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Examining the use of fecal pellet morphometry to differentiate age classes in Sonoran pronghorn

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Wildlife managers require knowledge of population demographics, yet for low-density, wide-ranging species procuring demographic information is challenging. While accurate abundance estimates can be costly and difficult to obtain, recruitment and survival trends can be used as an alternative indicator of a population’s trajectory. Physical capture has been the traditional practice for obtaining these demographic parameters, yet capture-related stress can lead to reduced levels of fitness, impaired locomotion, or even mortality for some species. Thus, noninvasive sampling methods may provide an alternative to physical capture. Population monitoring of endangered Sonoran pronghorn *Antilocapra americana sonoriensis* is critical for assessing the success of recovery efforts, and monitoring annual survival and recruitment by age class would provide information on the trajectory of population growth. We measured noninvasively collected Sonoran pronghorn fecal pellets collected post-fawning in Arizona, USA and matched to known age animals using fecal DNA genotyping to determine the feasibility of distinguishing age class by pellet dimensions. Based on cross-validation with logistic regression predictive models, we estimated a 98% probability of correct classification of fawn versus yearling and fawn versus adult using pellet width as a single explanatory variable. We could not, however, distinguish between yearling and adult. We additionally evaluated our ability to classify age class of fecal pellets by visual assessment only, and this approach was unreliable. Thus, we recommend measuring pellets for more accurate age classification. This measurement method is simple, relatively inexpensive, and shows potential for use in wild populations of pronghorn to discriminate fawns from other age classes. When combined with individual identification using fecal DNA, this approach could provide better knowledge of recruitment and age-specific survival for this and other species.

A comprehensive understanding of population demographic metrics, such as abundance, survival and recruitment, is important for managing wildlife species, yet these parameter estimates are often expensive and difficult to obtain. Alternatively, in large mammals, managers often monitor trends in recruitment and survival as an indicator of a population’s trajectory (Peak 2003, DeCesare et al. 2012). In some cases, trends may be more easily obtained, particularly for low-density, wide-ranging species which are inherently difficult to monitor due to low detection rates. However, recruitment can be difficult to document in some species as juveniles quickly become the same size as adults, and visual assessment of age is often incorrect (Smith 1988, Garel et al. 2006). Survival among age classes varies broadly, and thus knowledge of age structure is essential to accurately assess population demographics. It is widely accepted that for ungulates, adults generally have higher survival rates and elasticity than juveniles, and adult females typically have higher survival rates than adult males (Gaillard et al. 1998, 2000). However, due to high variability, juvenile survival typically has a greater impact on population dynamics (Gaillard et al. 1998, 2000, Raithel et al. 2007, Harris et al. 2008). Thus for endangered populations, estimates of juvenile survival are often a strong indicator of population health and viability and a valuable metric for managers.

Sonoran pronghorn exist exclusively in the Sonoran Desert of southern Arizona and northern Mexico and are federally listed under the Endangered Species Act (USFWS 1998) and as “most endangered” under CITES Appendix 1 (Hoffmann et al. 2008). While believed to number in the thousands in the 1800s (O’Gara and Yoakum 2004), the population declined from 250 animals in 1991 to fewer than 50 individuals in 2003 in the United States (US) range (USFWS 2015) purportedly due to drought, habitat loss and fragmentation due to fencing and human activity along the US–Mexico border (USFWS 1998, O’Gara and Yoakum 2004, Wilson et al. 2010). Subsequently, a 2.56-km² captive breeding pen was established on the Cabeza Prieta National Wildlife Refuge (CPNWR) to facilitate recovery efforts, and captive individuals are released annually into the wild.
(Otte 2006). In summer 2013, there were approximately 100 captive individuals in the pen. Population estimates for the wild population are derived from a biennial aerial count, and the population size in 2014 was estimated at 202 individuals (95% CI: 171–334; USFWS 2015). The aerial survey does not provide recruitment estimates as only a proportion of individuals are observed during the survey are classified to sex or age class due to the potential for disturbance caused by the aircraft needing to fly low to accurately perform these classifications (J. J. Hervert, Arizona Game and Fish Dept., pers. comm.). Survival is also not estimated.

Assigning age through traditional methods (e.g. examining tooth wear or tooth replacement) to an individual to track it throughout its lifetime usually involves capture and handling which can be expensive, dangerous and potentially injurious or lethal to the animal (Murray and Fuller 2000, Arnemo et al. 2006, Solberg et al. 2006). Pronghorn are especially sensitive to capture (Kreeger et al. 2002), and myopathy is not an uncommon consequence during capture and handling (Chalmers and Barrett 1977, Bright and Hervert 2005, Yoakum et al. 2013); thus, alternative methods, such as noninvasive sampling methods are appealing. Morphological measurements of feces (i.e. pellets, bolus) have been used to determine age class in a variety of ungulate species (Table 1). Most studies required use of multiple measurements for successful age classification. Success rates varied from 75–100% depending on age class, measurement, and study species. Pellet volume distinguished sex and age class of moose Alces alces (MacCracken and Van Ballenberghe 1987). A combination of length and width distinguished age classes in Manipur brow-antlered deer Cervus eldi eldi (Khan and Goyal 1993) and Svalbard reindeer Rangifer tarandus platyrhynchos (Morden et al. 2011). Pellet weight also corresponded to body weight in moose (MacCracken and Van Ballenberghe 1987) and Manipur brow-antlered deer (Khan and Goyal 1993). Bubenik (1982) additionally found sex-specific differences in pellet shape in elk Cervus elaphus. Ball (2010) and Morden et al. (2011) suggest combining aging by measurements with fecal DNA for estimating demographic parameters in ungulate species.

The use of noninvasively obtained DNA samples (e.g. feces, hair, saliva) as a tool for measuring population parameters has become common in mammal populations. Methods such as fecal DNA microsatellite analysis have proven to be a useful tool for estimating demographic parameters, such as abundance, survival or sex ratio (Waits and Paetkau 2005, Schwartz et al. 2007, Beja-Pereira et al. 2009, De Barba et al. 2010). One weakness of this method, however, is the difficulty to age individuals with noninvasive genetic samples, yet understanding the age structure of a population is central to understanding age-specific survival and recruitment. The ability to distinguish age classes greatly improves the applicability of noninvasive genetic sampling and thus would be especially useful in a monitoring program designed to measure key demographic parameters such as population size, survival rates and recruitment. Here, we examine the use of fecal pellet dimensions of endangered Sonoran pronghorn Antilocapra americana sonoriensis to distinguish between fawns, yearlings and adults.

Our specific objectives were to determine if pellet size (either single measurement, e.g. length, or a combination of measurements, e.g. length + width) and shape could be used to distinguish age class and to test our ability to determine age class by visual assessment of pellet size and morphology in the field. We expected adult pellets would be larger than yearling and fawn and yearling pellets would be larger than fawn. We also predicted we would be able to assign coarse age class (i.e. fawn <1 year or non-fawn = ≥ 1 year) in the field based on visual assessment of size and morphology.

### Methods

### Study area

The Sonoran Desert has average high temperatures of over 38°C in summer and is one of the hottest and driest regions of North America (INRMP 2003). From June to October, temperatures can exceed 32°C for more than 100 consecutive days on CPNWR (USFWS 2002). Average annual precipitation on the CPNWR ranges from 20 cm in the east to 7.5 cm in the west falling in monsoons during July–August and December–January (USFWS 2002). April to June is the dry season.

In May and June of 2012–2014, we collected pronghorn fecal pellets in the captive pen and from wild individuals at 12 developed watering holes (hereafter drinkers) and nine sites not associated with drinkers on Organ Pipe National Monument (ORPI), CPNWR, and the adjacent Barry M. Goldwater Range (BMGR) in southwest Arizona, USA (Fig. 1). The 2.56 km² captive pen holds two separate populations which total ~100 free ranging individuals (USFWS 2015), and the wild population in the US

<table>
<thead>
<tr>
<th>Species</th>
<th>Study</th>
<th>Age classes discriminated</th>
</tr>
</thead>
<tbody>
<tr>
<td>Boreal caribou Rangifer tarandus caribou</td>
<td>Ball 2010</td>
<td>calf, yearling, adult</td>
</tr>
<tr>
<td>Elk Cervus elaphus</td>
<td>Bubenik 1982, Alvarez 1994</td>
<td>juvenile, adult</td>
</tr>
<tr>
<td>Fallow deer Dama dama</td>
<td>Alvarez 1994</td>
<td>calf, juvenile, adult</td>
</tr>
<tr>
<td>Greater bilby Macrotis lagotis</td>
<td>Southgate 2005</td>
<td>immature, mature</td>
</tr>
<tr>
<td>Manipur brow-antlered deer Cervus eldi eldi</td>
<td>Khan and Goyal 1993</td>
<td>five age classes</td>
</tr>
<tr>
<td>Moose Alces alces</td>
<td>MacCracken and Van Ballenberghe 1987</td>
<td>yearling, adult</td>
</tr>
<tr>
<td>Mule deer Odocoileus hemionus</td>
<td>Sanchez-Rojas et al. 2004</td>
<td>yearling, adult</td>
</tr>
<tr>
<td>Sumatran elephant Elephas maximus sumatranus</td>
<td>Reilly 2002, Tyson et al. 2002</td>
<td>juvenile, sub-adult, adult</td>
</tr>
<tr>
<td>Svalbard reindeer Rangifer tarandus platyrhynchos</td>
<td>Morden et al. 2011</td>
<td>calf yearling, adult</td>
</tr>
<tr>
<td>White-tailed deer Odocoileus virginianus</td>
<td>Ezcurra and Gallina 1981</td>
<td>yearling, adult</td>
</tr>
</tbody>
</table>
was estimated at 202 individuals (95% CI: 171–334) in December 2014 (USFWS 2015). Arizona Game and Fish Department (AZGFD) and US Fish and Wildlife Service (USFWS) personnel provide alfalfa at feed stations both in the wild and the pen. Conditions in the pen are kept as natural as possible and portions of the pen are irrigated to encourage growth of natural forage. Fawning starts in the pen in mid-February and most fawns are born in March and April (Wilson et al. 2008). Wild pronghorn are fed alfalfa weekly during the dry months at six of the 12 drinkers outside the pen. Feeding and watering typically begin in April or May and continue through October or November depending on annual rainfall amounts (J. Atkinson, U.S. Fish and Wildlife Service, pers. comm.). In the wild, fawns are typically born between February and June with one birth reported as early as January in 2013 (Bright and Hervert 2005, USFWS 2015).

Sample collection

Each December, annual capture operations are conducted in the captive pen by AZGFD and USFWS, during which individuals are captured, radio-collared, and a blood sample is collected. Fawns captured in the pen are ear-tagged, and radio-collared if recaptured in subsequent captures. Young of the year are easily identified (e.g. size, horn development) and classified as fawn (0–11 months), and individuals captured as fawns in the previous year are known yearlings (12–23 months). However, not all fawns are caught during annual capture operations and consequently, an individual may not be handled until it is >1 year old and is potentially misclassified as to actual age. Thus, all captured animals of unknown age are classified as adults. Some captured (captive) individuals are subsequently released into the wild, at a ratio of approximately two males to one female (USFWS 2015). To obtain DNA, blood samples were collected from 58 captured individuals in December 2012 and 2013 when feasible (i.e. if health and safety of the animal was not at risk due to stress) (USFWS 2015). These samples provided a genotype of an individual of known age class for later matching to genotypes obtained from fecal pellets collected in the pen or in the wild.

In May 2012 in the captive pen, we collected five fecal pellets (per Morden et al. 2011) less than 24 h old from each of 185 fecal pellet piles in three pellet size groups defined visually as small, medium, and large. Size refers to the pellet, not size of pile, and all pellets from a pile were given the same size classification. While we recognize our size classification is subjective, we wanted to ensure collection of all age and sex classes, and this size classification was used only to structure collection and was not part of the analyses. To ensure collection of samples less than 24 h old, we cleared pellets from the area around feed stations in the captive pen on the day prior to collection. We collected fresh pellets as our pilot studies indicated low nuclear DNA (used for individual identification) PCR success rates (2–28%) by day 14 and 0% by day 60 (Woodruff et al. 2015, Woodruff unpubl.). We excluded piles that appeared to be from more than one non-drinker locations
▲ Drinkers
■ Captive pen
— Roads

Figure 1. Map depicting Sonoran pronghorn fecal sampling locations (drinkers and non-drinkers) in 2013 and 2014 on Barry M. Goldwater Range (BMGR), Cabeza Prieta National Wildlife Refuge (CPNWR), and Organ Pipe National Monument (ORPI), southern Arizona, USA. Each star represents general locations of multiple sampling areas.
individual based on pellet shape, color and size to minimize the potential for wasted effort in the laboratory, as these are likely mixed samples. We placed pellets in paper coin envelopes and stored them at room temperature in a plastic ziploc bag with – 250 ml of silica desiccant to minimize DNA degradation prior to analysis (Soto-Calderon et al. 2009, DeMay et al. 2013). To determine the age of the individual from which the fecal sample was collected, we matched 7–16 locus microsatellite genotypes of fecal samples to the blood samples.

Due to high rates of capture myopathy, only limited capture and radio-collaring of wild Sonoran pronghorn occurs, and the majority of marked animals in the wild are captive released individuals (USFWS 2015). To test our ability to assign age class, fawn or non-fawn (≥ 1 year old), in the wild population, we measured pellet samples collected from wild pronghorn in May and June of both 2013 and 2014 as part of a larger mark–recapture study (or detection–redetection since there is no physical capture; Woodruff et al. 2016). We visited drinkers when pronghorn had left the area and collected fecal pellets surrounding feed stations and drinkers. At least six pellets from each sample were collected, and stored as described above. We collected six pellets per sample to ensure we had a sufficient number of pellets to perform two DNA extractions (Woodruff et al. 2015). Samples were initially field-classified as fawn or non-fawn based on visual assessment of size and morphology. We had two sources of known age wild individuals: 1) genotypes of known age individuals released from the captive pen obtained from blood samples, which consisted of two known adults captive released in 2012 and sampled (i.e. through fecal pellets) in the wild in 2013, and 2) 15 individuals redetected in 2014 (thus, known non-fawn) from fecal pellet genetic analysis in 2013.

DNA extraction and genotyping

DNA extraction of blood samples was conducted using a Qiagen DNeasy blood and tissue kit (n = 10) or by over-night lysis with ProteinaseK (10 mg ml–1) at 55°C, followed by a modified protocol (n = 48) based upon the standard phenol/chloroform extraction and isoopropanol/sodium acetate precipitation (Sambrook et al. 1989). We used Phase Lock gel tubes (5-Prime) to aid in the separation between organic and aqueous phases and resuspended the DNA in Low TE (10 mM Tris-pH 8.0, 0.01 mM EDTA). Fecal pellet DNA was extracted using the QIAamp DNA Stool Mini Kit following methods described in Adams et al. (2011) and Woodruff et al. (2015). For individual ID, ten nuclear DNA (nDNA) microsatellite loci ranging in size from 90–278 base pairs and one sex ID locus were amplified in a single multiplex reaction (Lou 1998, Carling et al. 2003, Munguia-Vega et al. 2013). The 7 µl PCR reaction contained 1 × Qiagen Master Mix, 0.5 × Qiagen Q-Solution, 1.71 µM Anam97, 0.04 µM Anam50, 0.07 µM Anam82, 0.01 µM Anam79, 0.86 µM Aam13, 0.43 µM Aam11, 0.14 µM ADCYC, 0.26 µM Aam10, 0.04 µM Aam1, 0.04 µM Aam2, 0.29 µM KY (sex ID), and 1.5 µl DNA extract. The PCR profile included an initial denaturation of 95°C for 15 min, followed by a touchdown of 20 cycles with a 30 s denaturation at 94°C, 90 s annealing step at 63°C decreasing 0.5°C each cycle, and a 60 s extension at 72°C, followed by 34 cycles of a 30 s denaturation at 94°C, a 90 s annealing step at 53°C, and a 60 s extension at 72°C. The cycle finished with a 30 min final extension at 60°C.

We initially screened all samples with two PCR replicates to assess sample quality, and samples failing to amplify at ≥ 5 loci were dropped from additional genotyping to remove low quality, error-prone samples from the dataset. To obtain a consensus genotype, three to eight PCR replicates were performed per sample. Consensus genotypes were based on multiple runs of a sample as follows: 1) for homozygotes, the allele was present at least three times, and 2) for heterozygotes, we had to see each allele at least two times. We repeated this testing and evaluating process until we obtained a consensus genotype at a minimum of seven loci to meet the matching criteria of ≤ 0.01 probability of identity siblings (P(ID)sibs) (Waits et al. 2001). Consensus genotypes were determined in Microsoft Access (Skirbinske 2010), and matching and P(ID)sibs analysis was conducted in GenAIEx 6.5 (Peakall and Smouse 2006). Within the captive pen, fecal pellet genotypes were first matched to other fecal pellet genotypes. Then, unique genotypes were matched to genotypes of blood samples for individuals of known age. In the wild population, fecal pellet genotypes were matched to other fecal pellet genotypes, and individuals redetected in year two were then known to be ≥ 1 year old (non-fawn).

**Pellet dimension measurements**

Using digital calipers we measured (mm) maximum length (L), maximum width (W), and calculated length-to-width ratio (L/W), and approximate volume (V: 4/3π (L/2)(W/2)^2; volume of an ellipsoid – the approximate shape of pellets), where W is used for width and diameter, of each of the five pellets per sample from the captive pen within two hours of sample collection. We calculated the mean measurements of the five pellets to represent the sample. From wild individuals, we attempted to measure five pellets from at least one pellet group from each individual (confirmed by DNA analysis). However, we did not always have enough pellets post-extraction for measuring, and thus not all individuals had a measured sample. Samples with fewer than two pellets were discarded and we excluded broken, split, or partial pellets as recommend by Zahratka and Buskirk (2007). To evaluate the effect of time and desiccation on the measurements, we re-measured 16 samples after seven days of exposure to local field conditions and reran the calculations. Mean daily high temperature was 37.3°C (range: 33.6–40.7°C), and there was no precipitation during this 7-day period.

**Statistical analysis**

We conducted all statistical analyses using R ver. 3.1.2 (<www.r-project.org>). We used t-tests with an alpha of 0.05 to examine differences in pellet measurements between captive and wild individuals and between measured (fresh) and re-measured (7-day old) samples. We used logistic regression to distinguish between age classes. Using logistic regression (brglm function, R package brglm; Kosmidis 2013), we evaluated the predictive accuracy of models using
each measurement (e.g. length, volume) as a single predictor variable, as well as models using all possible combinations of the predictor variables (e.g. length and volume). Predictive accuracy was estimated using randomized five-fold cross-validation with the cvFit function from the R package cvTools (Alfons 2012) with 100 replications, and models were also ranked using Akaike’s information criterion corrected for small sample size (AICc). To avoid infinite parameter estimates due to complete or quasi-complete separation, the model parameters were estimated using bias-reduced maximum likelihood (Firth 1993, Kosmidis and Firth 2009). Probability of correct classification was compared to a null model (i.e. no explanatory variables) where the predictive accuracy depends only on the base rate of classes in the population (e.g. if 60% of the population are adults, then prediction accuracy would be 60% if every sample were called adult).

For age class, cross validation analyses indicated that it was not possible to reliably distinguish between yearlings and adults based on pellet morphology significantly better than using the base rate (probability = 0.69–0.72; null model probability = 0.67), so our age class analysis focused on separate analyses distinguishing fawn from yearling and fawn from adult. We used only known age captive individuals and known age captive released wild individuals for age class analysis (i.e. excluded wild individuals of unknown age). Finally, we combined yearlings and adults into a single age class (non-fawn) and tested the accuracy of our visual field-classification of wild samples by age class, i.e. fawn or non-fawn. We used the fitted model developed from known age individuals (captive fawns and captive and wild non-fawns) to predict age class of wild samples. We used models with single explanatory variables of width and volume because these were two of the best AICc ranked models for predicting age class. In cases where the model-predicted age class differed in multiple samples from the same individual (i.e. one predicted adult, one predicted fawn), we ran an additional model (length + length–width ratio).

Results

We collected 185 pellet groups from 58 small, 69 medium and 58 large (visual classification) fecal pellet piles from the captive pen (Supplementary material Appendix 1 Fig. A1). From these, we confirmed consensus genotypes for 176 which were matched and found to represent 67 individuals or approximately 85% of the pronghorn in the pen in 2012 (USFWS 2015). During physical capture of captive animals, we obtained 58 blood samples over two years from known age individuals. We genotyped and matched 33 of the blood samples (four adults, 17 yearlings, 12 fawns) to 87 fecal samples (Table 2). The remaining 98 pellet piles were not matched to known age individuals and were not included in our analyses. All samples collected in the captive pen were measured, and the number of samples per individual averaged 2.8 (range 1–8). From the wild, we measured 258 samples across years. We did not find significant differences between measurements of fresh pellets and pellets re-measured (n = 16) after seven days (p > 0.05).

<table>
<thead>
<tr>
<th>Age class</th>
<th>Captive</th>
<th>Wild</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>Male</td>
<td>Female</td>
</tr>
<tr>
<td>Fawn</td>
<td>14</td>
<td>7</td>
</tr>
<tr>
<td>Yearling</td>
<td>40</td>
<td>5</td>
</tr>
<tr>
<td>Adult</td>
<td>12</td>
<td>9</td>
</tr>
<tr>
<td>Total</td>
<td>66</td>
<td>21</td>
</tr>
</tbody>
</table>

Table 2. Total number of pellet samples measured in both the captive and wild Sonoran pronghorn populations in summer 2013 and 2014, Arizona, USA. Captive samples are known age class from individual ID genotype matching. Wild individuals classified as adult are non-fawn (i.e. yearlings and adults) and were field (visual) classified.

Pellet size and age class

For both captive and wild, length, width and volume of yearling and adult (or non-fawn) pellets were greater than fawns, but length–width ratio was larger for fawns, indicating fawn pellets were rounder than either yearling or adult pellets (Fig. 2, Table 3a). Probability of correct age class classification for captive fawn versus yearling using a single variable was 0.98 for width only and was the best AICc ranked model (Table 4a). Other single variable models and combinations of ≥2 explanatory variables performed similarly with predictive probability ranging from 0.91 to 0.98 and delta AICc (difference in AICc compared to the lowest AICc) from 1.66 to 7.16. Length as a single explanatory variable was the worst ranked model with probability of correct classification of 0.77 and delta AICc value of 7.16. The null model (i.e. the model with no explanatory variables) had probability of correct classification of 0.32 (i.e. if we
guess fawn, we will be right 32% of the time because 32% of our samples were fawn).

The best AICc ranked model for captive fawn versus adult again included pellet width only with probability of correct age class classification of 0.98 (Table 4b). The volume only model and combinations of ≥2 explanatory variables performed similarly with predictive probability ranging from 0.93 to 0.97 and delta AICcs from 1.46 to 8.41. Length and length–width ratio as single explanatory variables were the worst ranked models with probability of correct classification of 0.85 and 0.77 and delta AIC values of 17.77 and 26.65, respectively. The null model (i.e. the model with no explanatory variables) had probability of correct classification of 0.50 (i.e. 50% of pellet samples were adult and 50% were fawn).

Captive versus wild measurements of pellets

Mean length, width and volume of wild non-fawn pellets (≥1 year as evidenced by redetection) were smaller than those of captive non-fawns (Fig. 3). Mean length–width pellet ratio, however, was consistent for non-fawn captive (1.59) and wild (1.68) individuals. Pellet length, width and volume were significantly different between captive and wild non-fawns (Table 3b). For fawns, only pellet width was significantly different between captive and wild individuals.

Prediction of age class from wild pellet samples

In 2014, we redetected 15 individuals (n = 39 samples), thus known to be non-fawn (≥1 year old). All samples from redetected individuals were correctly model-predicted as non-fawn in either the width only or volume only model (Table 5, Supplementary material Appendix 1 Table A1).

Ten unique individuals (a total of 16 samples) had conflicting model-predicted (width only and volume only models) age class (i.e. one model predicted non-fawn, one predicted fawn). There was no consistency in which model (e.g. width only) predicted fawn or non-fawn in these cases. For eight of these individuals, we measured multiple samples,
In 2013, only one sample that was visually field-classified as non-fawn \( (n = 119) \) was model-predicted as fawn, whereas three samples visually field-classified as fawn were subsequently model-predicted as non-fawn (Table 5, Supplementary material Appendix 1 Table A1). Similarly, in 2014, 98\% \( (n = 62) \) of samples visually field-classified as non-fawn were model-predicted as non-fawn, but only 33\% \( (n = 22) \) of samples field-classified as fawn were model-predicted fawn. Nearly all samples (46 of 49 in the years combined) incorrectly visually field-classified as fawn were male.

Correct visual field-classification by age class varied considerably between years. In 2013, we measured 128 samples from 76\% \( (n = 72) \) of the total wild individuals we detected \( (\text{i.e. detected their genotype}) \) (Table 2). Ninety-six percent of visual field-based classifications matched the model predictions (Supplementary material Appendix 1 Table A1). In 2014, we measured 130 samples from 64 individuals, and 65\% of samples were correctly field-classified by visual observation \( (\text{i.e. matched the model prediction}) \).

Table 5. Comparison of visual field classification and measurement prediction for pellet samples collected for wild pronghorn in 2013 and 2014, Arizona, USA.

<table>
<thead>
<tr>
<th>Field classification</th>
<th>Measurement prediction</th>
<th>No. measured</th>
<th>No. non-fawn (%)</th>
<th>No. fawn (%)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Adult</td>
<td></td>
<td>122</td>
<td>119 (98)</td>
<td>3 (2)</td>
</tr>
<tr>
<td>Fawn</td>
<td></td>
<td>64</td>
<td>46 (72)</td>
<td>18 (28)</td>
</tr>
</tbody>
</table>

and we took the majority model-predicted classification \( (\text{i.e. four model-predicted as non-fawn, two model-predicted as fawn, we called it non-fawn}) \). The two other individuals were single detections \( (\text{i.e. we had only one sample}) \) for which we ran an additional fitted model \( (\text{length + length–width ratio}) \) and used the age class that was predicted twice \( (\text{i.e. volume and width models both predicted fawn, and length + length–width ratio predicted adult, we called the individual fawn}) \).

In 2013, only one sample that was visually field-classified as non-fawn \( (n = 119) \) was model-predicted as fawn, whereas three samples visually field-classified as fawn were subsequently model-predicted as non-fawn (Table 5, Supplementary material Appendix 1 Table A1). Similarly, in 2014, 98\% \( (n = 62) \) of samples visually field-classified as non-fawn were model-predicted as non-fawn, but only 33\% \( (n = 22) \) of samples field-classified as fawn were model-predicted fawn. Nearly all samples (46 of 49 in the years combined) incorrectly visually field-classified as fawn were male.

**Discussion**

Understanding age-specific survival rates is critical to management and identifying long-term trends in population growth, particularly in endangered species, yet documenting age-specific survival relies on knowing age structure. We successfully demonstrated the ability to distinguish fawns from yearlings or adults using morphometric pellet measurements of Sonoran pronghorn. While we did not have enough predictive power to separate yearlings and adults, an understanding of recruitment rates and fawn survival is important in determining the trajectory of the population. This measurement method provides a reliable way to document fawns, and the use of fecal DNA for individual identification further strengthens the method’s utility as individuals can be tracked over multiple years of sampling.

![Figure 3. Box plots showing mean measurements of Sonoran pronghorn fecal pellets collected in May and June 2012–2014 in Arizona, USA, from captive and wild fawn (F) and non-fawn (NF) for length, width, length–width ratio (LW), and volume. The box signifies the upper and lower quartiles and the median is represented by the thick black line. Black dots represent outliers.](https://bioone.org/journals/Wildlife-Biology on 15 May 2021 Terms of Use: https://bioone.org/terms-of-use)
**Pellet size and age class**

Our visual classification of pellets of captive pronghorn by size (small, medium, large) was a mechanism to facilitate collection of all age and sex classes and was not used in the analysis, but our results indicate this was an appropriate collection method for obtaining a range of age classes for our analyses. As expected, pellet size was larger for yearlings and adults than for fawns. The probability of correctly assigning age class models was similar for fawns and yearlings and fawns and adults and performed notably better than the null models (null probability 0.32 and 0.50, respectively). On the other hand, both Morden et al. (2011) and MacCracken and Van Ballenberghe (1987) found higher rates of correct model classification for adults compared to juveniles. Contrary to other studies (Delibes-Mattos et al. 2009, Morden et al. 2011), our cross-validation analysis showed that single measure models assigned age class with roughly the same accuracy as models with multiple variables. Our small sample size (n = 4) for adults is perhaps the reason we could not distinguish between adults and yearlings. Collecting and measuring fecal pellets from additional known adults would potentially allow for this distinction.

Some variables were correlated as would be expected given that two variables (length–width ratio and volume) are calculated using combinations of the other variables, and length and width also tend to be correlated. Thus, it is not surprising that models with additional variables generally did not substantially outperform the width-only model.

**Captive versus wild**

Similar to other research, our results suggest that the most accurate predictive models are built from samples collected from the target population (MacCracken and Van Ballenberghe 1987, Chapman 2004, Morden et al. 2011). Differences in diet (e.g. seasonal variation, captive versus wild) can affect size of fecal output (Campos-Arceiz et al. 2008, Morden et al. 2011) and defecation rate (Rogers 1987, Mayle et al. 1996, Chapman 2004, Ferretti et al. 2014) with both reduced (Irby 1981, Asa et al. 1985, Rogers 1987, Kitchen and Martin 1996, Chapman 2004) and defecation rate in captive individuals (Smith 1964, Dinerstein and Dublin 1982). Brashares and Arcese (1999a, b) suggest dominant male oribi *Ourebia ourebi* restrict defecation volume and occurrence in order to increase the frequency of territory marking events. Consequently, there could be differences in defecation rate, and potentially pellet size, between captive and wild individuals, as well as fed and unfed individuals in the wild. However, we believe that our approach to develop predictive models using captive individuals was justified and effective for multiple reasons. First, in our study system, any future fecal DNA sampling will be conducted at drinkers, nearly 50% of which also provide supplemental feed. Additionally, as conditions in the pen are semi-natural, captive pronghorn also feed on natural forage year-round due to irrigated forage plots. Second, we were able to test our methods on 15 wild animals known to be non-fawn. All were correctly identified as non-fawn by the model-prediction. Third, length–width ratios between captive and wild animals were not significantly different and indicated that pellet size changed consistently across measurements from captive to wild. Fourth, our results illustrate significant differences in pellet size between fawn and non-fawn in both the captive and wild populations, with wild fawn having the smallest measurements and thus having a lower chance of being misclassified as non-fawn compared to captive animals.

**Accuracy of visual-based field classification**

Our results suggest that age classification based on visual assessment is imprecise and likely influenced by individual differences in observers. In 2013, we had two people collecting samples who were trained simultaneously and extensively. In 2014, four new personnel collected samples and training was limited. Based on model predictions, visual-based field classification was more accurate in 2013 (96% correct) compared to 2014 (65% correct). Samples were more often field-classified as fawn and consequently model-predicted as, or known to be, non-fawn (n = 19) than vice versa (n = 2). While we detected twice as many males as females (Woodruff et al. 2016), 94% (46 of 49) of samples incorrectly visually field-classified as fawn were male suggesting fecal pellets of adult males are more difficult to classify than adult females, contrary to MacCracken and Van Ballenberghe (1987) where correct classification was similar for males and females. In 2014, two redetected individuals made up 24% (n = 11) of the samples incorrectly field-classified as fawn. One of these individuals was field-classified and model predicted as fawn in 2013 and would putatively be a yearling in 2014 perhaps providing explanation for the misclassification. The other individual was detected at a site with supplemental feed and field-classified and model predicted as non-fawn in 2013. However, in 2014 this individual was redetected at a site with no supplemental feed and was field-classified as fawn in all samples, but model-predicted as adult in 64% (n = 7) of samples. While based on only a single individual, this strengthens the idea that fed individuals produce larger pellets (Hummel et al. 2008). We recognize the weaknesses of our visual assessment method, including the skill level and subjectivity of observers, but we felt it was valuable to test the method. With in-depth training of observers (as seen in 2013), the accuracy of this method could be improved.

In turn, we recommend measuring pellets when trying to determine age class to increase accuracy. We also note that developing a model based on measurements from wild pronghorn would strengthen predictive power. However, we posit that when obtaining samples from wild individuals is not feasible, samples from captive animals provide a reasonable surrogate (Chapman 2004).

**Management implications**

In species that are highly sensitive to capture and disturbance, like Sonoran pronghorn, noninvasive genetic sampling methods are an appealing alternative approach for obtaining critical demographic data. Managers of Sonoran pronghorn do not have current sex and age-based survival estimates because these data cannot be obtained with the current aerial survey approach used for population estimation.
conjunction with individual identification from fecal pellets have lower DNA amplification success rates, and however, other research shows older Sonoran pronghorn days. Additional pilot studies are needed to evaluate this.ments, results could differ if pellets are older than seven days. Additional pilot studies are needed to evaluate this. However, other research shows older Sonoran pronghorn pellets have lower DNA amplification success rates, and if the objective is to use these measurement methods in conjunction with individual identification from fecal DNA, pellets older than seven days will be of little use (Woodruff et al. 2015).

While we could not determine sex using morphological methods (data not shown), sex is easily ascertained using genetic methods, and our method shows strong potential for use in wild populations of pronghorn to distinguish fawns from yearlings or adults. The ability to distinguish age structure using fecal pellet measurements greatly improves the applicability of fecal pellet collection and has significant implications for management of the Sonoran pronghorn population. We recommend analyzing pellets using both single width and volume variable models and running the combined model (length + width–width ratio) if the width and volume models disagree. A long-term monitoring scheme combining pellet measurements with individual identification using DNA would allow the identification of fawns that would remain known age individuals for the remainder of the monitoring years. This would provide a powerful tool for documenting the population’s age and sex structure, annual recruitment, and age-specific annual survival rates. As research has shown, distinguishing age classes is possible in other ungulate species, and as the use of noninvasive genetic methods increases in ungulates, these methods could potentially be applied to other species.

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