStemology project, Grant G0E8121N), and an FWO Senior Postdoctoral Fellowship (Grant 12ZA921N).

Acknowledgments

The authors acknowledge the constructive feedback from two anonymous reviewers. G. V. H., F. H., J. B., and P. M. acknowledge the financial support from the Belmont Forum and NORFACE Joint Research Programme on Transformations to Sustainability, co-funded by ANR, FWO, ISSC and the European Commission through Horizon 2020 (TruePAT H project, Grant 730211). G. V. H. and V. K. also acknowledge the financial support of an FWOFRQ bilateral Flanders-Québec research cooperation grant (eP EStemology project, Grant G0E8121N), and an FWO Senior Postdoctoral Fellowship (Grant 12ZA921N). The responsibility for the positions expressed rests exclusively with the authors.

ORCID iD

Gert Van Hecken  <https://orcid.org/0000-0001-8496-3194>

Notes

1. At a more epistemological level, RCTs are known to be particularly ill-suited for analyzing change in complex and dynamic social and ecological contexts (Bédécarrats et al. 2019).
2. At various occasions, several of us have attempted to engage in a discussion with the first author (Stefano Pagiola), ranging from emailing him our manuscripts and feedback on the RISEMP project, to approaching him at conferences, and trying to establish a meeting during one of his visits to Nicaragua. However, despite these attempts to engage over the years, we received no response, even if as a rebuttal or to push back on some of the claims we have made. These efforts illustrate that the authors cannot claim that they were unaware of other studies and the alternative explanations that have been identified. Indeed, a recent article by Rasch et al. (2021), analysing the effectiveness and permanence of the RISEMP project in Costa Rica, echoes many of the concerns we raise here and which we have raised previously. Thus we are not the only ones who have repeatedly and continue to make these nuanced claims regarding the outcomes of PES projects in the region.

References

- Barnaud, C., & Antona, M. (2014). Deconstructing ecosystem services: Uncertainties and controversies around a socially constructed concept. *Geoforum*, 56, 113–123.
- Bastiaensen, J., Huybrechs, F., Merlet, P., Romero, M., & Van Hecken, G. (2021). Fostering bottom-up actor coalitions for transforming complex rural territorial pathways. *Current Opinion in Environmental Sustainability*, 49, 42–49.
- Bastiaensen, J., Merlet, P., & Flores, S. (Eds.). (2015). *Rutas de desarrollo y territorios humanos en la vía láctea de Nicaragua: Estudios desde muy muy, matiguás y río blanco* [Development routes and human territories in the Nicaraguan milky way: Studies from muy muy, matiguás and Río Blanco]. UCA Publications.
- Battistoni, A. (2017). Bringing in the work of nature: From natural capital to hybrid labor. *Political Theory*, 45(1), 5–31.
- Bédécarrats, F., Guérin, I., & Roubaud, F. (2019). All that glitters is not gold. The political economy of randomized evaluations in development. *Development and Change*, 50(3), 735–762.
- Borras, S. M., Carranza, D., & Franco, J. C. (2007). Anti-poverty or anti-poor? The World Bank's market-led agrarian reform experiment in the Philippines. *Third World Quarterly*, 28(8), 1557–1576.
- Bromley, D. W. (2012). Environmental governance as stochastic belief updating: Crafting rules to live by. *Ecology and Environment*, 17(3), 14–21.
- Büscher, B. (2014). Selling success: Constructing value in conservation and development. *World Development*, 57, 79–90.
- Christie, A. P., Amano, T., Martin, P. A., Shackelford, G. E., Simmons, B. I., & Sutherland, W. J. (2019). Simple study designs in ecology produce inaccurate estimates of biodiversity responses. *Journal of Applied Ecology*, 56(12), 2742–2754.
- Cochet, H. (2012). The système agraire concept in francophone peasant studies. *Geoforum*, 43(1), 128–136.
- Engel, S., & Wünscher, T. (2015). Payments for ecosystem services as a mechanism to promote biodiversity conservation in a green economy—Potentials and limitations. In A. Gasparatos & K. J. Willis (Eds.), *Biodiversity in the green economy* (pp. 240–261). Routledge.
- Ervine, K. (2010). Participation denied: The global environment facility, its universal blueprint, and the Mexico-Mesoamerican biological corridor in Chiapas. *Third World Quarterly*, 31(5), 773–790.
- Ferdinand, M. (2019). *Une écologie décoloniale: Penser l'écologie depuis le monde caribéen* [A decolonial ecology: Thinking ecology from the Caribbean world]. Editions du Seuil.
- Joslin, A. J. (2019). Unpacking 'success': Applying local perceptions to interpret influences of water fund payments for ecosystem services in the Ecuadorian Andes. *Society & Natural Resources*, 32(6), 617–637.
- Kolinjivadi, V., Bissonnette, J.-F., Zaga Mendez, A., & Dupras, J. (2020). Would you like some fries with your ecosystem services? McDonaldization and conservation in Prince Edward Island. *Geoforum*, 111, 73–82.
- Levard, L., Lopez, Y. M., & Navarro, I. (2001). *Municipio de matiguás: Potencialidades y limitantes del desarrollo agropecuario* [Matiguás Municipality: Potentialities and Limits of Agricultural Development]. UCA/FIDA/Nitlapán.
- Li, H., Yao, S., Yin, R., & Liu, G. (2015). Assessing the decadal impact of china's sloping land conversion program on household income under enrollment and earning differentiation. *Forest Policy and Economics*, 61, 95–103.
- Li, T. M. (2014). What is land? Assembling a resource for global investment. *Transactions of the Institute of British Geographers*, 39(4), 589–602.
- Mairena, E., Pérez, F. J., Hernández, X., Wilson, C., & Pavón, K. (2010). *Contratos comunitarios forestales y sus arreglos de distribución de beneficios* [Community forestry contracts and their benefit sharing arrangements]. Nitlapan-UCA.

- Moore, J. W. (2015). *Capitalism in the web of life: Ecology and the accumulation of capital* (1st ed.). Verso.
- Myers, R., Larson, A. M., Ravikumar, A., Kowler, L. F., Yang, A., & Trench, T. (2018). Messiness of forest governance: How technical approaches suppress politics in REDD+ and conservation projects. *Global Environmental Change*, *50*, 314–324.
- Nelson, S. H., & Bigger, P. (2021). Infrastructural nature. *Progress in Human Geography*. Advance online publication. <https://doi.org/10.1177/0309132521993916>
- Pagiola, S., Honey-Roses, J., & Freire-Gonzalez, J. (2020). Assessing the permanence of land-use change induced by payments for environmental services: Evidence from Nicaragua. *Tropical Conservation Science*, *13*, 194008292092267–194008292092215.
- Pasgaard, M., Van Hecken, G., Ehammer, A., & Strange, N. (2017). Unfolding scientific expertise and security in the changing governance of ecosystem services. *Geoforum*, *84*, 354–367.
- Povorosa, J. (2013). *Opportunities and constraints for small and medium-sized dairy farmers in the context of the booming dairy value chain in Nicaragua: A case study from Matiguás* [PhD dissertation]. Institute of Policy Development and management (IOB), University of Antwerp.
- Rasch, S., Wünsch, T., Casasola, F., Ibrahim, M., & Storm, H. (2021). Permanence of PES and the role of social context in the regional integrated Silvo-pastoral ecosystem management project in Costa Rica. *Ecological Economics*, *185*, 107027.
- Sutherland, W. J., Alvarez-Castañeda, S. T., Amano, T., Ambrosini, R., Atkinson, P., Baxter, J. M., Bond, A. L., Boon, P. J., Buchanan, K. L., Barlow, J., Bogliani, G., Bragg, O. M., Burgman, M., Cadotte, M. W., Calver, M., Cooke, S. J., Corlett, R. T., Devictor, V., Ewen, J. G., . . . Wordley, C. (2020). Ensuring tests of conservation interventions build on existing literature. *Conservation Biology*, *34*(4), 781–783.
- Svarstad, H., & Benjaminsen, T. A. (2017). Nothing succeeds like success narratives: A case of conservation and development in the time of REDD. *Journal of Eastern African Studies*, *11*(3), 482–505.
- Tadaki, M., Allen, W., & Sinner, J. (2015). Revealing ecological processes or imposing social rationalities? The politics of bounding and measuring ecosystem services. *Ecological Economics*, *118*, 168–176.
- Vaessen, J., & Van Hecken, G. (2009). *Assessing the potential for experimental evaluation of intervention effects: The case of the regional integrated silvopastoral approaches to ecosystem management project (RISEMP)*. Impact Evaluation Information Document 15. Washington, DC.
- Van Hecken, G., & Bastiaensen, J. (2009). The potential and limitations of markets and payments for ecosystem services in agricultural landscape restoration. Critical reflections inspired by an assessment of the RISEMP program in Matiguás-Río blanco, Nicaragua. *Discussion Paper 2009.02*. IOB-University of Antwerp.
- Van Hecken, G., & Bastiaensen, J. (2010). Payments for ecosystem services in Nicaragua: Do market-based approaches work? *Development and Change*, *41*(3), 421–444.
- Van Hecken, G., Bastiaensen, J., & Huybrechts, F. (2015). What's in a name? Epistemic perspectives and payments for ecosystem services policies in Nicaragua. *Geoforum*, *63*, 55–66.
- Van Hecken, G., Merlet, P., Lindtner, M., & Bastiaensen, J. (2019). Can financial incentives change farmers' motivations? An agrarian system approach to development pathways at the Nicaraguan agricultural frontier. *Ecological Economics*, *156*, 519–529.
- Wunder, S. (2015). Revisiting the concept of payments for environmental services. *Ecological Economics*, *117*, 234–243.
- Yamamoto, W., Dewi, I. A., & Ibrahim, M. (2007). Effects of silvopastoral areas on milk production at dual-purpose cattle farms at the semi-humid old agricultural frontier in Central Nicaragua. *Agricultural Systems*, *94*(2), 368–375.